

1 Synthesis

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4 **What we use is not what we know: environmental predictors in plant distribution models**

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25 **ABSTRACT**

26 **Questions:** The choice of environmental predictor variables in correlative models of plant species
27 distributions (hereafter SDMs) is crucial to ensure predictive accuracy and model realism, as
28 highlighted in multiple earlier studies. Because variable selection is directly related to a model's
29 capacity to capture important species' environmental requirements, one would expect an explicit
30 prior consideration of all ecophysiological meaningful variables. For plants, these include
31 temperature, water, soil nutrients, light, and in some cases, disturbances and biotic interactions.
32 However, the set of predictors used in published correlative plant SDM studies varies considerably.
33 No comprehensive review exists of what environmental predictors are meaningful, available (or
34 missing), and used in practice to predict plant distributions. Contributing to answer these questions
35 is the aim of this review.

36 **Methods:** We carried out an extensive, systematic review of recently published plant SDM studies
37 (years 2010-2015; $n = 200$) to determine the predictors used (and not used) in the models. We
38 additionally conducted an in-depth review of SDM studies in selected journals to identify temporal
39 trends in the use of predictors (years 2000-2015; $n = 40$).

40 **Results:** A large majority of plant SDM studies neglected several ecophysiological meaningful
41 environmental variables, and the number of relevant predictors used in models has stagnated or
42 even declined over the last 15 years.

43 **Conclusions:** Neglecting ecophysiological meaningful predictors can result in incomplete niche
44 quantification and can thus limit the predictive power of plant SDMs. Some of these missing
45 predictors are already available spatially or may soon become available (e.g., soil moisture).
46 However, others are not yet easily obtainable across whole study extents (e.g., soil pH and
47 nutrients), and their development should receive increased attention. We conclude that more effort
48 should be made to build ecologically more sound plant SDMs. This requires a more thorough
49 rationale for the choice of environmental predictors needed to meet the study goal, and the

50 development of missing ones. The latter calls for increased collaborative effort between ecological
51 and geo-environmental sciences.

52

53 **Keywords:** covariate; environment; habitat suitability; independent variable; model; niche; plant;
54 predictor; species distribution;

55

56 **Abbreviations:** DEM = digital elevation model, GIS = geographic information system, SDM =
57 correlative species distribution modelling, WoS = ISI Web of Science

58

59 **Running title:** Variable selection and species distribution models

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61

62 **INTRODUCTION**

63

64 Correlative species distribution modelling (SDM; also called ecological niche, habitat suitability,
65 and (bio)climatic envelope modelling as well as various other names, hereafter all included under
66 the acronym ‘SDM’; see Guisan et al. 2013) is a topical approach in ecology and biogeography
67 (Franklin 2009, Peterson et al. 2011, Moquet et al. 2015). Over the last decades (Booth et al. 2014),
68 the number of correlative SDM studies has steadily increased, and SDM is currently one of the most
69 popular methods used to study the impact of various threats to biodiversity and to support related
70 conservation decisions (Guisan et al. 2013). In addition to a large number of case studies on species
71 distributions for conservation and risk assessment (Broennimann & Guisan 2008; Araújo et al.
72 2011; Jiménez-Valverde et al. 2011; Alagador et al. 2014), there is on-going discussion on
73 theoretical and technical issues, including modelling techniques, selection and evaluation of models,
74 handling of spatial autocorrelation and, most importantly, variable selection (Franklin 1995; Austin

75 2002, 2007; Guisan & Thuiller 2005; Araujo & Guisan 2006; Guisan et al. 2006, Dormann 2007;
76 Elith & Leathwick 2009; Zimmermann et al. 2010; Austin & Van Niel 2011a; Thibaud et al. 2014).
77 As SDMs statistically relate environmental variables to the presence/absence (or presence-only) of a
78 species to predict species distributions (Guisan & Zimmermann 2000), the selection of the most
79 appropriate set of environmental variables as predictors is essential (Dormann 2007).

