Seed dispersal distances: a typology based on dispersal modes and plant traits

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Abstract

Vittoz P. and Engler R. 2007. Seed dispersal distances: a simplification for data analyses and models. Bot. Helv. 1xx: xx–xx.

The ability of plants to disperse seeds may be critical for their survival under the current constraints of landscape fragmentation and climate change. Seed dispersal distance would therefore be an important variable to include in species distribution models. Unfortunately, data on dispersal distances are scarce, and seed dispersion models only exist for some species with particular dispersal modes. To overcome this lack of knowledge, we propose a simple approach to estimate seed dispersal distances for a whole regional flora. We reviewed literature about seed dispersal in temperate regions and compiled data for dispersal distances together with information about the dispersal mode and plant traits. Based on this information, we identified seven "dispersal types" with similar dispersal distances. For each type, upper limits for the distance within which 50% and 99% of a species' seeds will disperse were estimated with the 80th percentile of the available values. These distances varied 5000-fold among the seven dispersal types, but generally less than 50-fold within the types. Thus, our dispersal types represented a large part of the variation in observed dispersal distances. The attribution of a dispersal type to a particular species only requires information that is already available in databases for most Central European species, i.e. dispersal vector (e.g. wind, animals), the precise mode of dispersal (e.g. dyszoochory, epizoochory), and species traits influencing the efficiency of dispersal (e.g. plant height, typical habitats). This typology could be extended to other regions and will make it possible to include seed dispersal in species distribution models.

Key words: Anemochory, anthropochory, autochory, hydrochory, plant migration, zoochory.

Introduction

Plant dispersal has attracted scientists since long ago (Darwin 1859; Schmidt 1918; Ridley 1930; Müller-Schneider 1983) and is particularly relevant with relation to human-driven environmental changes. For example, the survival of plant metapopulations in fragmented landscapes strongly depends on their dispersal potential (Fischer et al. 1996; Couvreur et al. 2004; Soons and Ozinga 2005), and the predicted global warming will require considerable migration rates for plant species to remain under similar climatic conditions (Malcolm et al. 2002). Nevertheless, most models attempting to predict future plant distributions did not include dispersal, considering it as unlimited (Guisan and Theurillat 2000; Thuiller et al. 2005). Even without constraints on seed dispersal, these models already predict local extinctions, e.g. for isolated populations in mountains (Guisan and Theurillat 2000; Dirnbock et al. 2003; Thuiller et al. 2005). The actual extinction rates might be even higher if plant species cannot keep pace with rapid climate change due to limited dispersal. A more precise assessment of plant species extinction risk

thus calls for the incorporation of plant dispersal potential (Pitelka et al. 1997; Davis et al. 1998; Ronce 2001).

Many studies have measured or estimated dispersal distances of plants in the field (Schneider 1935; Stöcklin and Bäumler 1996; Jongejans and Telenius 2001), and several mathematical models have been developed to estimate these distances (Tackenberg et al. 2003; Mouissie et al. 2005a; Nathan et al. 2005; Soons and Ozinga 2005). However, all of these studies have considered only a limited number of species or dispersal vectors. No dispersal distance data exist for a complete regional flora. Müller-Schneider (1986) reviewed dispersal vectors for the entire flora of Graubünden (East of Switzerland), but his work includes only few dispersal distances, most of which stem from anecdotal observations. Likewise, Bonn and Poschlod (1998) and Bonn (2004) wrote important syntheses on seed dispersal in Central Europe, but dispersal distances were only provided for a few dispersal vectors, mainly from anecdotal observations. It is thus currently impossible to conduct an assessment of the extinction risk of plant species under landscape fragmentation or global warming that would take dispersal into account.

The distance over which plants disperse seeds depends on plant traits as well as environmental conditions and varies strongly in time and space. This variability can be represented by a dispersal curve (dispersal kernel), which gives the proportion of seeds reaching a given distance (Mouissie et al. 2005a). However, it would be highly time consuming, if not impossible, to determine dispersal kernels for each species of a region. Thus, a simplified approach is needed to estimate dispersal distances for a whole regional flora. For example, if dispersal curves could be classified into a limited number of types with similar dispersal distances, and if plant species could be attributed to these "dispersal types" based on generally available plant traits, it would be possible to estimate dispersal kernels for all of them.

In this paper, we develop such an approach for the Swiss flora based on an extensive review of seed dispersal literature. We propose a typology of dispersal curves that can be applied to most Swiss and Central European plants. This typology could be extended to other regions and could be used to account for dispersal distances in species distribution models, enabling refined extinction risk assessments to be made for large numbers of species.

Methods

Plant dispersal is generally achieved through seeds. These can be enclosed in fruits or larger structures (usually called "diaspores"), but for the sake of simplification, the term "seed" will be used here as a general denomination.

Data for seed dispersal distances were compiled by reviewing a large proportion of available literature from Switzerland and other European countries, including monographies (Müller-Schneider 1983, 1986), reviews and research articles. Swiss species or close relatives were considered first priority, since our aim was to develop a typology for this region. However, other species were included when data available for Swiss species were insufficient to assess dispersal distances for certain dispersal modes (see below). The complete data set (ca. 300 values) is presented in Appendix 1. Species nomenclature follows Aeschimann et al. (1996) for the Swiss species.

The data set proved to be very heterogeneous. A small proportion of the distances had been determined through experiments, detailed field observations of seed or seedling distributions, or mathematical models. In such cases, it is often possible to calculate a dispersal kernel. However, most of the available data represent isolated and often anecdotal observations, from which a precise dispersal kernel cannot be derived. Some of these isolated observations clearly represented long-distance dispersal events (LDD), i.e. extreme values reached only by a very small minority of seeds. We therefore classified the data into three categories: (1) mean, mode or median values, (2) maximum values (99th percentiles of distribution kernels) and (3) values for LDD (clearly above the potential dispersal of 99% of the seeds). LDD values were excluded from the further analysis of the data.

Our typology of dispersal curves was based on the dispersal modes recognised by Müller-Schneider (1983). The English translation of Müller-Schneider's German terminology generally follows Bonn et al. (2000). Müller-Schneider's (1983) classification of dispersal modes is primarily based on the dispersal vector (wind, water, animals, etc.), with additional subdivisions for the differing ways in which seeds are released and transported (e.g. on the fur or after ingestion by animals). Additional subdivisions were made for dispersal modes whose efficiency clearly depends on supplementary factors: plant height, pappus efficiency and environing vegetation structure for anemochory, and vector size for zoochory. Of the numerous possible subdivisions, only those considered most relevant were retained for our classification, as explained in the next section. This yielded a total of 21 refined dispersal modes (Tab. 1).

Tab. 1. Dispersal distances for seven dispersal types, estimated as the upper limits of the distances within which 50% and 99% of the seeds of a plant population are dispersed. Note that actual dispersal distances will usually be lower than those given here (cf. Fig. 1). The dispersal distances were estimated from the 80th percentile of the data compiled in Fig. 1 as well as additional qualitative information as explained in the text ('Dispersal modes and evaluation of published dispersal distances'). The dispersal modes included in each dispersal type are indicated; they are based on dispersal vectors (categories in parentheses) and plants traits that influence the efficiency of dispersal.

