Final Accepted Version: 1

2 Bätz, N., Verrecchia, E.P. and Lane, S.N., 2015. Organic matter processing and soil evolution in a braided river system. Catena, 126 86-97.

- Typeset article of record available at: 4
- 5 https://doi.org/10.1016/j.catena.2014.10.013
- 6

3

- 7
- Organic matter processing and soil evolution in a 8

braided river system 9

- *Nico Bätz*; Institute of Earth Surface Dynamics University of Lausanne, Switzerland. 10
- Email: nico.baetz@unil.ch 11

Eric P. Verrecchia; Institute of Earth Surface Dynamics - University of Lausanne, Switzerland. 12

Email: eric.verrecchia@unil.ch 13

14 Stuart N. Lane; Institute of Earth Surface Dynamics - University of Lausanne, Switzerland.

- Email: stuart.lane@unil.ch 15
- 16

Abstract 17

Traditionally, braided river research has considered flow, sediment transport 18 processes and, recently, vegetation dynamics in relation to river morphodynamics. 19 However, if considering the development of woody vegetated patches over a time 20 scale of decades, we must consider the extent to which soil forming processes, 21 particularly related to soil organic matter, impact the alluvial geomorphic-vegetation 22 system. Here we quantify the soil organic matter processing (humification) that occurs 23

24 on young alluvial landforms. We sampled different geomorphic units, ranging from the active river channel to established river terraces in a braided river system. For each 25 geomorphic unit, soil pits were used to sample sediment/soil layers that were analysed 26 27 in terms of grain size (<2mm) and organic matter quantity and quality (RockEval method). A principal components analysis was used to identify patterns in the dataset. 28 Results suggest that during the succession from bare river gravels to a terrace soil. 29 there is a transition from small amounts of external organic matter supply provided by 30 sedimentation processes (e.g. organic matter transported in suspension and deposited 31 on bars), to large amounts of autogenic in situ organic matter production due to plant 32 colonisation. This appears to change the time scale and pathways of alluvial 33 succession (bio-geomorphic succession). However, this process is complicated by: the 34 ongoing possibility of local sedimentation, which can serve to isolate surface layers via 35 aggradation from the exogenic supply; and erosion which tends to create fresh 36 deposits upon which organic matter processing must re-start. The result is a complex 37 pattern of organic matter states as well as a general lack of any clear chronosequence 38 within the active river corridor. This state reflects the continual battle between 39 deposition events that can isolate organic matter from the surface, erosion events that 40 can destroy accumulating organic matter and the early ecosystem processes 41 necessary to assist the co-evolution of soil and vegetation. A key guestion emerges 42 43 over the extent to which the fresh organic matter deposited in the active zone is capable of significantly transforming the local geochemical environment sufficiently to 44 accelerate soil development. 45

46

47 Highlights:

48	Grain size and organic matter characteristics are key pedo-biogeomorphic
49	variables;
50	Geomorphic processes can potentially add valuable resources to the young
51	ecosystem;
52	Thus, geomorphic processes can hamper but also facilitate alluvial soil
53	evolution;
54	Initial alluvial soils experience a transition from exogenic towards endogenic
55	matter input.
56	
57	Keywords: Alluvial soils; Braided river; Rock-Eval; Organic matter; River terraces;
58	Biogeomorphic succession

59 **1. Introduction**

Geomorphologically active systems, such as braided rivers, exhibit a complex mosaic of fluvial habitats (Tockner et al., 2010) including bare sediment surfaces, islands within the active zone at various vegetation succession stages, and established river terraces with floodplain forest and well-developed soils. Thus, the river landscape comprises a range of ages with reworked zones, and ages at the sub yearly timescale, to much more stable zones, potentially many decades old.

66

Recent research has established that the transition from a bare sediment surface to a
vegetated patch results in important changes in fluvial processes. Vegetation can be
seen as a type of ecosystem engineer, critically involved in this transition during fluvial
landform formation (Corenblit et al., 2014, 2011; Gurnell, 2014; Gurnell et al., 2012;
Jones et al., 1994; Osterkamp and Hupp, 2010) by: (i) stabilising sedimentary deposits

through rooting (Crouzy and Perona, 2012; Perona et al., 2012); and (ii) enhancing 72 fine sediment deposition due to above ground biomass induced energy losses that 73 lead to surface aggradation (Gurnell and Petts, 2002). Both plant-facilitated processes 74 allow habitat development within the most active zones of the floodplain by improving 75 local edaphic conditions (moisture and nutrient retention, reduced susceptibility to 76 erosion) so allowing the progress of succession – from pioneer island species to stable 77 terrace hardwood species (e.g. Francis, 2007; Francis et al., 2009; Gurnell et al., 2001; 78 Moggridge and Gurnell, 2009). Nevertheless, if the deposition rate is too high, 79 vegetation may get buried, leading to an optimal aggradation range for successional 80 processes (Gurnell and Petts, 2002). Conversely, if the erosion rate is too high, the 81 entire vegetated patch may be removed and its materials redeposit elsewhere, where 82 it may again facilitate plant development (large woody debris, e.g. Francis, 2007; 83 Francis et al., 2008). 84

85

These processes have been recently conceptualised into a biogeomorphological life 86 cycle model for Populus nigar, deemed to be valid for Salicaceae pioneer vegetation 87 in general (Corenbilt et al., 2013). The main phases of the life cycle identified are: (i) 88 in the geomorphological phase, seedlings are dispersed by floods and germinate on 89 suitable bar surfaces; (ii) in the following pioneer phase, seedlings are challenged by 90 water stress, erosion and deposition processes. During these first two phases, P. nigra 91 is completely exposed to the physical riverine processes without relevant feedbacks to 92 river morphology; (iii) in the third phase, interaction between plants and their physical 93 environment is highest, the biogeomorphological phase. Young trees take on an 94 engineering role by fixing sediments and trapping fine sediments. Symbiosis with 95 endomycorrhizal fungi improves their access to the soil nutrient pool and groundwater 96

97 (Harner et al., 2011). Finally, (iv), during the last or ecological phase, the vegetated 98 patch becomes relatively independent from the river and is able, via autogenic 99 ecosystem processes, to auto-sustain its own resource demands. Rare shallow 100 overland deposition or lateral erosion processes are the main riverine processes 101 affecting this phase (Corenblit et al., 2014).

102

The latter two stages imply timescales of the order of years to decades. At these longer 103 timescales soil, as an emergent property of the developing ecosystem, must also be 104 considered as an element of the braided river system (Bätz et al., 2014a). In stable 105 systems, such as river terraces of meandering systems, pedogenesis has been 106 extensively studied (Cierjacks et al., 2011, 2010; Gerrard, 1987). However, this is 107 much less the case in more dynamic alluvial environments, such as braided rivers, and 108 hence the question arises: are soil forming processes passive process that reacts to 109 stabilizing geomorphic conditions, or are they actively involved in controlling the rate 110 of biogeomorphic succession (sensu Corenblit et al., 2009)? In other words, is 111 pedogenesis able to change the rate of biogeomorphic succession? 112

Because the later stages of Corenbilt et al.'s (2013) model imply the presence of soil, 113 a second question follows: to what extent is pedogenesis involved in the first two of 114 Corenbilt et al.'s (2013) stages? For instance, flood pulses, that lead to deposition, 115 may provide water but also exogenously produced energy rich organic matter (plant 116 debris but also pedogenically transformed material), that is easily decomposed and 117 humified into plant available forms by the sediment/soil micro-flora and fauna (Cabezas 118 and Comín, 2010; Francis, 2007; Gregory et al., 1991; Langhans et al., 2012; Naegeli, 119 1997; Pusch et al., 1998; Tabacchi et al., 2000). This depositional process might 120 significantly enhance the nutrient pool of nutrient-poor, young mineral sediments, and 121

so accelerate initial ecosystem processes including soil forming processes (Doering et 122 al., 2011; Guenat et al., 1999; Guex et al., 2003; Langhans et al., 2012). However, 123 either massive deposition, leading to burial, or erosion events which lead to local loss 124 of pedogenically transformed organic matter, may potentially hamper fluvial planform 125 development. These interactions with the biogeomorphic component, can lead to a 126 multitude of pathways and trajectories of alluvial soil formation, so that it is better to 127 talk about soil evolution (Johnson, 1985; Schaetzl and Anderson, 2005). Yet, we know 128 surprisingly little about initial soil development on active surfaces of braided river 129 130 deposits and its interaction with biogeomorphological processes (Bätz et al., 2014a).