80

81 Many of the SDM (*sensu lato*) reviews published within the last 20 years have called for the use of
82 more ecologically meaningful predictors (Franklin 1995, 2009; Guisan & Zimmermann 2000;
83 Guisan & Thuiller 2005; Guisan et al. 2006, Elith & Leathwick 2009; Austin & Van Niel 2011a,
84 Peterson et al. 2011). For plants, seven environmental factors are generally considered essential for
85 growth and survival: temperature, water, nutrients, light, disturbances, biotic interactions and CO₂
86 (Körner 2014, see also Guisan & Zimmermann 2000; Austin & van Niel 2011a and Appendix S1).
87 However, although CO₂ is crucial for plant survival and productivity, it is not a limiting resource
88 under natural growth conditions at current and future atmospheric concentrations (e.g., Körner
89 2006; Norby & Zak 2011; Inauen et al. 2012; Bader et al. 2013). Under such conditions, the nutrient
90 cycle and climatic constraints control carbon capture, and therefore CO₂ is usually omitted in
91 correlative analyses of species distributions, such as SDMs, and will not be considered further in
92 this review. All of the other factors can be resources (i.e., can be consumed by the species; e.g.
93 nutrients, water, light) or regulators (i.e., can affect metabolic processes; e.g. temperature; see
94 Huston 2002) and can have direct (proximal) and indirect (distal) effects on plants (Austin 2002).
95 Thus, in standard SDMs, where species occurrence (and absence) is modelled principally as a
96 function of environmental conditions, the most realistic and accurate predictions should only be
97 achieved when all factors defining a species' niche and thus driving its distribution are accounted
98 for at the geographic scale considered (Pearson & Dawson 2003; McGill 2010). When considering
99 the environmental factors shaping species distribution from a niche modelling perspective, it is also

100 important to distinguish between bionomic (dynamically altered by the species through being
101 consumed or modified) and scenopoetic (constant, not affected by the species) variables (see
102 Hutchinson 1978; Peterson et al. 2011). In this review, by considering the environmental niche
103 (Grinnell 1917; Hutchinson 1957) of plants (Austin 1980; Austin & Smith 1989) in a wide sense,
104 we include both regulator and resource predictors, but because precise data on the dynamics of
105 environmental variables are scarce, we consider resources to remain constant (i.e. we do not
106 consider what could be consumed by the species itself) over the location and time period of the
107 study.

108

109 In addition to the importance of ecological justification for the use of ecophysiologicaly relevant
110 variables in SDMs, Austin (2002) and later Araujo & Guisan (2006) highlighted the importance of
111 acknowledging the biological significance of the selected variables, despite the diverse automated
112 and mathematically optimized variable selection methods developed for SDMs. Additionally,
113 Petitpierre et al. (in review) showed that selecting variables based on expert knowledge rather than
114 an automated selection from huge numbers of predictors can lead to better predictive performances
115 and be more reflective of biological and ecological understanding, especially for fine-scale studies
116 (see also Pearson & Dawson 2003 for the hypothesized higher importance of non-climatic variables
117 at finer scales; but see Harwood et al. 2014).

118

119 Although ecophysiological theory (Lambers et al. 2008; Körner 2014), community assembly
120 experiments (Fukami et al. 2005; Scherber et al. 2010) and biogeographical models (e.g. Franklin
121 1995; Bertrand et al. 2012; Dubuis et al. 2013; Wisz et al. 2013) stress the importance of various
122 groups of ecophysiologicaly essential predictors (Fig. 1), it seems that a large majority of SDMs
123 are built without consideration of the ecophysiological relevance and comprehensiveness of the set
124 of predictors (Pearson & Dawson 2003; Guisan & Thuiller 2005; Austin & Van Niel 2011a). The

125 most prominent explanation for this incomplete choice of predictors is the unavailability of some
126 data. It seems that largely available variables are frequently used in models (e.g., WorldClim;
127 Hijmans et al. 2005), while the use of less easily available or lacking environmental data is
128 understandably less frequent or absent in SDMs, respectively. This is however a working
129 hypothesis. Making further progress in SDM science therefore requires understanding the primary
130 causes of incomplete use of environmental information. Species distribution models are potentially
131 powerful tools to analyse and predict plant species and community distributions, but their strength,
132 validity and accuracy depend largely on the input data used. Yet, despite a long-standing knowledge
133 of which predictors should theoretically be used, no study has comprehensively reviewed which
134 ecophysiological meaningful variables are currently used and not used or missing, so that
135 recommendations can be made on where further development is required to obtain all important
136 predictors in a spatially explicit form.