Type	Dispersal distances [m]		Corresponding dispersal modes
	50%	99%	
1	0.1	1	Blastochory (autochory)
			Boleochory (anemochory) for species < 30 cm
			Ombrochory (hydrochory)
2	1	5	Ballochory (autochory)
			Cystometeorochory (anemochory)
			Chamaechory (anemochory) for fruits in grassland
			Boleochory (anemochory) for species > 30 cm
3	2	15	Pterometeorochory (anemochory) for herbs
			Myrmecochory (zoochory)
			Cystometeorochory (anemochory) ferns, Orchidaceae, Pyrolaceae, Orobanchaceae in forest
			Trichometeorochory (anemochory) in forest or little efficient plumes
			Epizoochory (zoochory) for small mammals
4	40	150	Chamaechory (anemochory) for seeds on snow or dry inflorescence
			Pterometeorochory (anemochory) for trees
			Dyszoochory (zoochory) for seeds not stocked and dispersed by small animals
5	10	500	Trichometeorochory (anemochory) in openland with efficient plumes
			Cystometeorochory (anemochory) ferns, Orchidaceae, Pyrolaceae, Orobanchaceae in openland
6	400	1500	Dyszoochory (zoochory) for seeds stocked by large animals
			Endozoochory (zoochory) for seeds eaten by birds and large vertebrates
			Epizoochory (zoochory) by large mammals
7	500	5000	Agochory (anthropochory)

Each dispersal distance in our data set was attributed to a dispersal mode, which was either the mode for which the distance had been determined (if mentioned in the original study) or the main dispersal mode of the species according to Müller-Schneider (1986). For species with more than one dispersal mode, distances that could not be clearly related to one of the modes were excluded from further data analysis. Dispersal types were then defined by grouping together dispersal modes with similar dispersal distances. This was done graphically by plotting the mean and maximal distances for each dispersal mode and identifying modes for which distances were in the same order of magnitude (Fig. 1).

Finally, we estimated upper limits of the distances, within which 50% and 99% of the seeds would disperse, by using the 80th percentiles of the available mean, mode or median values and of the maximum values. Results were rounded to one significant digit to reflect their approximate nature. Our aim was to provide a conservative estimate of the dispersal constraint experienced by most species belonging to a dispersal type. Therefore we did not take the average of the published values (Fig. 1), but rather the 80th percentile of the distribution, as this allowed us to exclude the most extreme values. In some cases, a comparison of the results with qualitative information from the literature or with the authors' experience indicated that the available data were not quite

representative for a certain dispersal type; values were then adjusted to obtain more realistic estimates. Such decisions are explained in the next section of the text for the individual dispersal modes.
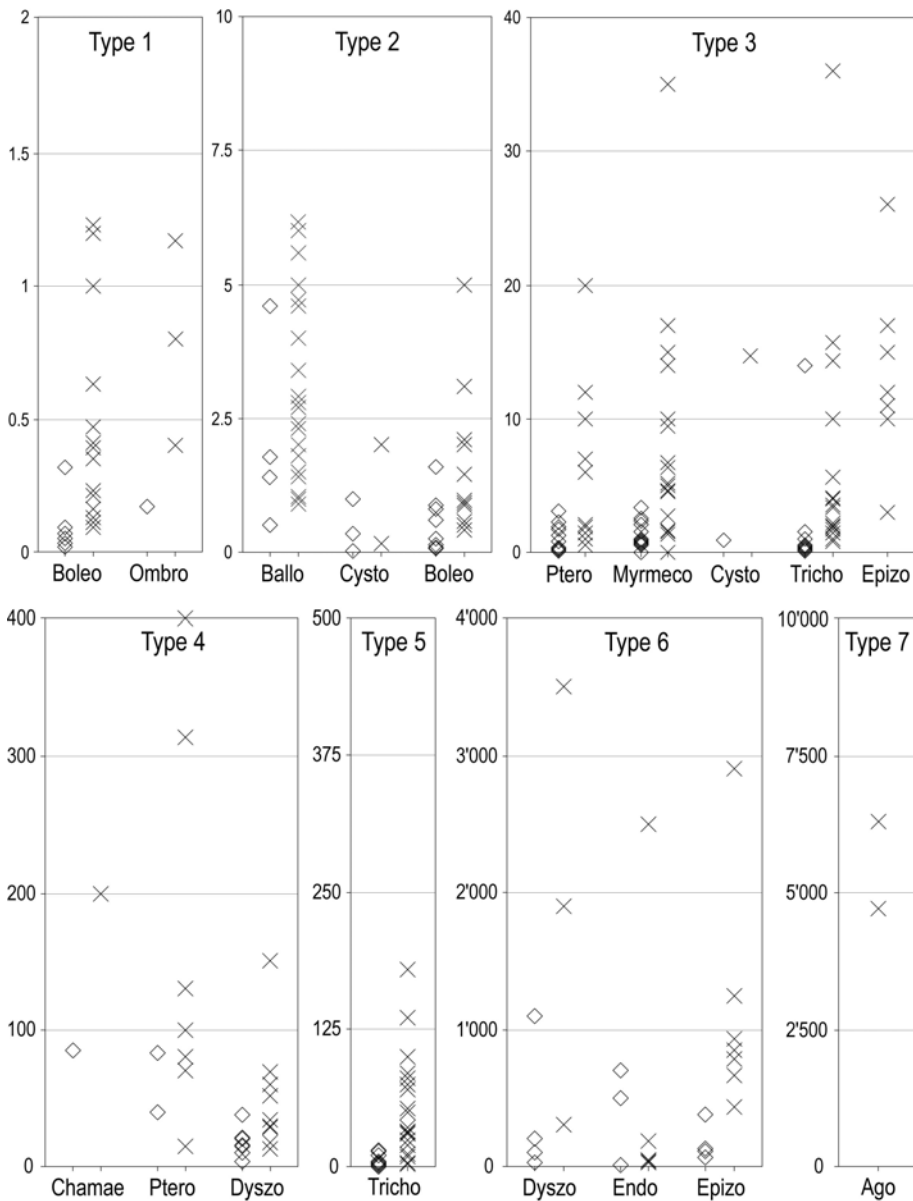

Fig. 1. Distribution of the dispersal distances found in literature (Appendix 1) for each dispersal mode, and subdivision of the data set into seven dispersal types. Diamonds are for mean, median or mode values, and crosses for 99% or maximum values (without long-distance dispersal). Four retained maximum values of type 5 are outside of the graph: 1714 m, 2112 m, 2194 m and 3673 m. See Table 1 for definitions of the dispersal modes.

Dispersal modes and evaluation of published dispersal distances

Autochory

Autochorous plants disperse seeds without the help of an external vector. As a result, dispersal is limited to very short distances.

In **blastochory**, the stem of the plant grows or crawls on the ground to deposit the seeds as far as possible from the mother plant (e.g. *Cymbalaria muralis*, *Polygonum aviculare*, *Veronica hederifolia*; Müller-Schneider 1983). No data were found in the literature, but since the dispersal

distance corresponds to the length of the stem, although species-specific, it is mostly very short and blastochory can hence be classified as type 1 (Tab. 1). This dispersal mode is, however, frequently completed by another one (Müller-Schneider 1986).

In **ballochory,** the explosion of the fruit ejects the seeds (ballistichory, ballistic dispersal). This explosion may be due to the turgescence of tissues (*Impatiens sp., Cardamine sp.*) or the tension between cells or different cell layers when the fruit is drying (*Viola sp., Vicia sp., Lotus sp.*). Published values are scattered and very variable (maximum 0.89-6.2 m; Fig. 1). Ballochory is classified in dispersal type 2 (Tab. 1).

Two further dispersal modes are **barochory** (seeds fall from the plant) and **herpochory** (seeds creep on the soil by the movement of organs in a succession of dry and wet conditions). However, since these strategies are not very efficient and always combined with other dispersal modes (Müller-Schneider 1986), they were not retained here.

Anemochory

Anemochorous seeds are dispersed by wind, often with the help of specific organs. This dispersal vector is the most studied as it is easily observable and measurable, at least over short distances (e.g. Bullock and Clarke 2000; Jongejans and Telenius 2001). Moreover, it relies on physical processes that can be translated into models (Tackenberg et al. 2003; Nathan et al. 2005; Soons and Ozinga 2005). Anemochory is subdivided according to the organs used to slow down the falling of seeds.