131

Addressing the question of initial soil evolution requires a multi-angled approach 132 (Bernasconi and Biglink Project Members, 2008). In this paper we focus upon the 133 question of organic matter processing (humification), which is considered an important 134 part of young ecosystems. Initial soil evolution is a result of these processes and we 135 can consider how soil horizons reflect and record organic matter processing of the 136 developing fluvial landform (biogeomorphic succession), through its transition from a 137 barren sediment surface to a vegetated soil covered patch. As explained above, 138 organic matter may profoundly transform the local abiotic environment, increasing the 139 nutrient pool, ameliorating water and nutrient retention through soil aggregate and soil 140 141 structure formation, but also through production of humic acids, which may enhance weathering rates (Bätz et al., 2014a). 142

A Swiss braided river system, the Allondon River (Canton Geneva Switzerland), has been analysed for this study. We use a chronosequence (space for time substitution) approach, ranging from young surfaces close to the active zone of the river, to older stable floodplain terraces. On each area along the chronosequences, soil properties, 147 mainly in terms of grain size and organic matter quality, were analysed. Principal 148 components analysis is used to generalise the data obtained and to develop a model 149 for organic matter processing in braided rivers soils. Moreover, we try to identify the 150 time scales for soil formation and its link to the biogeomorphic succession.

151 **2. Material and Methods**

152 **2.1 Study site**

The gravel bed Allondon River is located to the west of Geneva (Switzerland). A large 153 part of the catchment is located in the calcareous French Jura Mountains. A number 154 of small (karstic) torrents flow from the Jura and combine into a single river at the 155 French/Swiss border. From this point, a 3km long reach of braided floodplain is formed 156 157 before its confluence with the Rhône River. The catchment area above the study reach is about 120 km² (FOEN, 2013). This reach, incised by about 60 m into fluvioglacial 158 159 sediments, overlays the Swiss Molassic basin. Fluvioglacial sediments were deposited during the last glacial cycle (the Würm and Riss glaciations), and their origin typical of 160 the Rhône basin geology (CJB (eds.), 1990; Coutterand, 2010). 161

162

Erosion into these fluvioglacial sediments (valley side slopes) is thought to lead to slope failures, exacerbated during localised saturation during storm events and river lateral undercutting processes. These are thought to be the main source of sediments in the braided reach. There are terraces of fluvial origin within the reach and the river has a potentially wide range of surface ages, ranging from active sites with a high turnover, to mature floodplain terraces, which are much older (Beechie et al., 2006). Following the biogeomorphic succession model proposed by Corenblit et al. (2009), there is evidence of rapid vegetation colonisation on exposed sediments of engineering species (*Salix elaeagnos, Salix purpurea*) and progressive plant facilitated stabilisation of some braid bar deposits which eventually lead to more stable fluvial landforms such as alluvial terraces (*Alnus glutinosa, Corylus avellana, Quercus robur, Fraxinus excelsior and Carpinus betulum*). However, terraces covered by dry grasslands can also be found (CJB (eds.), 1990). Moreover, there is clear evidence of both contemporary and historical soil development (e.g. buried soils).

177

The hydrology is pluvio-nival, having two maximum flood probabilities: (i) during spring 178 due to snowmelt and especially rain on snow in the Jura Mountains; and (ii) during 179 autumn, when heavy and prolonged rainfalls occur. The catchment hydrology 180 181 responds quickly, causing rapid hydrograph rise and high magnitude flood peaks, with return periods of 45 m³/s (2 years), 66 m³/s (5 years), 81 m³/s (10 years), and 123 m³/s 182 (50 years). Baseflow conditions are between 0.5 and 7.5 m³/s (FOEN, 2013; 183 Fourneaux, 1998). The river flow is also closely coupled to groundwater. There is clear 184 evidence of surface flow loss to groundwater in the upper part of the reach and the 185 return flow of calcareous groundwater to the main river in the lower part (e.g. 186 Fourneaux, 1998; Hottinger, 1998). 187

188

Land use in the catchment is mostly forest, prairies and pasture (70%), and agriculture (15%). Nevertheless, both industry and the CERN research centre use the water of the Allondon River and, despite extensive wastewater treatment, several polluting events impacted the river ecology between 1970 and 1990 (DIM, 2010). Generally, there is little river management within the 3 km reach considered in this study and most of the interventions (e.g. spur dykes) were removed in 2000 during a revitalisation and 195 renaturalisation programme. The spur dykes did not significantly hamper braiding processes of the sections studied, which are still very active. However, whilst still in a 196 braided/wandering state, there is evidence that the study reach is evolving from a 90m 197 wide bar braided system (1957) to a c. 40 m narrower river with vegetated islands. 198 These changes may be caused by land use changes of the alluvial terraces (pasture 199 to forest) and changes in the hydrological regime. Nowadays, the river corridor is 200 recognised both nationally (eg. Federal Inventory of Alluvial sites of National 201 Importance) and internationally as a protected site (DIM, 2010). 202

203

Figure 1 about here

205

2.2 Identification of sample sites

By means of an analysis of a series of historical aerial images, a Digital Elevation 206 Model of 1m resolution (DEM; SITG, 2012; SWISSTOPO, 2012), Electrical Resistivity 207 Analysis (ERT) and field observations (vegetation type and microtopography), four 208 chronosequences covering a large range of ages has been defined. The historical 209 aerial images and the DEM where used to reconstruct recent geomorphological 210 development and to trace fluvial landform boundaries. Additional field observations 211 212 have been integrated to validate the remote sensing data. Moreover, ERT data with 0.5m electrode spacing, following the method proposed by Laigre et al. (2013), have 213 been used to visualise belowground soil/sedimentary structures along each 214 chronosequence (see also Bätz et al. (2014b) for an example). Within each mapped 215 landform a representative site, in terms of microtopography, vegetation type and 216 soil/sediment thickness/structure has been chosen for detailed soil analysis. 217

The chronosequence with the sample numbers 11 to 16 includes an active zone and three terraces (Figure 1). The increasing altitude is an indication of increasing age of the surface. Sites 12 (1.3m elevation above the river), 11 (2.5m) and 13 (3.7m) are already visible in the oldest aerial image (1957 Figure 1), consequently, only a minimum age can be assigned of about 60 years for the lowest terrace (12).

Site 16 (1.1m) is part of a large mid-bar in 1999 (Figure 1) and experienced fast vegetation colonisation, potentially due to (fast resprouting) large woody debris. Due to its island character, we expect this landform to interact (engineering action) with the floods (geomorphological processes). Site 14 (0.4m) and 15 (0.3m) are the youngest sites on this chronosequence. In 2001 they appear to part of the main channel system. From 2005 both sites are exposed, but different rates of vegetation colonisation can be observed.

The chronosequence with the samples 31 to 34, 21 and 22, cover medium to long time 230 spans. 21 appears to be an old terrace at 2.2m from the river bed and is already visible 231 in the oldest aerial image (1957 in Figure 1). In the aerial image of 1980, sites 22 (0.8m) 232 and 31 (0.75m) are part of the main channel, while sites 34 (1.15m) and 33 (1.31m) 233 recently experienced an avulsion process and are in the former outer and inner 234 meander respectively. The island or cut-off terrace that has formed in this period has 235 been almost completely eroded during an avulsion between 1980-1996. The same 236 event created the base for the development of sites 31 (0.75m) and 32 (0.88m), which 237 in 1996 show already vegetation colonisation of the avulsion channel with 32 being in 238 239 the former inner and 31 in the former outer meander.

The shorter chronosequence of samples 41 to 43 cover mid to short time spans. In 1996 (Figure 1) site 41 appears as a barren surface whilst 42 shows already a pioneer vegetation cover. Based on the elevations above the river, which are respectively 1.1m for 41 and 1.59m for 42, and the avulsion pattern observed earlier for the sites 31-32, we might think of site 41 being part of former main channel and 42 being a former point bar. Site 43 (0.98m) has experienced a series of revegetation processes, but erosion
processes reset the local system in 2006.

The chronosequence 51 to 55 covers young time spans of a decade. These sites have 247 been continuously exposed to geomorphological disturbance and revegetation 248 processes (Figure 1). However, site 5.3 (1.77m) is the oldest site on this 249 chronosequence. In 2001 it is part of the main channel, but subsequent lateral erosion 250 exposed the site (2005 – Figure 1). Vegetation colonisation appears to be slow on this 251 site. The lateral erosion observed between 2001-2005 is also the starting point for the 252 development for sites 55 (1.23m) and 54 (1.43) as part of the point bar system. In 2006 253 both sites became isolated from the main channel by avulsion processes. Only larger 254 floods have impacted these two during recent years. Site 51 (0.86m) is still part of the 255 256 main channel system in 2009, while site 52 (1.45m) appears to be a mid-bar as part of a riffle system. Between 2009-2011 site 51 became exposed due to avulsion. 257

258

2.3 Soil sampling strategy

In each of the six areas, a soil pit was dug at a representative location to a minimum 259 depth of 50cm (P1 through P6). For each pit, soil/sedimentary layers were identified 260 and described using standard soil description methods. The following parameters were 261 recorded in the field: layer depth (H-layer), layer thickness, root density (nr./dm²) and 262 volume of soil stones (% > 2mm). Further data, such as the distance of each soil pit 263 from the main river along the most likely line of connection, were acquired using a 264 differential global positioning system (dGPS). For each layer in each profile, a sample 265 was taken for laboratory analyses in terms of grain size distribution (<2mm) and 266 organic matter quality, as explained in the next section. 116 soil/sediment samples 267 were obtained in total. 268

269

270

2.4 Laboratory analysis

Each of the soil/sediment samples was analysed using laser diffraction to determine the grain size distribution (Malvern Mastersizer 2000) and by pyrolysis to determine organic matter quality and pools (Rock-Eval 6).