137

138 Here, we evaluate whether the predictors used in correlative plant SDM studies correspond to the
139 known ecophysiological needs of plant species and whether additional constraints, such as biotic
140 factors and disturbances, are included. Simultaneously, we aim to identify which of the
141 ecophysiological relevant variables are missing and whether their omission is due to the
142 unavailability of data in a mapped format or to other causes. We do not either intend to review
143 exhaustively the literature to exemplify good from bad modelling practices, nor to provide examples
144 from our own analyses. We concentrate on niche-based species distribution models of plants
145 (vascular plants and bryophytes) and mainly consider direct abiotic variables – both regulator and
146 resource (sensu Austin 1980) – as well as biotic and disturbance variables. Plants form the basis of
147 primary production and the food chain and, as such, are important for other species, biodiversity
148 and environmental conservation in general. Focusing solely on plants also allows for a more in-
149 depth review. We acknowledge the importance of other, non-niche processes influencing plant

150 distributions, such as dispersal and (evolutionary) history (Soberón & Peterson 2005), but we do not
151 examine these processes explicitly here, as we consider them to be outside the scope of this review,
152 which centres on environmental niche predictors. Further, although efforts towards incorporating
153 the environmental predictors discussed here are also in progress in the field of mechanistic
154 modelling (see, e.g., D'Amen et al. in press), this review only considers correlative SDMs.

155 MATERIALS AND METHODS

156

157 We performed two web searches to extract original articles (excluding reviews, opinions and
158 perspectives) dealing with SDMs of vascular plants and bryophytes. The target of the first search
159 was to record recently published (2010-2015) articles in high-quality ecological journals (see
160 Appendix S2 for the journals used), while the target of the second search was to examine the
161 temporal changes in the variables used in the SDMs. The first search was performed using the query
162 ("*species distribution model**" OR "*habitat model**" OR "*ecological niche model*" OR "*niche*
163 *model**" OR "*habitat distribution model**" OR "*habitat suitability model**" OR "*niche-based*
164 *model**" OR "*bioclimatic envelope model**") AND (*vegetation* OR *plant** OR *vascular* OR
165 *bryophyte**) following Guisan et al. (2013) in the ISI Web of Science (WoS), restricting the time
166 range and journals to meet the filters specified above. This search resulted in 745 papers (hereafter
167 called the 'recent search'). The second WoS search used the same search words, but the results were
168 limited to two journals, *Journal of Vegetation Science* and *Journal of Biogeography*, after
169 preliminary queries showed the high number of plant SDM studies published in these journals,
170 accounting for the years 2000-2015. The second search was also repeated in other search engines to
171 increase the number of articles and to complement missing years, resulting in a total of 171 articles
172 (hereafter called the 'temporal search').

173

174 For all of the selected articles, we recorded the environmental predictors that were used in the
175 SDMs. To standardize the results, we divided the predictors into eight variable categories, partially
176 following Austin and Van Niel (2011a, see also Appendix S1): temperature, water, substrate
177 (including nutrients but not moisture), radiation, biotic interactions, disturbance (including
178 anthropogenic factors), topography and land use (Table 1, see detailed list of different variables in
179 Appendix S3). The temperature and water categories were further divided into mean, extreme and

180 seasonality variables, and the water category had two additional sub-classes: water balance and soil
181 moisture. The substrate-related category was divided into two classes: bedrock/pH and nutrients.
182 The category of biotic variables accounted for all variables expressing the influence of other
183 biological agents (e.g., cover of vegetation or certain plant species, species richness, and presence
184 or abundance of animal species). The disturbance category accounted for processes that primarily
185 destroy vegetation, such as fire, geomorphological disturbance and human activities, although these
186 processes can also have a positive impact on certain species (e.g., ruderals; Grime 1977).
187 Topographic and land-use related variables do not represent direct or resource variables for plants,
188 but because these are regularly included in SDMs (Franklin 1995) and have an indirect impact on
189 plant distribution through altering the distribution of temperature, moisture, nutrients and light, they
190 were also recorded here (Moeslund et al. 2013). All generally ecophysiological meaningful
191 predictor variables could be assigned to 16 classes (Table 1). Predictors that were meaningful for
192 the target of the original study but not for our review (such as fragmentation and distance to
193 waterbodies) were not recorded but are included in the total number of predictors.

194

195 From each selected SDM study, we further recorded the taxonomic group of species of interest and
196 the resolution of the input/environmental data. Only studies that used species distribution data
197 (presence-absence or presence-only) were included in further analyses, i.e. studies on species
198 richness or abundance were not considered. To avoid bias in our analyses due to the tendency to
199 highlight the use of climate variables only, we restricted our searches to studies conducted up to a
200 resolution of 1 km² (~30 arc seconds). Studies at coarser resolution (and often larger scale)
201 effectively tend to include only climatic variables due to data availability and the scale-dependence
202 of different predictors (Pearson & Dawson 2003, Thuiller et al. 2004; but see Harwood et al. 2014).
203 From the 745 ‘recent’ articles found in the WoS, 182 met our requirements (that is, they involved
204 actual SDMs concerning plants and had a maximum 1 km² resolution). Hereafter, however, our

205 analyses include 200 studies due to some articles using distinct sets of predictors for different
206 species or different spatial resolutions. Each of these studies were divided into separate studies. Of
207 the ‘temporal’ articles, forty pertained to plants and were conducted at a maximum resolution of 1
208 km². The resulted dataset was used to examine the number and type of predictors included in the
209 models. Especially, this was done in order to distinguish which predictors are frequently used in the
210 SDMs, and on the other hand, which predictors are not used and might require further developing.