An air filled structure lightens small seeds in **cystometeorochory** (balloon-like). This dispersal mode is little studied. Maximum calculated distances are below 2 m (Soons and Ozinga 2005), but extreme values were measured up to 80 m for *Calluna vulgaris* (Bullock and Clarke 2000). This mode is certainly less efficient in forests, as wind is weaker, but it seemed useless to subdivide these already low values and thus cystometeorochory as a whole was attributed to type 2.

The tiny seeds of Orchidaceae, Pyrolaceae and Orobanchaceae also have a low falling velocity (0.2-0.31 m/s for Orchidaceae; Müller-Schneider 1986). But only a calculated dispersion distance is available (median 0.95 m and 99-percentile 14.7 m for *Cephalanthera damasonium*, Soons and Ozinga 2005). However, because it is thought that very light seeds (<0.05 mg), even without corresponding adaptation for anemochory, are as efficient in wind as plumed seeds (Bonn and Poschlod 1998; Greene and Calogeropoulos 2002), we decided to classify these plant families with trichometeorochory in type 5 in openland but decreased to type 3 for forest species (Tab. 1). Fern spores can be included in cystometeorochory as well, but no data exist on their dispersal capacity except a calculated distance of 330 km for *Lycopodium sp.* based on its very low falling velocity (1.8 cm/s; Schmidt 1918). This value seems exaggerated and in the absence of a more precise value, we attributed the ferns to the same types as orchids.

Plumed seeds are more efficient for wind dispersal. In **trichometeorochory**, seeds are completed with a hairy structure (e.g. pappus) to reduce falling velocity. These organs have very variable efficiency, however, with falling velocity varying from 8 cm/s for *Epilobium angustifolium* to 165 cm/s for *Pulsatilla alpina* (Müller-Schneider 1986). With an arbitrary separation at 30 cm/s, on the basis of our own observations, we distinguished species with less efficient plumes from those with efficient plumes (long plumes for small seeds). The first group has maximum distances between 1-15.7 (36) m, corresponding to type 3, and the second mainly between 20 and 179 m (Fig. 1). However, because some species have much higher calculated potentiality (e.g. up to 3600 m for *Typha latifolia*; Soons and Ozinga 2005), we retained intermediate values and assigned trichometeorochory to type 5. Forest species were classified with trichometeorochory for less efficient plumes (type 3).

In **pterometeorochory** (or pterochory), seed dispersal is improved through wings. Trees are frequent in this category, but herbs are present as well, with a generally higher falling velocity. Because tree seeds are often large and easy to find, many available maximum dispersal distances are to be classified as LDD (e.g. Müller-Schneider 1983). Reviewed maximum distances ranged mainly between 80-314 m for trees and 1-12 m for herbs (Fig. 1). Pterometeorochory was thus classified as dispersal type 3 for herbs and type 4 for trees (Tab. 1).

A much less studied dispersal mode is **chamaechory**, with diaspores rolling on the ground pushed by the wind. This diaspore can be either a circular-shaped fruit (*Colutea arborescens, Astragalus alpinus*), the fruit with calyx (*Anthyllis vulneraria*) or the complete, dry, inflorescence (synaptospermie of *Eryngium campestre, Carlina acaulis*). Chamaechory is especially common and efficient in steppes where nothing hampers dispersal (Müller-Schneider 1983), but it also occurs in mountains, with small seeds on snow (e.g. *Saxifraga bryoides, S. exarata, Sempervivum montanum*). The only available data are from Greene and Johnson (1997), who observed *Betula alleghaniensis* seeds and calculated a possible dispersion of 38 m for spherical 1mg-seeds on snow. Dispersal is usually restricted because seeds get stuck in irregularities. For chamaechory, we retained dispersal distance type 2 for fruits in grassland and type 3 for seeds on snow or carried by dry inflorescences (Tab. 1), but supplementary data would be necessary to get more precise values.

Boleochory (semachory) is another mode used by anemochorous plants. The small seeds without particular features are spread when the fruit is shaken by wind. At maturity, the stem of such plants is often rigid but elastic and sways in the wind, acting like a catapult. As animals or others may shake the capsules as well, some classify this mode independently (semachory; Bonn et al. 2000). Although small, the seeds are dense and have a high falling velocity (1.2-5 m/s; Müller-Schneider 1983; Tackenberg 2001). Consequently, Soons and Ozinga (2005) calculated very short dispersal distances, generally <0.5 m, but without considering the catapult effect. Yet, this effect is certainly important, as measured distances sometimes exceed 10 m and are always higher than calculations for the same or close species. Since the catapult effect strongly depends on the stem size, we distinguished small species (<30 cm) whose seeds rarely go beyond 1 m (type 1) from taller species (>30 cm), whose seeds may reach up to 3-5 m (type 2).

Hydrochory

Water can disperse seeds in various ways. In wetland plants, seeds are often light enough to float and move on rivers, lakes or ponds (**nautochory** of *Alisma plantago-aquatica, Carex flava, C. elata, Iris pseudacorus, Sparganium sp.*). Some seeds can float and survive for one year or more (Müller-Schneider 1983). Similarly, running water may carry many different types of seeds with heavy rains (**bythisochory**), sometimes to rivers and down to lowland areas. Bythisochory is complementary to other dispersal modes and randomly affects many different species dwelling on slopes. It is through this vector that high mountain species are frequently observed on gravel areas along rivers (Bill et al. 1999). Although the dispersal distances may be important, we did not attribute dispersal types to hydrochorous dispersal modes because distances are highly unpredictable and never documented. Moreover, nautochory is geographically limited and the bythisochory downslope restricted.

Rain may contribute to disperse seeds through the shock generated by the rain droplets hitting the fruits (**ombrochory**). Some species (e.g. *Caltha palustris, Veronica serpyllifolia, Prunella vulgaris, Thlaspi perfoliatum*) have developed fruit shapes and elastic fruitstalks in order to use this energy to eject seeds. Very few measurements are published for ombrochory, but they are all below or close to 1 m (type 1).

Zoochory

Animals are frequent and efficient vectors of dispersal, either voluntary when foraging or involuntary when carrying seeds on their fur or in their guts. Even though zoochory has often been observed and studied, estimating dispersal distances nevertheless remains difficult, as they highly depend on the disperser's behaviour. Zoochory can be split into four subcategories.

Many seeds are foraged as food by animals, which sometimes hide them as stock for the winter and forget about them, or lose them during transport (**dyszoochory** or dysochory). Vectors are mainly rodents or birds, and the dispersal distance is thus strongly dependent on the vector size. Small rodents, like voles or mice (*Clethrionomys sp., Microtus sp., Apodemus sp.*), generally disperse seeds less than 30 m (Cain et al. 1998; Xiao et al. 2004), and squirrels (*Sciurus sp.*) a little farther. In most cases, small birds disperse seeds by chance when feeding, for example when tits or woodpeckers are looking for a convenient place to break a nut. The rare available data do not

exceed 60 m. However, some larger species are more efficient dispersers by hiding fruits for winter stocks. The most famous examples are the nutcracker (*Nucifraga carcyocatactes*; Müller-Schneider 1986; Mattes 1992) and the jays (*Garrulus glandarius*; Müller-Schneider 1949; Kollmann and Schill 1996). The literature contains different data, but those are unfortunately too often extreme values (Mattes 1992), and most of the seeds are probably hidden within a few hundred meters. We thus retained type 4 when the vector of dyszoochory is a small animal and type 6 when seeds are stocked by a large animal (Tab. 1).

A particular case of dyszoochory is **myrmecochory**, or dispersal by ants. Generally interested by the elaiosome, a fatty appendix of the seeds, ants transport the seeds before eating the elaisome but leaving the rest of seed untouched and still able to germinate. Seeds may be used as building material for their nest as well, without loosing their germination potential (Müller-Schneider 1963; Cherix 1981). This dispersal mode has been extensively studied in the world, but only rarely are distances available for European plants, and they rarely exceed 10 m. Some exceptional observations nevertheless give values up to 70 m (Müller-Schneider 1983; Bonn and Poschlod 1998), and myrmecochory was hence classified as type 3.