274

For the grain size distribution, 1g of <2mm sieved material of the bulk sample was used to remove organic matter with a H_2O_2 solution (first at 15% then at 35% concentration), making sure that the pH did not drop below 3. Before performing particle size analysis, samples were dispersed in a calgon solution (sodium hexametaphosphate) for one night. Measurement outputs were grouped into 5 classes of apparent diameters: clay (0.01-4µm), silt (4-63µm), fine sand (63-250µm), medium sand (250-500µm), and coarse sand (500-1000µm).

282

In order to analyse the organic matter quality, the Rock-Eval method was followed (Disnar et al., 2003; Lafargue et al., 1998; Sebag et al., 2006). Here, only a short summary of the main principles and methodological steps is given (for details see Disnar et al., 2003 and Sebag et al., 2006). The analysis needs little pre-treatment of the samples: untreated bulk raw samples were sieved at <2mm and then grinded (with an agate mortar) to powder.

289

The principle of the Rock-Eval method lies in the fact that the quality of organic matter components is closely related to their thermal stability. Thus, in a first step, samples are slowly heated to 650° C in an inert atmosphere (N₂) – the pyrolysis step. Organic matter components gradually undergo cracking while hydrocarbon, CO and CO₂ emissions are continuously measured. Hydrocarbon emissions can then be plotted versus temperature (labelled as the S2 curve) and, using signal deconvolution, different proportions of organic matter pools can be identified: labile fresh litter (A1), stable litter components such as cellulose and lignin (A2), humified litter (namely humic and fulvic acids; A3), stable humus components such as humins (A4) and resistant humus components/geopolymers such as black carbon and charcoal (A5). Note, that these data are expressed as proportions (pools) of the fraction of Total Organic Carbon (TOC).

302

The temperature at which most hydrocarbons are liberated (TpS2) is determined using the S2 curve. TpS2 refers to the predominant organic matter pool in the sample. Following Sebag et al. (2006), two indices are used to describe the relationships between the five different organic matter pools :

$$I = \log\left(\frac{A1 + A2}{A3}\right)$$

308

309 *"I"* stands for immature and this I index represents the amount of fresh organic matter.310 Another index, "R" defined as:

[1]

[2]

311
$$R = \log\left(\frac{A3 + A4 + A5}{100}\right)$$

312

313 represents the thermo-resistant or more humified stable/resistant organic matter314 components (R stands for thermo-resistant).

315

In a second step, the residue of the pyrolysis is heated from 450°C to 750°C in an

317 oxygen enriched environment, allowing combustion of resistant organic matter. Again,

318 CO and CO₂ emissions are measured continuously.

From the combined pyrolysis and oxidation steps, fractions of soil organic (%TOC) and mineral carbon (%MINC), a Hydrogen Index (HI) and an Oxygen Index (OI) can be obtained. The higher the HI, the less organic matter is transformed (fresh litter). The OI expresses the oxidation rate of organic matter and thus the humified and more resistant organic matter pools.

325

326

2.5 Statistical analysis

The field and laboratory data are highly inter-correlated. Thus, they were analysed 327 using a principal component analysis (PCA; Matlab R2012b). A total of 22 variables 328 were available for the 116 samples. The five grain sizes and the five organic matter 329 pool variables, expressed as proportions, were first transformed by a centred log ratio 330 function into a new metric system to avoid problems of co-linearity and the closed data 331 effect. In a second step, all variables were standardized assuming a Gaussian 332 distribution of the data set. The PCA was performed using eigenvectors and 333 eigenvalues of the correlation matrix. The Pearson correlation coefficients between 334 principal components and variables were calculated by multiplying eigenvectors by the 335 square root of their associated eigenvalue. We set the level of significance (p=0.005) 336 for the correlation of the variables with the principal components at r>0.2540 based on 337 100 degrees of freedom. However, we also consider correlations r> 0.8 to be strongly 338 correlated with the related component. For the interpretation, we included all the 339 components explaining more than 5% of the total variance. 340

319

341 **3. Results**

Figure 1 shows the analysed profiles and their position in the floodplain. The related 342 data are summarised in Annexes A and B, while some examples of the analysed 343 profiles are given in Figure 2. Based on the USS Working Group WRB (2006), the old 344 terrace soils (11, 13, 21, 21) can be considered as mollic fluvisol skeletic, because they 345 have a think, organic matter enriched topsoil. The very young sites (14, 43, 51, 52) can 346 be defined as leptosol skeletic. Most mid aged soil profiles would best fit the name 347 fluvisol skeletic (15, 31, 32, 33, 34, 41, 42, 53, 54). However, profile 16 would be a 348 classical fluvisol and profile 55 also shows stagnic properties. 349

The vegetation of the terraces (11, 12, 13, 21) is dominated by *Corylus avellana*, *Quercus robur and Fraxinus excelsior*. Location 16 is dominated by *Alnus glutinosa* and some *Robinia pseudoacacia* stands, while site 22 is dominated by a *Populus alba* stand. Most of the very young stands (14, 43, 51) do not show significant vegetation cover – only a few sporadic grass stalks. Site 53 is covered by dry grass land. Sites 14, 52, 54, 55, 41, 32, 31, 33, 42, 34, are dominated by *Salix elaeagnos and/or Salix purpurea* with increasing ages (14 young and 34 the oldest stand).

357

358 Figure 1 and Figure 2 about here

359

Grain size distributions (<2mm) and organic matter quality and quantity, but also data related to the position in the landscape (see Annexes A and B) have been used as input variables for the principal component analysis (PCA). Some samples have low Total Organic Carbon values (TOC<0.02%; Annex B) and thus, this absolute value should be considered with caution (Disnar et al., 2003; Sebag et al., 2006). Nevertheless, these data are coherent with the conceptual model described below and, as such, were not excluded from the analysis: they simply indicate samples with
 exceptionally low organic matter content.

368

369 Table 1 and Figure 3 about here

370

The first four Principal Components (PC) explain 72.7% of the total variance (Table 1). 371 From the fifth PC on, explained total variance drops below 5% and so these PCs are 372 not considered further. The first PC explains 40.5% of the data variability (Table 1; 373 Figure 3A). It correlates positively with the proportions of clay, silt and fine sands, TOC, 374 with the proportions of more stable biopolymers and humified litter (A2 and A3), the 375 root density and to a lesser extent with the HI. It correlates negatively with the fractions 376 of coarse sands, the mineral carbon content (MINC), the OI, the resistant humus 377 components (A5) and to a lower degree with the proportion of medium sand and the 378 percentage of stones in the layer (%stones). Thus, the first PC appears to be 379 representing the level of pedogenesis in the system with higher levels of pedogenesis 380 associated with higher scores on this PC. Evidence for this is related to the higher 381 quantities of organic matter and related humification processes, as shown by the 382 presence of several organic matter decomposition stages (A2, A3 and to a lesser 383 extent A4) a key process in early pedogenesis. 384

385

The second PC explains less variance, 17.2% as compared with PC1 (Table 1; Figure 3A). It is mainly associated with stable humus components (A4%) the R-index and it is negatively associated with fresh organic matter input (A1%; I-Index), but also to a lesser extent with TOC and HI. Thus, PC2 appears to be a measure of the fraction of fresh organic matter (negative values) versus stable organic components (positive values): high values of PC2 describe a lack of fresh organic matter input and where
organic matter processing has transformed the available fresher pools towards more
stable ones (R-Index). It is surprising that the fresh and resistant organic matter
qualities (PC2) are, to a certain extent, independent from pedogenesis (PC1). This
might be due to the fact that there are two main sources for organic matter: fluvial input,
and endogenic vegetation production (e.g. Langhans et al., 2012; Naegeli, 1997; Pinay
et al., 1992; Steiger and Gurnell, 2003)

398

PC3 explains 8.1% (Table 1; Figure 3B) of the variance and is mainly correlated with layers that have high values of clay and silt. A negative correlation exists with the medium sand content. We suggest that this PC represents the sedimentary processes the layers are or have been exposed to. With positive values for environments with rather silt and clay rich deposits (e.g. shallow overland deposits on terraces, or plantcolonised old avulsion channels) and negative values for sites exposed to more powerful events.

406

The PC4, explains 6.9% of the total variance (Table 1) and correlates positively with 407 distance and elevation from the river. PC4 can be interpreted as representing the 408 chronosequence, with the older terrace profiles being higher and more distant from the 409 410 actual channel. However, it is interesting to note that this component actually explains very little of the total variance. We attribute this to the morphodynamics of the active 411 channel, e.g. avulsion processes, which prevent simple substitution of time by space 412 and create a complex mosaic of states of organic matter within the active zone. 413 Depending on flood magnitude and position in the landscape, landforms can have 414 different amounts and qualities of sedimentary inputs (Cabezas and Comín, 2010; 415

Steiger and Gurnell, 2003). It is only when a part of the active zone is stable forsufficient time that the effect of time becomes dominant.