Animals are important dispersal vectors when eating fruit or even the complete plant (Janzen 1984), and seeds go undamaged through their gut (**endozoochory**). Many authors have studied the survival of the seeds through vertebrate guts and the importance of this vector (see Janzen 1984; Pakeman 2001). As the consumer can be anything from a worm to a snail, mammal or bird, dispersal distance is very dependent on its size and mobility. No data exist for small animals and they are scarce and mostly anecdotal for larger ones such as birds or foxes. Models based on seed-retention time is a possibility for getting dispersal distance estimates, but they are still rare (e.g. Hickey et al. 1999; Vellend et al. 2003), and seed-retention time depends on seed and animal species (Bonn 2004; Mouissie et al. 2005b). Moreover, these models usually calculate linear distance, but animals generally live in a limited territory and do not move linearly. We chose type 6 to translate potential dispersal by large mammals or birds (Tab. 1).

Seeds are also frequently transported by animals in fur (**epizoochory).** This is partly the result of specific structures, with seeds or fruits bearing hooks or glandulous hairs (*Galium aparine, Arctium sp., Saxifraga tridactylites,...*) but seeds without an appendix can attach to fur as well (e.g. Fischer et al. 1996; Mouissie et al. 2005a; Römermann et al. 2005). Observations in natural conditions are rare and most of the data are from retention time measurements with sheep, cattle or dummies (e.g. Fischer et al. 1996; Mouissie et al. 2005a). Although small rodents may disperse seeds as well (Kiviniemi and Telenius 1998), the most efficient epizoochory is obtained with taller animals. The maximum distance calculated by models based on seed retention time in fur is between 435-1242 m, but can be longer with sheep whose long and curled wool is particularly efficient at retaining seeds (Fischer et al. 1996; Mouissie et al. 2005a). The habitual dispersal distance is thus estimated as type 6 for epizoochory with large animals, but much longer distances can occasionally be achieved by sheep during transhumance (Fischer et al. 1996).

Anthropochory

Seed dispersal by humans certainly always occurred, but it strongly increased during the last centuries, and became particularly important a few decades ago with the market globalisation and the intercontinental transport of goods (e.g. Hodkinson and Thompson 1997; Tinner and Schumacher 2004).

Müller-Schneider (1983, 1986) distinguished three modes of anthropochory: plants or seeds being sold for agriculture and gardening (**ethelochory**), seeds being involuntarily mixed with the previous ones (**speirochory**), or seeds travelling hidden in goods, cars, soil under soles, with hay, etc. (**agochory**). All three means can potentially lead to very long dispersal distances and are, for example, responsible for the advent of neophytes in Switzerland and Europe. But while ethelochory and speirochory mostly concern urban and cultivated areas, agochory is probably more important in natural or semi-natural ecosystems. Seed dispersal distance through anthropochory is strongly dependent on the type of human activity but, in general, agricultural activities are the most susceptible to spreading seeds in semi-natural ecosystems due to movements between fields or

meadows (McCanny and Cavers 1988). We can thus limit most of the dispersal distance to the approximate size of a farming property (type 7).

Dispersal types and estimated dispersal distances

Despite the heterogeneous origin of the data compiled here, dispersal distances for individual dispersal modes proved to be rather consistent, mostly belonging to the same order of magnitude. Across the entire data set, maximal dispersal distances ranged between 0.09 and 6300 m (LDD excluded), corresponding to a factor of 70'000 between the highest and lowest value. After classification into dispersal types, this variation was reduced to a factor of 10 for type 1, 40 for type 2, 70 for type 3, 20 for type 4, 1700 for type 5, 200 for type 6, and 1 for type 7 (very few data). This variability within types may still seem considerable, but it is small compared to the 5000-fold difference in dispersal distances between types 1 and 7. Furthermore, the high value for type 5 (trichometeorochory with efficient plumes) reflects the high variability of pappus efficiency in this category and the high variability found within species (e.g. *Taraxacum officinale*). The typology presented here thus expresses a large part of the variation in seed dispersal distances. Accordingly, attributing species to dispersal types makes it possible to describe interspecific variation in dispersal capacity.

The estimated distances in Table 1 do of course not represent the dispersal kernel of one single plant, nor even the mean pluri-annual dispersal kernel of a particular plant population. They were estimated as the upper limits (80th percentile) of the dispersal distance values (Fig. 1), meaning that they represent the dispersal potential of the plant species grouped into a dispersal type. Most plant populations will disperse over smaller distances than those indicated in Table 1, but data and models indicate that they could potentially disperse 50% or 99% of their seeds inside the retained distances. Estimating upper limits to dispersal, rather than average distances, is justified when dispersal is included as a possible constraint to species survival in predictive models of species distributions. In this case, upper dispersal limits yield a constraint that holds for all species of a certain dispersal type. This ensures that dispersal constraints will not be overestimated.

Alternative dispersal modes

Multiple dispersal vectors

About 40% of the species considered by Müller-Schneider (1986) have two or more dispersal modes. The species can either use them alternatively depending on the available vector (*Picea abies* is anemochorous or dyszoochorous with the red squirrel or some birds) or on its phenology (*Urtica dioica* is anemochorous and avoided by animals when green but grazed and endozoochorous once dry), or it can rely on them successively to improve dispersal (*Leucojum vernum* is firstly blastochorous and lately myrmecochorous; Müller-Schneider 1983).

If the most obvious dispersal mode can often be inferred from the seed or fruit morphology, finding out what the alternative dispersal modes of a species are generally requires precise observations. For example, *Campanula rotundifolia* and *Primula elatior* are considered endozoochorous by Müller-Schneider (1986) but not *Campanula scheuchzeri* or *Primula veris*, which are only described as boleochor species. This difference, probably incorrect, strongly affects their dispersal potential, as endozoochory is much more efficient than boleochory (Tab. 1), and shows the gaps in our attainments.

Recent results showed that this problem appears with other dispersal modes too. Tackenberg et al. (2003) modelled wind dispersion of seeds on the basis of their falling velocity and release height. They concluded that some species normally not considered as anemochorous could be as efficient as species traditionally thought-of as wind dispersed. Another example is given by Higgins et al. (2003), who demonstrated that a 7.8 g *Carya glabra* nut is able to disperse 647 m if uplifted by strong winds. Similarly, epizoochory concerns more species than what diaspore morphology indicates, and many plumed seeds for anemochory or smooth seeds are transported as well (Fischer et al. 1996; Couvreur et al. 2004; Mouissie et al. 2005a; Römermann et al. 2005).

When multiple vectors are recognized, it is logical to classify the species into the dispersal distance type corresponding to the most efficient one (e.g. dyszoochory for *Picea abies* or endzoochory for *Campanula rotundifolia*). But this can not consider the unsuspected supplementary vectors.

Long-distance dispersal (LDD) and Reid's paradox

The inadequacy between the dispersal potential of plants and their post-glacial recolonisation, also known as "Reid's paradox" (Clark et al. 1998), is an issue that has been recognized for a long time (Reid 1899; Skellam 1951; Cain et al. 1998). Lang (1994) calculated the migration rate of anemochorous trees through Europe and found per-generation travel distances of 0.5-5 km for *Tilia sp*, 1.2-9 km for *Abies alba*, 10-20 km for *Acer sp.* or 15-60 km for *Pinus sylvestris*. This is much higher than the 200 m considered in table 1 for 99th percentile. Similarly, dyszoochorous species with an estimated potential dispersal of 1 km (Tab. 1) showed post-glacial colonisation rate of 2.2-15 km per generation for *Quercus sp.* or 7-14 km for *Fagus sylvatica* (Lang 1994). However, as was recently found for *Fagus sylvatica* (Magri et al. 2006), it is possible that those recolonisation rates are overestimated because some glacial refugia remain yet unknown (Clark et al. 1998; Stewart and Lister 2001; Pearson 2006).

Recent data for invasive species show similar high rates of spread for many species. Pyšek and Hulme (2005) listed 16 species with colonisation superior to 1 km/y for long-distance dispersal, with a maximum of 167 km yr^{-1}. They showed that the rate of spread may be similarly high for wind, water or animal dispersed plants. But the landscape structure and human activity influence this spreading, with higher rates found in densely inhabited or particularly economically active regions (Williamson et al. 2005).

A solution to resolve this discrepancy between estimated dispersal distances and observed migration rates is to consider that dispersal vectors indicated by seed morphology mainly explain the short dispersal distances, with the rare events responsible for LDD relying on other vectors (Cain et al. 1998; Higgins et al. 2003). For example, 78% of the plants that arrived on Surtsey island (Iceland) were transported by water when only one quarter of those taxa were morphologically adapted for water dispersal (Higgins et al. 2003). Birds can transport seeds in mud sticking to their feet (Carlquist 1967), ingest some anemochorous seeds (Wilkinson 1997) or use them to build their nest (*Salix sp.* or *Clematis vitalba*; Müller-Schneider 1983; Dean et al. 1990). Seed plumes or pappus are not only very efficient for wind dispersal (anemochory), but also for fixing on animal fur (Fischer et al. 1996; Couvreur et al. 2004). Finally, humans also are efficient involuntary dispersal vectors nowadays (e.g. Hodkinson and Thompson 1997), but were also vectors during the post-glacial recolonisation, like for *Corylus avellana* or agricultural weeds (Braun-Blanquet 1970; Lang 1994; Clark et al. 1998).

Taking into account the influence of LDD on plant migration in a better manner can possibly be achieved by improving the models fitted on dispersal observations (Kot et al. 1996; Clark et al. 1998; Higgins and Richardson 1999). These improved models would help to propose dispersal distance values for the remaining 1% of the seeds (Tab. 1). But up to now, the necessary values for this improvement are missing for most species and dispersal modes, and hence we cannot propose realistic values for our dispersal distance types. Yet, even though this improvement in LDD estimation would be achieved, it could explain only a part of LDD, as the randomness of unconventional dispersal vectors cannot be standardised for all species. The importance of these accidental dispersions is not known in nature. It may be an important factor for colonising large, new areas (Higgins et al. 2003) or disturbed areas (Bergelson et al. 1993; Williamson et al. 2005), but it is certainly less frequent in closed, natural vegetation. Takahashi and Kamitani (2004) observed the colonisation of native herbaceous species in an artificial pine forest. They found that the distances dispersed by species using various dispersal vectors were similar to what we proposed for our dispersal types (Tab. 1), thus indicating that the unconventional dispersal vectors were certainly not predominant.

Conclusions

Although the data compiled in this paper are certainly incomplete, they are the most comprehensive data set currently available for the Central European flora. Our method for estimating dispersal distances based on dispersal types is less precise than the calculation of species-specific dispersal models. On the other hand, our typology can be applied to almost all European plant species, which is not the case of a species specific model. As discussed above, our typology is able to represent a large fraction of the interspecies variation in dispersal distances as long as long-distance dispersal is ignored.

Future research on dispersal mechanisms as well as the inclusion of our estimates in species distribution models will show whether the use of this typology leads to predicted migration rates that are close to the observed ones. If differences prove to be important for some of the dispersal types, our typology could be improved by adjusting the corresponding dispersal distances. Alternatively, if observed migration rates are consistently underestimated by the use of our typology, this would suggest that long-distance dispersal is much more important for long-term plant displacements than the dispersal modes presented here. Our typology is therefore certainly not the final one, but an important basis for improving predictive models of species distributions.

Résumé

La capacité des plantes à disperser est un facteur important à leur survie dans un paysage fragmenté ou sous l'influence des changements climatiques. Il est donc important de pouvoir tenir compte des distances de dispersion dans les modèles de répartition des espèces, mais les valeurs existantes, mesurées ou calculées, sont rares. Nous proposons donc une approche simple permettant d'estimer ces distances pour l'ensemble d'une flore régionale. Nous avons recherché dans la littérature les données disponibles pour la flore des régions tempérées (avant tout pour les espèces suisses) et associé les distances de dispersion trouvées avec le mode de dispersion et des traits biologiques. Sept types de dispersion ont pu être identifiés sur la base de ces informations, chaque type regroupant des espèces avec des distances de dispersion proches. Les distances à l'intérieur desquelles 50 % et 99 % des graines sont dispersées ont été estimées sur la base du 80e percentile des valeurs disponibles au sein de chaque type. Ces distances varient d'un facteur 5000 entre les sept types de dispersion, alors que les valeurs à disposition pour chaque type ne dépassent généralement pas un facteur de 50. Nos types de dispersion conservent donc une large part de la variation existante dans la dispersion des graines. L'attribution d'une espèce à un type de dispersion ne nécessite que des informations couramment disponibles, comme le vecteur de dispersion (vent, animaux, …), le mode précis de la dispersion (dyszoochorie, épizoochorie, …) et des traits biologiques influençant la dispersion (hauteur de la plante, habitat, …). Cette typologie pourrait être étendue à d'autres régions et permet d'inclure la dispersion des graines dans les modèles de répartition des espèces.

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Appendix

App. 1. Literature data on seed dispersal distances. The examples are mainly from the Swiss flora, except when data were insufficient for certain dispersal modes. Some supplementary species were thus added, mostly from temperate regions. Asterisks indicate values that were considered to represent long-distance dispersal and were therefore excluded from data analysis.

This Appendix can be downloaded freely from http://www.birkhauser.ch/BH, "Electronic supplementary material".