418

Figure 3C plots the sample sites onto the first two PCs for each of the 3 analysed 419 zones. This plot spreads the samples based mainly on the organic matter quantity and 420 quality. Organic matter enriched topsoil layers (Ah layers) of the older terraces (notably 421 13, 12, 11, 22) are grouped into the lower right corner of the plot. However, also some 422 younger and fluvially exposed sites (54, 55) are found in this area of the plot. They are 423 characterized by high values on PC1, that is they tend to have higher proportions of 424 clay, silt and fine sands, TOC, higher proportions of the A2 and A3 organic matter 425 compounds and lower mineral content and a lower thermo-resistant organic matter 426 fraction (A5). They are also characterised by low values on PC2 that is they typically 427 have a high fraction of fresh organic matter (A1% and I-Index). We argue that these 428 samples are typical of the upper horizons of soils where pedogenesis processes tied 429 to organic matter processing are very active, with high in situ organic matter production, 430 which explains the relatively lower fraction of stable organic matter compounds. 431

432

Samples in the upper and lower left side of the plots shown in Figure 3C, tend to be 433 the subsurface layers (C horizons) of these older terrace soils (13, 12, 11, 22) but also 434 the C horizons of younger profiles (53, 54, 55, 14), with low values of PC1, that is 435 where TOC is low, with a high fraction of resistant organic components (OI and A5). 436 These samples distribute along the entire range of PC2, indicating that subsurface 437 layers are very variable in terms of fresh (negative PC2) versus stable humus 438 components (positive PC2). Whilst these sites may be beneath surface layers where 439 pedogenesis and in situ organic matter production is occurring, the young nature of 440

these alluvial soils means that their C horizons can still be clearly distinguished as 441 relatively inert. However, some M samples mix within this group. These are sandy 442 deposits mixed with some exogenous pedogenically altered organic matter. Due to 443 sediment transport processes only the most stable pools can be found in these layers, 444 because they are attached to finer grains (e.g. Asselman and Middelkoop, 1995; Pinay 445 et al., 1995; Steiger and Gurnell, 2003). However, because of the low TOC content on 446 the right side of the plot, the fractions of the different organic matter pools need to be 447 considered with caution. 448

449

Samples in the higher right corner of the plots shown in Figure 3C have high values on 450 PC1 and PC2: organic matter humification is active but, because of a lack or scarcity 451 of fresh organic matter input (A1), the organic matter pool tends to be biased towards 452 the mature and resistant components. The lack of fresh organic matter input and 453 organic matter humification has transformed the available labile matter (first A1 then 454 A2) to more stable compounds. Samples in this area of the plot include very young (2-455 5 years old) A horizons (51(Ai), 14(Ai)) and some Ah horizons of profiles that are still 456 impacted by shallow depositional events (41, 15, 22), mostly composed by humified 457 litter and stable humus components in combination with fine sands. Nevertheless, most 458 sites are related to the buried surface horizons (Ahb) and organic matter enriched 459 sandy deposits (M). The first (Ahb) have been progressively decoupled from fresh 460 organic matter input via burial, while the second (M) have been deposited together with 461 large quantities of stable humus components. 462

Summarising, sites close to positive values of PC2 can arrive through three
fundamentally different processes: (1) a stable surface deposit where pedogenesis
has formed an initial organic matter enriched topsoil after a period of stability with good

466 fresh organic matter input (transition from sediments into a soil); (2) a buried soil horizon and (3) deposits of upstream eroded organic matter enriched sands, with its 467 fresh organic matter supply being cut off and the accumulated organic matter 468 becoming progressively more resistant. Thus, even though the soil horizons show 469 similar properties in this zone, their evolution may be very different. Indeed, the 470 accumulation of stable humus compounds in Ahb layers is due to recessive pedogenic 471 processes, which lead to degradation of the organic matter (Johnson, 1985; Schaetzl 472 and Anderson, 2005). On the other hand, the initial accumulation of stable humus 473 compounds in layer Ai is due to progressive pedogenesis, in which the sparse 474 vegetation cover and/or fluvial litter input provides easily and fast transformable 475 organic matter (Gregory et al., 1991; Langhans et al., 2012; Sebag et al., 2006). 476

477

When plotting the samples on PC1 versus PC3 (Figure 3D) the entire range of 478 sedimentary composition gets clear and can almost be read like a classical grain size 479 triangle. The organic matter enriched topsoils, which are placed on the positive axe of 480 PC1, show high variability in clay to medium sand content (PC2), but with initial A 481 horizons (55Ai, 51(Ai), 42 Ai(M), 14(Ai)C1, 16Ai) mainly placed in the upper right part 482 of the plot (high fine sand content). The buried soil horizons (Ahb) distribute, as the Ah 483 on the entire right part of the plot. Most of the M horizons, as being organic matter 484 485 enriched sandy deposits, concentrate in the fine to medium sand range.

486

The plot PC1/PC4 of the samples (not shown here) show the distribution of sedimentary and /or soil layers along the chronosequence, with samples being older on the positive axis of PC4. The plot is not shown here because it simply reflects the chosen sampling strategy along the chronosequences.

491 **4. Discussion**

492 **Pedo-geomorphic interactions of the Allondon River**

The shape of figure 3C need special mention: as PC1 is defined as an axis of 493 pedogenesis, as we move from the higher middle of the point cloud towards higher 494 organic matter contents (from the left to the right), there is simultaneously an increase 495 in amount of fresh organic matter supply (negative values of PC2) because of the 496 establishment of vegetation with higher biomass productivity (eg. Van Breemen and 497 Finzi, 1998). This is supported by the correlation with the root density. It also suggests 498 higher organic matter humification rates indicated by the presence of all the organic 499 matter humification steps (A1-A2-A3). However, the same change in organic matter 500 supply can be observed if moving towards negative values on PC1. Although, this 501 increase in fresh organic matter supply is characterised by no increase in TOC and 502 503 related humification steps (pools A2 and A3). Samples located in the lower right corner (14 and 22) tend to be geomorphic stable but still influenced by the fluctuating ground 504 water (elevation from river is respectively 0.9m and 1.2m). Possible sources are, in the 505 case of the terrace soils, illuviation products from the upper Ah horizons, while for 506 younger sites dissolved organic matter from the fast infiltrating (gravelly material) river 507 flow and related processes of the hyporheic zone may be a potential source. 508

It is interesting to notice that there is no clear correlation of topsoil maturity, thus moving from the top of the point cloud of Figure 3C towards the lower right corner, and landform age observed form the historical aerial images. We can find young sites with about one decade of development (54), mid aged (15-30 years) sites (22, 32, 33) and old terrace soils (13, 11, 21) in the same area of the plot. In connection with the low degree of explained variance of PC4, representing the chronosequence, this indicates that there is no clear impact of time on young alluvial soils. Local and reach scale geomorphological setting in conjunction with different flood magnitudes, create a
complex pattern of geomorphological impact (Steiger and Gurnell, 2003), thus directing
local soil evolution trajectory (Johnson, 1985; Schaetzl and Anderson, 2005).

519

The shape of Figure 3D does not show a clear pattern. However we can notice the 520 affinity of fine sand, silt and clay for higher TOC as observed in other studies (e.g. 521 Cabezas et al., 2010; Pinay et al., 1995). Moreover, the old terrace topsoil appears to 522 have higher contents of fines (higher right corner), while younger sites are rather 523 composed of medium to fine sands (lower right corner), indicating the kind of 524 depositional processes each site is/has been exposed to. The C horizons, located in 525 the left part of the plot (Figure 3D) are generally high in coarse sands, but show a wide 526 range of silt/clay versus medium sand content. It is interesting to notice that there is a 527 trend: section 2-3 and 4-5 have higher contents of silt and clay if compared to the C 528 layers of section 1. 529

530

Based on the above interpretation, we can conceptualise braided river soil evolution 531 (Figure 4). Pedogenesis starts from a large range of deposits in terms of grain size and 532 organic matter quantity and quality (top of the PC1/PC2 plot – Figure 3C). If conditions 533 are favourable (vegetation establishment, water and nutrient supply from low 534 magnitude depositions events) a soil starts to form (towards the lower right corner of 535 the PC1/PC2 plot – Figure 3C) and eventually form a mature terrace soil (Figure 4). 536 However, young deposits can also get buried before pedogenic alteration occurred due 537 to high magnitude events (taking a position in the left corner of the point cloud). Burial 538 or erosion and re-deposition of pedogenically altered material can also occur whilst 539 following the pedogenic trajectory, indicated by the presence of Ahb (buried surface 540

541 horizons) and M (organic matter enriched sandy deposits) horizons in the right lower to mid right part of Figure 3C. Burial leads to the ageing of the organic matter pools of 542 the ancient surface (Ahb) (e.g. Schaetzl and Anderson, 2005). The rates of ageing, 543 thus following positive PC2 (A4 and R-Index), diminish in relation to the availability of 544 unstable organic matter compounds and because of the missing fresh organic matter 545 input. The thick organic matter enriched sandy deposits (M) may also contain rather 546 stable compounds, because they tend to be attached to sand particles (e.g. Asselman 547 and Middelkoop, 1995; Pinay et al., 1995; Steiger and Gurnell, 2003). 548

The geomorphic activity in this area of the plot suggests that there are close 549 interactions between geomorphic processes and braided alluvial soil evolution. 550 Especially during early stages of braided alluvial pedogenesis these appear critical. 551 552 Major geomorphological events might reset the system, through either deposition or erosion. However, weaker deposition events may supply organic matter enriched fine 553 sediments (Cabezas and Comín, 2010; Langhans et al., 2012; Pinay et al., 1992; 554 Steiger and Gurnell, 2003) forming cumulic soils (Daniels, 2003; Jacobson et al., 2005), 555 and act as an exogenic organic matter source that can accelerate initial ecosystem 556 processes (Bätz et al., 2014a). This processes can potentially be facilitated by the 557 engineering action of vegetation, thus during the biogeomorphic phase proposed in the 558 biogeomorphic succession model of Corenblit et al. (2009). 559

560

Bätz et al., (2014b) reconstructed the geomorphological, soil evolutionary and vegetation successional development of a section which includes profile 15 of this study. High temporal frequency historical images were used to infer past geomorphological changes. These were combined with a 2D grain size distribution model (Electrical Resistivity Analysis - ERT) of the entire section and soil profile 15 of

this study to deduce soil maturity and distribution. The analysed section is 566 characterised by two distinguished areas. A first low altitude zone with shrubby willow 567 stands and a soil covered surface (Ah/M/C profile); this site can be attributed to the 568 biogeomorphic phase proposed in the biogeomorphic succession model of Corenblit 569 et al. (2009). A second higher elevated gravel/cobble barren surface with only a few 570 grass stalks with no soil. The historical images show that the barren higher located 571 area formed about 10 years before sampling. The lower located soil covered area was 572 been created 8 years before sampling and experienced sand aggradation over the 573 entire period due to the rapid colonisation of river engineering species. If the deposit is 574 not too thick it can be integrated into the soil (Daniels, 2003; Jacobson et al., 2005), 575 then the organic matter component of the deposit may act as an exogenic input. The 576 close interaction between geomorphological processes (fine sediment and resources 577 supply such as organic matter), vegetation development (trapping and fixing sediments 578 with its biomass but also producing fresh litter) and soil processes (nutrient 579 transformation and storage) has led to the rapid development of a more productive 580 local ecosystem. Conversely the older site, which was cut-off from the river supply 581 early, experienced a slow development (Bätz et al., 2014b). This close interaction 582 between these three components during fluvial landform formation has been defined 583 as a coevolutive process (Bätz et al., 2014a; Corenblit et al., 2014, 2009). The time 584 scale of coevolution of the two zones changes due to different rates and forms of 585 interaction between geomorphic processes (deposit quality - Asselman and 586 Middelkoop, 1995; Langhans et al., 2012; Pinay et al., 1995, 1992; Steiger et al., 2001; 587 Steiger and Gurnell, 2003), vegetation colonisation (Corenblit et al., 2011, 2009; 588 Gurnell, 2014; Gurnell et al., 2012) and soil evolution (Bätz et al., 2014a, 2014b; 589 Langhans et al., 2012; Mardhiah et al., 2014). 590

Similar observations can be made in the point distribution of Figure 3C. Samples 54Ah 592 and 55Ai, with an approximate age of 10 years, estimated by a series of historical 593 images, show similar properties (OM quality/ quantity and grain size) as mature 594 terraces soils (13Ah, 11Ah, 21Ah, >50 years), thus indicating the changes in speed of 595 the biogeomorphic succession due to pedo-biogeomorphic feedbacks. Thus, 596 sedimentation processes appear to either constrain or contribute to initial soil 597 development by facilitating the accumulation of an organic matter stock before isolation 598 from the river flux (shown by the intensity triangles in Figure 4). Initial deposits are 599 commonly very low in organic matter content (Figure 3 – annexe A). Fluvial fresh input 600 of exogenic organic matter may then be a main source. The young deposits in Figure 601 602 3, are exposed to (weak) fluvial sedimentary processes. The associated organic matter deposition ages quickly because of the limited fresh input and the relatively high 603 biological activity of alluvial environments (Bullinger-Weber et al., 2007; Guenat et al., 604 1999; Langhans et al., 2012; Pusch et al., 1998). The more stable litter compounds, 605 such as lignin (A2), accumulate in the topsoil with time (Doering et al., 2011; Langhans 606 et al., 2012) because transformed more slowly into more stable compounds such as 607 humic and fulvic acids and humins (A3+A4) - higher right part of Figure 3C. Moreover, 608 the deposited material, as the M horizon shows, has a high fraction of more stable and 609 resistant humus components (A4 and A5). However, when plants establish and the 610 biomass production increases, the organic matter transformation chain is established 611 and organic matter pools arrange in the order A2-A3-A4 (Annex B), as the position of 612 mature terrace soils in the lower right part of Figure 3C suggests. 613

614

591

615 Plant colonisation, is particularly dependent on ambient and geomorphic conditions (Cierjacks et al., 2011; Francis, 2007; Gurnell et al., 2012; Hupp and Rinaldi, 2007). 616 These might be favourable from the beginning (e.g. access to water in a gravely 617 618 environment) but can also develop within a few decades (e.g. water retention through fine sand or silt deposition), changing the timing in the shift of main fresh organic matter 619 input (exogenic to endogenic as indicated by the intensity triangles in Figure 4). This 620 means that there are locations in which soil evolution can be fast (high exogenic and 621 endogenic input) and others which develop relatively slowly (low exogenic and 622 endogenic input), thus changing the time scale and pathway of landform coevolution 623 processes in gravelly braided river systems. 624

625

626 **Comparison with other river systems and future research challenges**

Clearly a major question arises from this research: even though the amounts of total 627 organic matter are relatively low in the most recent deposits (Annexe B) as compared 628 with those sites where in situ production can occur (Cierjacks et al., 2011), are these 629 sufficient to enhance the rate of initial soil-forming processes? Tabacchi et al. (2000) 630 noted that the riparian corridor may be seen as a recycling zone of exogenous inputs 631 coming from the entire upland catchment. They argue that there will be a dependence 632 of fluvial habitats on such organic matter supplies, with supply closely related to 633 geomorphological riverine processes. Cierjacks et al. (2011) found that the difference 634 in organic matter quantity and quality found in alluvial soils of the Danube River (Austria) 635 are related to flooding history. Floods can import significant amounts of particulate OM 636 (Hein et al., 2003). Based on sediment trap analysis of meandering rivers Garonne 637 (France) and Severn (UK), Steiger et al. (2001) and Steiger and Gurnell (2003), found 638 a strong relationship between the quantity of deposit fine grain sizes (silt/clay) and 639

640 quantity of TOC and organic nitrogen. Only phosphorous showed a dependence on the quantity of deposition, regardless of its grain size distribution. This research also 641 emphasises the importance of the geomorphological setting and flood magnitude in 642 643 defining the spatial distribution of deposits and related quality (grain size, and the elements NPC) confirming findings of Pinay et al. (1995, 1992). Low magnitude events 644 in constrained meandering river reaches, show deposition peaks on landforms closely 645 located to the main channel (e.g. point bars). As flood magnitude increases, the 646 deposition peak shifts towards higher located and distal fluvial landforms (e.g. side 647 channels, higher benches). However, less constrained and more geomorphologically 648 complex river reaches will have a less clear shift in the sedimentation peak (Steiger et 649 al., 2001; Steiger and Gurnell, 2003). 650

Cabezas et al. (2010) and Cabezas and Comín (2010) analysed the spatial pattern of 651 deposition in terms of OM quantity/quality and grain size for different landforms of the 652 meandering Ebro River (Spain). Similar to this research historical aerial images were 653 used to estimate landform age and to deduce hydrological connectivity. Moreover, 654 sediment traps were used to quantify fluvial deposits. Results indicated that on old 655 sites, endogenic OM production dominates, while on younger sites river depositional 656 events were the main source of OM. Their results also show that the exogenic input is 657 also dependent on the position in the fluvial landscape, with higher values for fluvially 658 659 exposed sites (point bars compared to side channels), and from the channel-floodplain morphology. However, in their research, time since formation was well-correlated with 660 net carbon accumulation (Cabezas et al., 2010; Cabezas and Comín, 2010). This 661 might be related to the higher geomorphological activity of braided river (e.g. avulsion). 662

663

664 Bechtold and Naiman (2009) developed a modelling approach, which combines the CENTURY model for predicting changes in the soil organic matter pools with a simple 665 fluvial deposition model and a forest growth/production model. The model was tested 666 for the meandering Queens River (Washington - USA). Results confirmed that fluvial 667 OM, but also the deposition of fines, are especially important during the first decades 668 of fluvial landform development, because they influence moisture retention and nutrient 669 regime. These results are deemed to be less clear in higher energy systems (Bechtold 670 and Naiman, 2009). 671