Species	Distance [m] *long-distance dispersal	Remarque (without indication, the values are measured)	Reference
Autochory			
Ballochory			
Cardamine amara	1.4	Maximum	Schneider 1935
Cardamine impatiens	2	Maximum	Schneider 1935
Cardamine pratensis	2.4	Maximum	Schneider 1935
Cardamine resedifolia	1.04	Maximum	Schneider 1935
Cardamine resedifolia	< 1	Maximum	Stöcklin & Bäumler 1996
Geranium maculatum	1.4 / 6.2	Mode and maximum	Stamp & Lucas 1983 and unpubl. in Willson 1993
Geranium maculatum	4.6 / 5.6	Mode and maximum	Stamp & Lucas 1983 and unpubl. in Willson 1993
Geranium molle	1.78 / 2.8	Mean and maximum	Stamp & Lucas 1983 in Cain et al. 1998
Geranium robertianum	6	Maximum	Ridley 1930 in Müller-Schneider 1983
Geranium rotundifolium	1.8	Maximum	Müller-Schneider 1933 in Müller-Schneider 1983
Geranium sylvaticum	2.7	Maximum	Müller-Schneider 1983
Impatiens parviflora	3.4	Maximum	Schneider 1935
Lathraea clandestina	4	Maximum	Guttenberg 1926 in Müller-Schneider 1983
Lathyrus vernus	1.5	Maximum	Müller-Schneider 1986
Mercurialis annua	2.9	Maximum	Müller-Schneider 1983
Mercurialis perennis	4	Maximum	Ridley 1930 in Müller-Schneider 1983
Mercurialis perennis	0.89	Maximum	Müller-Schneider 1986
Oxalis acetosella	2.3	Maximum	Moor 1940 in Müller-Schneider 1983
Oxalis acetosella	5	Maximum	Berg 2000
Viola arvensis	2.4	Maximum	Stapf 1887 in Müller-Schneider 1983
Viola canina	4.7	Maximum	Ulbrich 1928 in Müller-Schneider 1983
Viola riviniana	4.6	Maximum	Ulbrich 1928 in Müller-Schneider 1983
Viola stricta	0.5 / 3.4	Mode and maximum	Stamp & Lucas 1983 in Willson 1993
Anemochory			
Cystometeorochory			
Calluna vulgaris	1 / 80*	90% of the seeds and maximum	Bullock & Clarke 2000
Calluna vulgaris	0.35 / 2	Calculated median and 99-percentile	Soons & Ozinga 2005
Erica cinerea	1 / 80*	90% of the seeds and maximum	Bullock & Clarke 2000
Sanguisorba minor	0.03 / 0.17	Calculated median and 99-percentile	Soons & Ozinga 2005
Orchidaceae			
Cephalanthera damasonium	0.95 / 14.7	Calculated median and 99-percentile	Soons & Ozinga 2005
Trichometeorochory			
Little efficient			
Anthoxanthum odoratum	0.3 / 2	Mode and maximum	Antonovics & Ellstrand 1985 in Willson 1993
Carex frigida	< 1	Maximum	Stöcklin & Bäumler 1996
Carlina vulgaris	0.22 / 1.7	Calculated median and 99-percentile	Soons & Ozinga 2005
Carlina vulgaris	1.47	Maximum with a 16.4 km/h wind	Sheldom & Burrows 1973 in Cain et al. 1998
Crepis paludosa	0.31 / 2.2	Calculated median and 99-percentile	Soons & Ozinga 2005
Geum reptans	4	Maximum	Stöcklin & Bäumler 1996
Hieracium aurantiacum	0.1 / 1.9	Mode an maximum	Stergios 1976 in Willson 1993
Hieracium murorum aggr.	0.27 / 1.9	Calculated median and 99-percentile	Soons & Ozinga 2005
Hieracium murorum aggr.	10	Maximum	Stöcklin & Bäumler 1996
Hieracium pilosella	0.21 / 1.7	Calculated median and 99-percentile	Soons & Ozinga 2005
Hieracium sabaudum	1.5 / 15.7	Calculated median and 99-percentile	Soons & Ozinga 2005
Leontodon autumnalis	0.12 / 0.81	Calculated median and 99-percentile	Soons & Ozinga 2005
Picris hieracioides	0.40 / 3.54	Calculated median and 99-percentile	Soons & Ozinga 2005
Pulsatilla sp.	80*	Maximum	Hegi 1906-1938 in Müller-Schneider 1986
Senecio jacobea	0.49 / 4.1	Calculated median and 99-percentile	Soons & Ozinga 2005
Senecio jacobea	14 / 36	Mode and maximum	McEvoy & Cox 1987 in Willson 1993
Senecio vulgaris	0.27 / 3.4	Calculated median and 99-percentile	Soons & Ozinga 2005
Senecio vulgaris	0.34 / 2	Mean and maximum	Bergelson et al. 1993
Tragopogon pratensis	0.41 / 3.4	Calculated median and 99-percentile	Soons & Ozinga 2005
Forest plants			
Mycelis muralis	0.99 / 14.3	Calculated median and 99-percentile	Soons & Ozinga 2005
Solidago virgaurea	0.58 / 5.6	Calculated median and 99-percentile	Soons & Ozinga 2005
Highly efficient			
Adenostyles leucophylla	85	Maximum	Stöcklin & Bäumler 1996
Carduus nutans	0.83 / 9.6	Calculated median and 99-percentile	Soons & Ozinga 2005
Carduus nutans	10 / 40-100	Mode and maximum of different measures	Smith & Kok 1984 in Willson 1993
Cirsium arvense	2.0 / 53.4	Calculated median and 99-percentile	Soons & Ozinga 2005
Cirsium arvense	11.4	Maximum with a 16.4 km/h wind	Sheldom & Burrows 1973 in Cain et al. 1998
Cirsium spinosissimum	30	Maximum	Stöcklin & Bäumler 1996
Cirsium vulgare	1.8 / 31.6	Calculated median and 99-percentile	Soons & Ozinga 2005
Cirsium vulgare	1 / 32	Mean and maximum	Klinkhammer et al. 1988 in Cain et al. 1998
Clematis sp.	100	Maximum	Müller-Schneider 1986
Clematis vitalba	10.2 / 100	Calculated median and 99-percentile	Soons & Ozinga 2005
Epilobium angustifolium	7.48 / 2112	Calculated median and 99-percentile	Soons & Ozinga 2005
Epilobium ciliatum	3.65 / 179	Calculated median and 99-percentile	Soons & Ozinga 2005
Epilobium fleischeri	50	Maximum	Stöcklin & Bäumler 1996
Epilobium hirsutum	4 / 136	Calculated median and 99-percentile	Soons & Ozinga 2005
Epilobium montanum	1.6 / 49.6	Calculated median and 99-percentile	Soons & Ozinga 2005
Erigeron acer s.l.	75	Maximum	Stöcklin & Bäumler 1996
Erigeron annuus	1.6 / 35.4	Calculated median and 99-percentile	Soons & Ozinga 2005
Eriophorum angustifolium	1.1 / 24.2	Calculated median and 99-percentile	Soons & Ozinga 2005
Eupatorium cannabinum	1.6 / 23.7	Calculated median and 99-percentile	Soons & Ozinga 2005
Hieracium staticifolium	75	Maximum	Stöcklin & Bäumler 1996
Myricaria germanica	100	Maximum	Stöcklin & Bäumler 1996
Phragmites australis	13.9 / 1714	Calculated median and 99-percentile	Soons & Ozinga 2005
Poa nemoralis	50	Maximum	Stöcklin & Bäumler 1996
Salix sp.	100	Maximum	Stöcklin & Bäumler 1996
Solidago gigantea	4.2 / 136	Calculated median and 99-percentile	Soons & Ozinga 2005
Solidago virgaurea subsp. minuta	4	Maximum	Stöcklin & Bäumler 1996
Taraxacum officinale	0.22 / 2.2	Calculated median and 99-percentile	Soons & Ozinga 2005
Taraxacum officinale	50	Maximum	Stöcklin & Bäumler 1996
Tussilago farfara	20	Maximum	Stöcklin & Bäumler 1996
Tussilago farfara	10 / > 4000*	Mode and maximum	Bakker 1961 in Willson 1993
Typha angustifolia	11.3 / 2194	Calculated median and 99-percentile	Soons & Ozinga 2005
Typha latifolia	14.7 / 3673	Calculated median and 99-percentile	Soons & Ozinga 2005
Pterometeorochory			
Herbs			
Agrostis rupestris	< 1	Maximum	Stöcklin & Bäumler 1996
Angelica sylvestris	2.29	Median with a 5.3 m/s wind	Jongejans & Telenius 2001
Angelica sylvestris	0.31 / 1.91	Calculated median and 99-percentile	Soons & Ozinga 2005
Bromus inermis	1.72 / 7	Mean and maximum	Hume & Archbold 1986 in Cain et al. 1998
Bromus sterilis	20	Maximum	Howard et al. 1992 in Bullock & Clarke 2000
Heracleum sphondylium	0.38 / 2.11	Calculated median and 99-percentile	Soons & Ozinga 2005
Heracleum sphondylium	3.12	Median with a 4.7 m/s wind	Jongejans & Telenius 2001
Laserpitium latifolium	1.9	Median with a 4.3 m/s wind	Jongejans & Telenius 2001
Oxyria digyna	1	Maximum	Stöcklin & Bäumler 1996
Pastinaca sativa	3.05	Median with a 4.7 m/s wind	Jongejans & Telenius 2001
Peucedanum palustre	0.25 / 1.49	Calculated median and 99-percentile	Soons & Ozinga 2005
Peucedanum palustre	1.31	Median with a 3.4 m/s wind	Jongejans & Telenius 2001
Rumex acetosa	0.18 / 0.99	Calculated median and 99-percentile	Soons & Ozinga 2005
Rumex scutatus	12	Maximum	Stöcklin & Bäumler 1996
Scabiosa columbaria	1.9	Maximum	Verkaar et al. 1983
Selinum carvifolia	0.1 / 0.5	Calculated median and 99-percentile	Soons & Ozinga 2005
Selinum carvifolia	0.79	Median with a wind of 2.6 m/s	Jongejans & Telenius 2001
Trifolium badium	10	Maximum	Stöcklin & Bäumler 1996
Trifolium pallescens	6	Maximum	Stöcklin & Bäumler 1996
Trees			