672

Similar observations have been made by Doering et al. (2011) and Langhans et al. 673 (2012) for the braided Tagliamento River (Italy). Riparian forest and especially 674 vegetated islands have a high productivity. Pre-processed fresh organic matter can 675 then be exported from these sites by wind and fluvial processes and deposited on less 676 productive sites such as river bars. This influx could provide a high quality resource 677 (Doering et al., 2011; Langhans et al., 2012) to initialise ecosystem processes such as 678 soil forming processes, and notably to enhance local weathering processes and plant 679 nutrient availability due to the decomposition processes. Naegeli (1997) has studied 680 the spatial and temporal variability of particular organic matter in gravel bed river 681 deposits of the Necker River in Switzerland. The spatial and temporal variations of 682 particular organic matter stocks were found to be closely related to river dynamics. 683 This resource reservoir can be activated and transformed by microbial activity (Bridge, 684 1993; Doering et al., 2011; Gregory et al., 1991; Naegeli, 1997; Pusch et al., 1998; 685 Uehlinger, 2000). Especially, earthworms, bacteria, algae and biofilms can transform 686 such organic matter inputs and make resources available for plants (Bätz et al., 2014a; 687 Bullinger-Weber et al., 2007; Pusch et al., 1998). Even if sediments free of fresh 688

organic matter are considered to be highly reactive (as for those found on glacier
forefields for instance; Burga et al., 2010; Mavris et al., 2010), the organic matter found
in alluvial deposits will provide an even more accessible resource (Bardgett et al., 2007;
Gregory et al., 1991; Guelland et al., 2013), potentially functioning as a start-up for
plants and thus the biogeomorphological succession. The mycorrhizal fungi symbiosis
that *Salicaceae* can establish may further promote resource up-take from the sediment,
soil and groundwater stock (Harner et al., 2011, 2010).

696

Whilst the dynamics of organic matter in fluvial systems have been addressed, there 697 are fewer studies that investigated the extent to which it is important or, more precisely, 698 the conditions under which it is important. For instance, if the source of exogenous 699 organic matter is the erosion of river banks and/or terraces, then this will require the 700 highest flows. Nevertheless, these are the flows that are also likely to mobilise large 701 amounts of sediments, so diluting the organic matter concentration. The large 702 variations of initial substrate (C and M) of our analysis show the large variability of 703 fluvial inputs in terms of grain size distribution and organic matter quality and quantity. 704 Additionally the amount and frequency of the deposition determines whether or not the 705 active river surface aggrades or is even buried (Bätz et al., 2014a). We know very little 706 about organic matter delivery in braided rivers during flood events and its relationship 707 708 to sediment delivery rates and thus the balance between the benefits (organic matter delivery) and risks (burial) of deposition. If it can be shown that exogenous organic 709 matter does enhance pedogenesis, even at relatively low concentrations, then we may 710 conclude that geomorphological processes not only disturb negatively the soil-711 vegetation development (through erosion for instance), but may also facilitate, or even 712 trigger, more rapid plant colonisation in braided rivers. In turn, because of the 713

engineering action of plants (Gurnell, 2014; Gurnell et al., 2012) overall landform
development speed and trajectory (biogeomorphic succession *sensu* Corenblit et al.
(2014, 2009) might be affected.

717

A second broad group of unanswered questions relates to the likelihood of erosion and 718 burial, which will equally locally reset the surface organic matter dynamics. Deposition 719 is an interesting process, because it may have a range of competing effects. On the 720 one hand, it may progressively aggrade the landform surface and so isolate it from the 721 active river channel (Daniels, 2003; Jacobson et al., 2005). This may make it more 722 stable, inundated less frequently and, provided water supply and nutrient stock does 723 724 not become a limiting factor, colonised by plants. It is established that typical early 725 colonisers are able to adapt their root network so as to avoid water logging whilst also follow the falling water table to avoid drying (Glenz et al., 2006; Pasquale et al., 2012). 726 But if the deposition is too great, or it involves grains that are too coarse, this may 727 prevent plant recovery and/or development as well as isolate the ancient topsoil from 728 the potential benefits of exogenous organic matter input. Indeed, Gurnell and Petts 729 (2002) proposed an optimal range of (fine) sediment aggradation and vegetation 730 recruitment types (propagules, resprouting of woody debris) that promote rapid fluvial 731 landform development. Similar for the development of a topsoil in a fluvial setting: only 732 733 deposition events that are within a certain magnitude, frequency and quality (grain size and organic matter) can be integrated into cumulic topsoil (Bätz et al., 2014a; Daniels, 734 2003; Steiger and Gurnell, 2003). 735

736

In relation to erosion, sites closer to the riverine matter flux, will also be exposed to a
certain risk; soil profile 15 (Bätz et al., 2014b), for instance, has now been removed by

739 a later channel migration during a large flood in winter 2013. Whilst such a site may have higher resource availability due to riverine matter fluxes, thus a faster pedo-740 biogeomorphic evolution, it also bears a higher risk of being eroded and destroyed. 741 Hence, braided rivers morphodynamics can be described as a continuous battle 742 between destructive processes (e.g. powerful floods) and stabilising processes 743 (aggradation, vegetation establishment and soil matter transformation) that coevolve 744 (from a landform development perspective) as fast as possible to prevent damage 745 and/or destruction. 746

747 5. Conclusion

In this paper, we have considered the relationship between geomorphic processes and 748 organic matter processing (humification) in relation to initial braided alluvial soil 749 750 evolution. Soils form as an emergent property and are deemed to represent the state of ecosystem organic matter processing. In the field, identified soil and sedimentary 751 layers were sampled and analysed in the laboratory for organic matter quality and 752 quantity (Rock-Eval method) and grain size distribution <2mm. Additionally, variables 753 describing the position of the layers in the fluvial landscape, and the rooting density 754 were added to the dataset. To explore possible relations, a principal components 755 analysis has been performed. Results lead to three main conclusions: 756

- the amount and quality of organic matter, as well as grain size distributions, may
 be key variables in understanding pedo-biogeomorphological feedbacks;
- results suggest that geomorphological processes may add organic matter and
 fine grained sediments to the normally inert sedimentary sites and change the
 rate of fluvial landform development (biogeomorphic succession *sensu* Corenblit et al., 2014, 2009). Research needs to establish the extent to which

this additional resource can facilitate initial fluvial landform ecosystem matter
turnover processes and soil formation. The enhanced overall ambient
conditions (enhanced nutrient and water availability), potentially promotes
vegetation growth/succession and eventually may lead to a river independent
and self-sustained fluvial landform (e.g. river terrace).

3. data suggest that, during the coevolution mentioned above, there might be a
shift from small amounts of external organic matter supply during fluvial
deposition (e.g. bars) to large in situ production due to vegetation colonisation
and related biomass production (terraces). The extent to which these two
organic matter sources interact, changes biogeomorphic succession time
scales and pathways.

Although the data show an impact of organic matter dynamics on fluvial landform development, the spatial and temporal variability of these processes is not yet clear. Our future work will focus on quantifying the impact of initial organic matter supply on the vegetation-soil system and its relation to fluvial geomorphic processes. Moreover, the general validity of the concept should be investigated, by trying to identify these processes in other river systems but also, for instance, in alluvial fans.

780 Acknowledgments

Special thanks to B. Lovis, G. Grob, J. Gerber, L. Laigre, J.-B. Bosson, N. Deluigi, D. Balin, J. Roberts, J. Heyman, M. Bühler, F. Dietrich, N. Diaz, T. Adatte, H-M. Saleh and B. Putlitz for help in the field and laboratory; N. Micheletti for helping in the orthorectification procedure of the historical aerial images; H.-R. Pfeifer for information and discussions on the Allondon River; Canton Geneva - Department of Nature and Landscape for data and access to the natural reserve for research purposes; and

- 787 Canton Vaud for funding the research. Two reviewers provided very valuable
- comments on an earlier draft of this paper.