Species	Distance [m]	Remarque (without indication, the values are measured)	Reference
Abies alba	7000*	Maximum	Bouget & Davy de Virville 1926 in Müller-Schneider 1983
Acer pseudoplatanus	5000*	Maximum	Braun-Blanquet 1913 in Müller-Schneider 1986
Acer pseudoplatanus	400-500*	Maximum	Firbas 1935 in Müller-Schneider 1986
Acer rubrum	83 / 314 / 11'371*	Calculated median, 99-percentile et maximum	Higgins et al. 2003
Alnus viridis	70	Maximum	Stöcklin & Bäumler 1996
Betula sp.	40 / 100	Limit for the majority and maximum	Greene & Calogeropoulos 2002
Carpinus betulus	130	Maximum	Müller-Schneider 1986
Fraxinus excelsior	725*	Maximum	Geiger 1960 in Müller-Schneider 1986
Fraxinus sp.	40 / 100	Limit for the majority and maximum	Greene & Calogeropoulos 2002
Larix decidua	15	Maximum	Stöcklin & Bäumler 1996
Picea abies	1500* / 800*	Maximum horizontal and vertical dispersion	Braun-Blanquet 1913 in Müller-Schneider 1986
Picea glauca	475*	Maximum	Greene & Johnson 1995 in Cain et al. 1998
Pinus sp.	40 / 100	Limit for the majority and maximum	Greene & Calogeropoulos 2002
Pinus sylvestris	2000*	Maximum	Firbas 1935 in Müller-Schneider 1983
Tilia platyphyllos	80	Maximum observed	Müller-Schneider 1986

Chamaechory
Species	Distance [m]	Remarque	Reference
Betula alleghaniensis	85 / 200	Optimum and maximum	Greene & Johnson 1997

Boleochory
Short species
Species	Distance [m]	Remarque	Reference
Achillea erba-rotta subsp. moschata	< 1	Maximum	Stöcklin & Bäumler 1996
Achillea millefolium	0.07 / 0.39	Calculated median and 99-percentile	Soons & Ozinga 2005
Achillea nana	4*	Maximum	Stöcklin & Bäumler 1996
Arabis alpina	< 1	Maximum	Stöcklin & Bäumler 1996
Arabis hirsuta	0.09 / 0.47	Calculated median and 99-percentile	Soons & Ozinga 2005
Arenaria serpyllifolia	0.03 / 0.13	Calculated median and 99-percentile	Soons & Ozinga 2005
Bellis perennis	0.02 / 0.09	Calculated median and 99-percentile	Soons & Ozinga 2005
Campanula rotundifolia	0.07 / 0.35	Calculated median and 99-percentile	Soons & Ozinga 2005
Capsella bursa-pastoris	0.05 / 0.23	Calculated median and 99-percentile	Soons & Ozinga 2005
Cerastium fontanum subsp. vulgare	0.03 / 0.16	Calculated median and 99-percentile	Soons & Ozinga 2005
Cerastium arvense	< 1	Maximum	Stöcklin & Bäumler 1996
Cerastium pedunculatum	< 1	Maximum	Stöcklin & Bäumler 1996
Eranthis hiemalis	0.32 / 1.23	Median and maximum in natural wind	Emig et al. 1999
Gentiana germanica	1.2	Maximum	Verkaar et al. 1983
Linaria alpina	12*	Maximum	Stöcklin & Bäumler 1998
Linum catharticum	0.02 / 0.13	Calculated median and 99-percentile	Soons & Ozinga 2005
Linum catharticum	0.41 / 0.63	Maximum in dense vegetation or open micro-sites	Verkaar et al. 1983
Primula veris	0.03 / 0.12	Calculated median and 99-percentile	Soons & Ozinga 2005
Sagina saginoides	10*	Maximum	Stöcklin & Bäumler 1996
Saxifraga sp.	40*	Maximum	Stöcklin & Bäumler 1996
Saxifraga tridactylites	0.02 / 0.11	Calculated median and 99-percentile	Soons & Ozinga 2005
Sempervivum sp.	< 1	Maximum	Stöcklin & Bäumler 1996
Silene rupestris	10*	Maximum	Stöcklin & Bäumler 1996

Tall species
Species	Distance [m]	Remarque	Reference
Aquilegia vulgaris	0.07 / 0.41	Calculated median and 99-percentile	Soons & Ozinga 2005
Campanula trachelium	0.25 / 1.45	Calculated median and 99-percentile	Soons & Ozinga 2005
Daucus carotta	0.15 / 0.93	Calculated median and 99-percentile	Soons & Ozinga 2005
Oenothera biennis	0.15 / 0.98	Calculated median and 99-percentile	Soons & Ozinga 2005
Oenothera biennis	1.6 / 5	Mode an maximum	Platt & Weis 1977 in Willson 1993
Papaver argemone	0.6 / 3.1	Mode an maximum	Salisbury 1942 in Willson 1993
Papaver dubium	0.9 / 2.1	Mode an maximum	Salisbury 1942 in Willson 1993
Papaver rhoeas	0.1 / 0.5	Calculated median and 99-percentile	Soons & Ozinga 2005
Papaver somniferum	2	Maximum with limited wind	Müller-Schneider 1983
Rhododendron ferrugineum	25*	Maximum	Stöcklin & Bäumler 1996
Silene pratensis	0.1 / 0.57	Calculated median and 99-percentile	Soons & Ozinga 2005
Verbascum thapsus	0.15 / 0.87	Calculated median and 99-percentile	Soons & Ozinga 2005

Hydrochory
Ombrochory
Species	Distance [m]	Remarque	Reference
Eranthis hiemalis	0.4	Maximum	Müller-Schneider 1936 in Müller-Schneider 1983
Eranthis hiemalis	0.17 / 1.17	Median and maximum dispersion under the rain	Emig et al. 1999
Thlaspi perfoliatum	0.8	Maximum	Müller-Schneider 1936 in Müller-Schneider 1983

Zoochory
Dyszoochory
Small animals
Species	Distance [m]	Remarque	Reference
Fagus sylvatica	4.13 / 13	Mean and maximum by rodents (*Clethrionomys sp.* and *Apodemus sp.*)	Jensen 1985 in Cain et al. 1998
Helianthus annuus	20	Mean by nuthatch (*Sitta europaea*)	Müller-Schneider 1949
Juglans nigra	15 / 38.1 / 151	Minimum, mean and maximum by fox squirrels (*Sciurus niger*)	Stapanian & Smith 1978, 1986
Picea abies	60	By great spotted woodpecker (*Dendrocopos major*) to open the cones	Müller-Schneider 1983
Pinus strobus	15 / 30	70 % and maximum by rodents (*Peromyscus sp.* and *Clethrionomys sp.*)	Abbott & Quink 1970
Pinus jeffreyi	21 / 69	Mean and maximim by rodents (chipmunk, *Tamias sp.*)	Vander Wall 1993 in Cain et al. 1998
Pinus sp.	1800*	Maximal dispersion by red squirrels (*Sciurus vulgaris orientis*)	Hayashida 1988
Quercus serrata	10 / 28.5	80% of the seeds and maximum by small rodents (mice and rats)	Xiao et al. 2004
Quercus macrocarpa	10 / 52	Mean and maximum by fox squirrels (*Sciurus niger*)	Stapanian & Smith 1986
Quercus petraea	18	Maximum by rodents (*Apodemus sp.*)	Kollmann & Schill 1996
Quercus sp.	15.3 / 34	Mean and maximum by rodents (*Apodemus sp.* and *Clethrionomys sp.*)	Jensen & Nielsen 1986

Large animals
Species	Distance [m]	Remarque	Reference
Corylus avellana	15000*	Maximum dispersion by nutcrackers (*Nucifraga caryocatactes*)	Mattes 1982
Fagus grandifolia	4000*	Maximum dispersion by blue jays (*Cianocitta cristata*)	Johnsonn & Adkinson 1985 in Clark et al. 1998
Fagus sylvatica	32	Dispersion by jay (*Garrulus glandiaria*)	Müller-Schneider 1949
Juglans regia	200	By carrion crow (*Corvus corone*) to break the nut	Müller-Schneider 1983
Pinus albicaulis	100 / 3500	Mean and maximum by birds	Hutchins & Lanner 1982 in Cain et al. 1998
Pinus cembra	12'000*	Maximum dispersion by nutcrackers (*Nucifraga caryocatactes*)	Sutter & Ammann 1953 in Müller-Schneider 1986
Quercus palustris	1100 / 1900	Mean and maximum dispersion by birds	Darley-Hill & Johnson 1981 in Cain et al. 1998
Quercus sp.	4000*	Maximum dispersion by jay (*Garrulus glandarius*)	Müller-Schneider 1983
Quercus petraea	300	Maximum by jay (*Garrulus glandarius*)	Kollmann & Schill 1996

Myrmecochory
Species	Distance [m]	Remarque	Reference
Allium ursinum	1.52-4.61	Different observations with *Formica rufa*	Müller-Schneider 1971
Allium ursinum	0.95	One observation with *Formica cinerea*	Müller-Schneider 1971
Asarum canadense	1.54 / 35	Mean and maximum	Cain et al. 1998
Carex pilulifera	0.75 / 1.4	Mean and maximum dispersed by *Myrmica ruginodis*	Kjellsson 1985 in Ness et al. 2004
Chelidonium majus	80*	Maximum	Senander 1906 in Bonn & Poschold 1998
Daphne striata	6.38	One observation with *Formica lugubris*	Müller-Schneider 1963
Euphorbia characias	2.1 / 4.6	Mean and maximum dispersed by *Aphaenogaster senilis*	Gomez & Espadaler 1998 in Ness et al. 2004
Euphorbia characias	2.1 / 9.4	Mean and maximum dispersed by *Messor barbarus*	Gomez & Espadaler 1998 in Ness et al. 2004
Euphorbia characias	0.79 / 1.6	Mean and maximum dispersed by *Tapinoma nigerrimum*	Gomez & Espadaler 1998 in Ness et al. 2004
Melica nutans	70*	Maximum	Senander 1906 in Bonn & Poschold 1998
Mercurialis annua	3.4 / 14	Mean and maximum dispersed by *Messor structor*	Lisci & Pacini 1997 in Ness et al. 2004
Rhamnus alaternus	1 / 5	Mean and maximum	Gomez et al. 2003
Sanguinaria canadensis	17	Maximum	Pudlo et al. 1980 in Cain et al. 1998
Sanguinaria canadensis	2.57 / 6.7	Mean and maximum dispersed by *Formica subsericea*	Ness 2004 in Ness et al. 2004
Viola hirta	70*	Maximum	Senander 1906 in Bonn & Poschold 1998
Viola sp.	0.75 / 1.5	Mean and maximum	Culver & Beattie 1978 in Cain et al. 1998
Various species	70*	Maximum dispersion by *Formica rufa*	Sernander 1906 in Müller-Schneider 1983
Various species	15	Maximum dispersion by *Lasius niger*	Sernander 1906 in Müller-Schneider 1983
Various species	0.96 / 77*	Mean and maximum in world literature	Gomez & Espadaler 1998
Various species	0.64 / 2.7	Mean and maximum in mesic deciduous forest in Japan	Higashi et al. 1989 in Ness et al. 2004
Various species	0.91 / 4.5	Mean and maximum in Oak–Pine temperate woodlands in USA	Gibson 1993 in Ness et al. 2004
Various species	0.53 / 5.2	Mean and maximum in temperate deciduous forest in USA	Kalisz et al. 1999 in Ness et al. 2004
Various species	2.4 / 10	Mean and maximum in temperate deciduous forest in USA	Kalisz et al. 1999 in Ness et al. 2004

Endozoochory
Species	Distance [m]	Remarque	Reference
Prunus avium	> 1000*	Altitudinal shift by fox	Vittoz, unpublished observation
Prunus avium	30 / 100	Mean and maximum by birds	Turcek 1968 in Bonn & Poschold 1998

Species	Distance [m]	Remarque (without indication, the values are measured)	Reference
Prunus serotina	7.1 / 35	Mean and maximum by birds	Smith 1975 in Cain et al. 1998
Rubus idaeus	> 900*	Altitudinal shift by alpine chough (*Pyrrhocorax graculus*)	Müller-Schneider 1983
Parthenocissus quinquefolia	9 / 24	Mean and maximum by birds	Hoppes 1988 in Cain et al. 1998
Phytolacca americana	33	Maximum by birds	Hoppes 1988 in Cain et al. 1998
Trillium grandiflorum	700 / 2500 / 3750*	Median, 99-percentile and maximum by deer (*Odocoileus virginianus*)	Vellend et al. 2003
Vaccinium sp.	500	Median by marten (*Martes americana*)	Hickey et al. 1999
Vitis vulpina	24	Maximum by birds	Hoppes 1988 in Cain et al. 1998
Various species	50 / 180	Maximum and extrem by blackbird (*Turdus merula*)	Müller-Schneider & Lenggenhager 1959 in Bonn & Poschold 1998

Epizoochory
Small mammals

Species	Distance [m]	Remarque	Reference
Agrimonia eupatoria	11	Maximum by wood mouse (*Apodemum flavicollis*)	Kiviniemi & Telenius 1998
Daucus carotta	17	Maximum by wood mouse (*Apodemum flavicollis*)	Kiviniemi & Telenius 1998
Geum rivale	26	Maximum by wood mouse (*Apodemum flavicollis*)	Kiviniemi & Telenius 1998
Sanicula europaea	15	Maximum by wood mouse (*Apodemum flavicollis*)	Kiviniemi & Telenius 1998
Torilis japonica	15	Maximum by wood mouse (*Apodemum flavicollis*)	Kiviniemi & Telenius 1998
Triglochin palustris	3	Maximum by wood mouse (*Apodemum flavicollis*)	Kiviniemi & Telenius 1998
Various species	12	99-percentile by wood mouse (*Apodemum flavicollis*)	Mouissie et al. 2005a

Large mammals

Species	Distance [m]	Remarque	Reference
Agrimonia eupatoria	932	Maximum by fallow deer (*Dama dama*)	Kiviniemi 1996 in Kiviniemi & Telenius 1998
Agrimonia eupatoria	780	Maximum by cattle	Kivieniemi & Eriksson in Kiviniemi & Telenius 1998
Bidens sp.	109	Mean	Bullock & Primack 1977 in Cain et al. 1998
Geum rivale	660	Maximum by cattle	Kivieniemi & Eriksson in Kiviniemi & Telenius 1998
Jurinea cyanoides	10 / 17	99-percentile and maximum	Eichberg et al. 2005
Triglochin palustris	1242	Maximum by fallow deer (*Dama dama*)	Kiviniemi & Telenius 1998
Various species	380 / 2900	Mode and 99-percentile by sheep	Mouissie et al. 2005a
Various species	65 / 435	Mode and 99-percentile by fallow deer (*Dama dama*)	Mouissie et al. 2005a
Various species	125 / 850	Mode and 99-percentile by cattle	Mouissie et al. 2005a

Anthropochory
Agochory

Species	Distance [m]	Remarque	Reference
Bromus tectorum	6300	Calculated annual migration rate	Mack 1986 in Malcolm et al. 2002
Veronica filliformis	4700	Calculated annual migration rate	Williamson et al. 2003 in Pyšek & Hulme 2005

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