789 **References**

- Asselman, N.E.M., Middelkoop, H., 1995. Floodplain sedimentation: Quantities, patterns and
 processes. Earth Surf. Process. Landf. 20, 481–499. doi:10.1002/esp.3290200602
- Bardgett, R.D., Richter, A., Bol, R., Garnett, M.H., Bäumler, R., Xu, X., Lopez-Capel, E.,
 Manning, D.A.C., Hobbs, P.J., Hartley, I.R., Wanek, W., 2007. Heterotrophic microbial
 communities use ancient carbon following glacial retreat. Biol. Lett. 3, 487–490.
 doi:10.1098/rsbl.2007.0242
- Bätz, N., Verrecchia, E.P., Lane, S.N., 2014a. The role of soil in vegetated gravelly river braid
 plains: more than just a passive response? Earth Surf. Process. Landf. n/a–n/a.
 doi:10.1002/esp.3631
- Bätz, N., Verrecchia, E.P., Lane, S.N., 2014b. Soil in braided rivers: an overlooked component
 of braided river morphodynamics, in: River Flow 2014.
- Bechtold, J.S., Naiman, R.J., 2009. A Quantitative Model of Soil Organic Matter Accumulation
 During Floodplain Primary Succession. Ecosystems 12, 1352–1368.
 doi:10.1007/s10021-009-9294-9
- Beechie, T.J., Liermann, M., Pollock, M.M., Baker, S., Davies, J., 2006. Channel pattern and
 river-floodplain dynamics in forested mountain river systems. Geomorphology 78, 124–
 141. doi:10.1016/j.geomorph.2006.01.030
- Bernasconi, S.M., Biglink Project Members, 2008. Weathering, soil formation and initial
 ecosystem evolution on a glacier forefield: a case study from the Damma Glacier,
 Switzerland. Mineral. Mag. 72, 19–22. doi:10.1180/minmag.2008.072.1.19
- Bridge, J.S., 1993. The interaction between channel geometry, water flow, sediment transport
 and deposition in braided rivers. Geol. Soc. Lond. Spec. Publ. 75, 13–71.
 doi:10.1144/GSL.SP.1993.075.01.02
- Bullinger-Weber, G., Le Bayon, R.-C., Guenat, C., Gobat, J.-M., 2007. Influence of some physicochemical and biological parameters on soil structure formation in alluvial soils.
 Eur. J. Soil Biol. 43, 57–70. doi:10.1016/j.ejsobi.2006.05.003
- Burga, C.A., Krüsi, B., Egli, M., Wernli, M., Elsener, S., Ziefle, M., Fischer, T., Mavris, C.,
 2010. Plant succession and soil development on the foreland of the Morteratsch glacier
 (Pontresina, Switzerland): Straight forward or chaotic? Flora Morphol. Distrib. Funct.
 Ecol. Plants 205, 561–576. doi:10.1016/j.flora.2009.10.001
- Cabezas, A., Angulo-Martínez, M., Gonzalez-Sanchís, M., Jimenez, J.J., Comín, F.A., 2010.
 Spatial variability in floodplain sedimentation: the use of generalized linear mixedeffects models. Hydrol. Earth Syst. Sci. 14, 1655–1668. doi:10.5194/hess-14-16552010
- Cabezas, A., Comín, F.A., 2010. Carbon and nitrogen accretion in the topsoil of the Middle
 Ebro River Floodplains (NE Spain): Implications for their ecological restoration. Ecol.
 Eng., Carbon, nutrient and metal retention in wetlands in a restoration context 36, 640–
 652. doi:10.1016/j.ecoleng.2008.07.021
- Cierjacks, A., Kleinschmit, B., Babinsky, M., Kleinschroth, F., Markert, A., Menzel, M.,
 Ziechmann, U., Schiller, T., Graf, M., Lang, F., 2010. Carbon stocks of soil and

- vegetation on Danubian floodplains. J. Plant Nutr. Soil Sci. 173, 644–653.
 doi:10.1002/jpln.200900209
- Cierjacks, A., Kleinschmit, B., Kowarik, I., Graf, M., Lang, F., 2011. Organic matter
 distribution in floodplains can be predicted using spatial and vegetation structure data.
 River Res. Appl. 27, 1048–1057. doi:10.1002/rra.1409
- CJB (eds.), 1990. Editions des Conservatoire et Jardin botaniques de la Ville de Genève (CJB);
 Sentier naturaliste, Vallon de la Roulavaz: Dardagny.
- Corenblit, D., Baas, A.C.W., Bornette, G., Darrozes, J., Delmotte, S., Francis, R.A., Gurnell, 837 A.M., Julien, F., Naiman, R.J., Steiger, J., 2011. Feedbacks between geomorphology 838 839 and biota controlling Earth surface processes and landforms: A review of foundation Rev. concepts and current understandings. Earth-Sci. 106. 307-331. 840 doi:10.1016/j.earscirev.2011.03.002 841
- Corenblit, D., Steiger, J., González, E., Gurnell, A.M., Charrier, G., Darrozes, J., Dousseau, J.,
 Julien, F., Lambs, L., Larrue, S., Roussel, E., Vautier, F., Voldoire, O., 2014. The
 biogeomorphological life cycle of poplars during the fluvial biogeomorphological
 succession: a special focus on Populus nigra L. Earth Surf. Process. Landf. 39, 546–567.
 doi:10.1002/esp.3515
- Corenblit, D., Steiger, J., Gurnell, A.M., Tabacchi, E., Roques, L., 2009. Control of sediment
 dynamics by vegetation as a key function driving biogeomorphic succession within
 fluvial corridors. Earth Surf. Process. Landf. 34, 1790–1810. doi:10.1002/esp.1876
- Coutterand, S., 2010. Etude géomorphologique des flux glaciaires dans les alpes nordoccidentales au pléistocène récent Du maximum de la dernièr glaciation aux premières
 étapes de la déglaciation. l'Université de Savoie, Le Bourget du Lac.
- Crouzy, B., Perona, P., 2012. Biomass selection by floods and related timescales. Part 2:
 Stochastic modeling. Adv. Water Resour. 39, 97–105.
 doi:10.1016/j.advwatres.2011.09.018
- Baniels, J.M., 2003. Floodplain aggradation and pedogenesis in a semiarid environment.
 Geomorphology 56, 225–242. doi:10.1016/S0169-555X(03)00153-3
- BIM, 2010. Département de l'Intérieur et de la mobilité (DIM); L'Allondon Fiche-rivière n.
 1 4 édition.
- Bisnar, J.R., Guillet, B., Keravis, D., Di-Giovanni, C., Sebag, D., 2003. Soil organic matter
 (SOM) characterization by Rock-Eval pyrolysis: scope and limitations. Org. Geochem.
 34, 327–343. doi:10.1016/S0146-6380(02)00239-5
- B63 Doering, M., Uehlinger, U., Ackermann, T., Woodtli, M., Tockner, K., 2011. Spatiotemporal
 heterogeneity of soil and sediment respiration in a river-floodplain mosaic (Tagliamento,
 NE Italy). Freshw. Biol. 56, 1297–1311. doi:10.1111/j.1365-2427.2011.02569.x
- FOEN, 2013. Environmental data from the Federal Office for the Environment (FOEN).
- Fourneaux, J.C., 1998. Analyse du fonctionnement hydrogeologique du bassin de l'Allondon.
 Arch. Sci.-Geneve 171–183.
- Francis, R.A., 2007. Size and position matter: riparian plant establishment from fluvially
 deposited trees. Earth Surf. Process. Landf. 32, 1239–1243. doi:10.1002/esp.1557
- Francis, R.A., Corenblit, D., Edwards, P., 2009. Perspectives on biogeomorphology, ecosystem
 engineering and self-organisation in island-braided fluvial ecosystems. Aquat. Sci. Res.
 Boundaries 71, 290–304–304.
- Francis, R.A., Tibaldeschi, P., McDougall, L., 2008. Fluvially-deposited large wood and
 riparian plant diversity. Wetl. Ecol. Manag. 16, 371–382. doi:10.1007/s11273-0079074-2
- 877 Gerrard, J. (Ed.), 1987. Alluvial soils, Van Nostrand Reinhold Soil Science Series. Van
 878 Nostrand Reinhold, New York.

- Glenz, C., Schlaepfer, R., Iorgulescu, I., Kienast, F., 2006. Flooding tolerance of Central
 European tree and shrub species. For. Ecol. Manag. 235, 1–13.
 doi:10.1016/j.foreco.2006.05.065
- Gregory, S.V., Swanson, F.J., McKee, W.A., Cummins, K.W., 1991. An Ecosystem
 Perspective of Riparian Zones. BioScience 41, 540–551. doi:10.2307/1311607
- Guelland, K., Hagedorn, F., Smittenberg, R.H., Göransson, H., Bernasconi, S.M., Hajdas, I.,
 Kretzschmar, R., 2013. Evolution of carbon fluxes during initial soil formation along
 the forefield of Damma glacier, Switzerland. Biogeochemistry 113, 545–561.
 doi:10.1007/s10533-012-9785-1
- Guenat, C., Bureau, F., Weber, G., Toutain, F., 1999. Initial stages of soil formation in a riparian
 zone: Importance of biological agents and lithogenic inheritance in the development of
 the soil structure. Eur. J. Soil Biol. 35, 153–161. doi:10.1016/S1164-5563(10)70001-7
- Guex, D., Weber, G., Musy, A., Gobat, J.-M., 2003. Evolution of a swiss alpine floodplain over
 the last 150 years: hydrological and pedologocal considerations. Presented at the
 International conference "Towards natural flood reduction strategies", 6-13 September
 2003, Warsaw.
- Gurnell, A.M., 2014. Plants as river system engineers. Earth Surf. Process. Landf. 39, 4–25.
 doi:10.1002/esp.3397
- Gurnell, A.M., Bertoldi, W., Corenblit, D., 2012. Changing river channels: The roles of
 hydrological processes, plants and pioneer fluvial landforms in humid temperate, mixed
 load, gravel bed rivers. Earth-Sci. Rev. 111, 129–141.
 doi:10.1016/j.earscirev.2011.11.005
- Gurnell, A.M., Petts, G.E., 2002. Island-dominated landscapes of large floodplain rivers, a
 European perspective. Freshw. Biol. 47, 581–600.
- Gurnell, A.M., Petts, G.E., Hannah, D.M., Smith, B.P.G., Edwards, P.J., Kollmann, J., Ward,
 J.V., Tockner, K., 2001. Riparian vegetation and island formation along the gravel-bed
 Fiume Tagliamento, Italy. Earth Surf. Process. Landf. 26, 31–62. doi:10.1002/10969837(200101)26:1<31::AID-ESP155>3.0.CO;2-Y
- Harner, M.J., Mummey, D.L., Stanford, J.A., Rillig, M.C., 2010. Arbuscular mycorrhizal fungi
 enhance spotted knapweed growth across a riparian chronosequence. Biol. Invasions 12,
 1481–1490. doi:10.1007/s10530-009-9559-4
- Harner, M.J., Opitz, N., Geluso, K., Tockner, K., Rillig, M.C., 2011. Arbuscular mycorrhizal
 fungi on developing islands within a dynamic river floodplain: an investigation across
 successional gradients and soil depth. Aquat. Sci. 73, 35–42. doi:10.1007/s00027-0100157-4
- Hein, T., Baranyi, C., Herndl, G.J., Wanek, W., Schiemer, F., 2003. Allochthonous and autochthonous particulate organic matter in floodplains of the River Danube: the importance of hydrological connectivity. Freshw. Biol. 48, 220–232. doi:10.1046/j.1365-2427.2003.00981.x
- Hottinger, M., 1998. Etude hydrogéologique de la nappe alluviale de l'Allondon genevoise
 (GE). Centre d'hydrogéologie Université de Neuchâtel.
- Hupp, C.R., Rinaldi, M., 2007. Riparian Vegetation Patterns in Relation to Fluvial Landforms
 and Channel Evolution Along Selected Rivers of Tuscany (Central Italy). Ann. Assoc.
 Am. Geogr. 97, 12–30. doi:10.1111/j.1467-8306.2007.00521.x
- IUSS Working Group WRB, 2006. World reference base for soil resources 2006, 2nd edition.
 ed. IUSS Working Group WRB, FAO, Rome.
- Jacobson, R.B., O'Connor, J.E., Oguchi, T., 2005. Surficial Geologic Tools in Fluvial
 Geomorphology, in: Kondolf, G.M., Piégay, H. (Eds.), Tools in Fluvial Geomorphology.
 John Wiley & Sons, Ltd, pp. 23–57.
- Johnson, D.L., 1985. Soil thickness processes. Catena Supply 6, 29–40.

- Jones, C., Lawton, J., Shachak, M., 1994. Organisms as Ecosystem Engineers. Oikos 69, 373–
 386. doi:10.2307/3545850
- Bargue, E., Marquis, F., Pillot, D., 1998. Rock-Eval 6 Applications in Hydrocarbon
 Exploration, Production, and Soil Contamination Studies. Oil Gas Sci. Technol. 53, 421–437. doi:10.2516/ogst:1998036
- Laigre, L., Reynard, E., Arnaud-Fassetta, G., Baron, L., Glenz, D., 2013. Caractérisation de la paléodynamique du Rhône en Valais central (Suisse) à l'aide de la tomographie de résistivité électrique. Géomorphologie Relief Process. Environ. 405–426. doi:10.4000/geomorphologie.10020
- Langhans, S.D., Richard, U., Rueegg, J., Uehlinger, U., Edwards, P., Doering, M., Tockner, K.,
 2012. Environmental heterogeneity affects input, storage, and transformation of coarse
 particulate organic matter in a floodplain mosaic. Aquat. Sci. 75, 335–348.
 doi:10.1007/s00027-012-0277-0
- Mardhiah, U., Caruso, T., Gurnell, A., Rillig, M.C., 2014. Just a matter of time: Fungi and roots significantly and rapidly aggregate soil over four decades along the Tagliamento River, NE Italy. Soil Biol. Biochem. 75, 133–142. doi:10.1016/j.soilbio.2014.04.012
- Mavris, C., Egli, M., Plötze, M., Blum, J.D., Mirabella, A., Giaccai, D., Haeberli, W., 2010.
 Initial stages of weathering and soil formation in the Morteratsch proglacial area (Upper Engadine, Switzerland). Geoderma 155, 359–371. doi:10.1016/j.geoderma.2009.12.019
- Moggridge, H.L., Gurnell, A.M., 2009. Controls on the sexual and asexual regeneration of
 Salicaceae along a highly dynamic, braided river system. Aquat. Sci. 71, 305–317.
 doi:10.1007/s00027-009-9193-3
- Naegeli, M.W., 1997. Sediment structure, organic matter storage, and metabolism in the
 hyporheic zone of a prealpine gravel-bed river. (PhD Thesis). ETH Zürich, Zürich.
- Osterkamp, W.R., Hupp, C.R., 2010. Fluvial processes and vegetation Glimpses of the past,
 the present, and perhaps the future. Geomorphology 116, 274–285.
 doi:10.1016/j.geomorph.2009.11.018
- Pasquale, N., Perona, P., Francis, R., Burlando, P., 2012. Effects of streamflow variability on
 the vertical root density distribution of willow cutting experiments. Ecol. Eng. 40, 167–
 172. doi:10.1016/j.ecoleng.2011.12.002
- Perona, P., Molnar, P., Crouzy, B., Perucca, E., Jiang, Z., McLelland, S., Wüthrich, D., Edmaier,
 K., Francis, R., Camporeale, C., Gurnell, A.M., 2012. Biomass selection by floods and
 related timescales: Part 1. Experimental observations. Adv. Water Resour. 39, 85–96.
 doi:10.1016/j.advwatres.2011.09.016
- Pinay, G., Fabre, A., Vervier, P., Gazelle, F., 1992. Control of C,N,P distribution in soils of
 riparian forests. Landsc. Ecol. 6, 121–132. doi:10.1007/BF00130025
- Pinay, G., Ruffinoni, C., Fabre, A., 1995. Nitrogen cycling in two riparian forest soils under different geomorphic conditions. Biogeochemistry 30, 9–29. doi:10.1007/BF02181038
- Pusch, M., Fiebig, D., Brettar, I., Eisenmann, H., Ellis, B.K., Kaplan, L.A., Lock, M.A., Naegeli,
 M.W., Traunspurger, W., 1998. The role of micro-organisms in the ecological
 connectivity of running waters. Freshw. Biol. 40, 453–495. doi:10.1046/j.13652427.1998.00372.x
- Schaetzl, R.J., Anderson, S., 2005. Soils: Genesis and Geomorphology. Cambridge University
 Press.
- 973 Sebag, D., Disnar, J.R., Guillet, B., Di Giovanni, C., Verrecchia, E.P., Durand, A., 2006.
 974 Monitoring organic matter dynamics in soil profiles by "Rock-Eval pyrolysis": bulk
 975 characterization and quantification of degradation. Eur. J. Soil Sci. 57, 344–355.
 976 doi:10.1111/j.1365-2389.2005.00745.x
- 977 SITG, 2012. Système d'Information du Territoire à Genève (SITG); Aerial pictures Allondon
 978 River.

- Steiger, J., Gurnell, A.M., 2003. Spatial hydrogeomorphological influences on sediment and nutrient deposition in riparian zones: observations from the Garonne River, France.
 Geomorphology 49, 1–23. doi:10.1016/S0169-555X(02)00144-7
- Steiger, J., Gurnell, A. m., Petts, G. e., 2001. Sediment deposition along the channel margins
 of a reach of the middle River Severn, UK. Regul. Rivers Res. Manag. 17, 443–460.
 doi:10.1002/rrr.644
- SWISSTOPO, 2012. The Swiss Federal Office of Topography (SWISSTOPO); Aerial pictures
 Allondon River.
- Tabacchi, E., Lambs, L., Guilloy, H., Planty-Tabacchi, A.-M., Muller, E., Décamps, H., 2000.
 Impacts of riparian vegetation on hydrological processes. Hydrol. Process. 14, 2959–
 2976. doi:10.1002/1099-1085(200011/12)14:16/17<2959::AID-HYP129>3.0.CO;2-B
- Tockner, K., Lorang, M.S., Stanford, J.A., 2010. River flood plains are model ecosystems to
 test general hydrogeomorphic and ecological concepts. River Res. Appl. 26, 76–86.
 doi:10.1002/rra.1328
- Uehlinger, U., 2000. Resistance and resilience of ecosystem metabolism in a flood-prone river
 system. Freshw. Biol. 45, 319–332. doi:10.1111/j.1365-2427.2000.00620.x
- Van Breemen, N., Finzi, A.C., 1998. Plant-soil interactions: ecological aspects and
 evolutionary implications. Biogeochemistry 42, 1–19.

997