211

212 To account for environmental and spatial coverage, we recorded the continent and biome of origin
213 of the data. The articles included study areas from all continents. Most studies were from Europe (*n*
214 = 84) and North America (*n* = 53), with fewer studies from Australia (*n* = 25), Africa (*n* = 20), Latin
215 America (*n* = 15) and Asia (*n* = 12). All biomes were covered with an expected bias towards
216 European and North American biomes (temperate, boreal, Mediterranean, alpine, arctic) where
217 more studies have been conducted overall.

218

219 **RESULTS**

220

221 In the ‘recent’ articles, the average number of predictors included in the models was eleven (Fig. 2).
222 The number of predictors considered in the models varied from one to 75. The different classes of
223 variables covered in the models varied from one to thirteen (out of the 16 defined in this study),
224 with only two studies covering all eight of our categories (Fig. 2). Several variables under one class
225 and/or category were often simultaneously included as predictors. Variables from the five most
226 essential categories (temperature, water, substrate, radiation, biotic interactions) were included in
227 seven studies, with all of these also including disturbance, topography and/or land-use related
228 variables. Overall, the reviewed studies represent considerable variability in the different variables

229 used. In particular, the ‘water balance’ and ‘biotic’ classes included various sets of different types
230 of factors (see Appendix S3).

231

232 Most of the ‘recent’ studies included temperature- and water-related variables (both were included
233 in 88.5 % of studies). Each of the temperature sub-classes appeared in more than half of the SDMs.

234 The most frequently included water-related variables were monthly or annual mean precipitation
235 (68.5 %), with extreme and seasonal precipitation and water balance appearing in approximately
236 one third of the studies (Fig. 3). Approximately one third of the studies included only climatic
237 variables (derived from temperature and/or precipitation). Measurements or approximations of
238 actual or potential soil water or soil moisture were incorporated in 15 studies.

239

240 Substrate-related variables were used in ~ 40 % of the studies, and variables directly representing
241 bedrock/pH or nutrients were included in approximately one quarter of the studies. Only 60 studies
242 involved variables representing light. One fifth of the studies included some biotic component as a
243 predictor variable. Variables representing natural disturbances were included in 17 studies.

244 Variables related to human activity were included in 19 studies.

245

246 After climatic variables, topographic factors were most commonly included in the SDMs screened
247 in this study (44.5 %). Land use was included in 32 studies, with one study using land use as a mask
248 to exclude certain areas.

249

250 There were no significant differences in the number of variable classes used among the continents
251 (Fig. 4). Only Latin America (LAm) had a significantly lower number of variable categories
252 compared with the other continents.

253

254 The ‘temporal search’ showed no increase in the number of categories accounted for in the SDMs
255 through time (2000-2015). On the contrary, the number of variables from different categories
256 showed a decreasing trend (Spearman’s rank correlation -0.40^* ; Fig. 5). Exceptions were the SDM
257 studies from 2011 (by Austin and Van Niel (2011b), Meier et al., Mellert et al. and Ohmann et al.),
258 which increased the number of categories included; all studies discussed the importance of selecting
259 variables on an ecological basis or the impacts of omitting meaningful predictors in the models and
260 thus included variables from multiple categories.

261

262

263 **DISCUSSION**

264

265 Ecological theory, supported by experimental and correlative studies, stresses that multiple
266 environmental factors drive the distribution of species (e.g., Larcher 1975, Fitter & Hay 2002,
267 Schulze et al. 2005, see also e.g., Guisan & Zimmermann 2000; Elith & Leathwick 2009; Franklin
268 2009; Austin & Van Niel 2011a; Bertrand et al. 2012; Dubuis et al. 2013; le Roux et al. 2013a, b),
269 particularly temperature, water, nutrients, light, biotic interactions and disturbances (see Appendix
270 S1). In recently published SDM studies, many of these factors were omitted or replaced with rough
271 surrogates (e.g., precipitation for plant available water). Indeed, more than half (53 %) of the plant
272 SDM studies reviewed here based their predictions solely on the categories of temperature and
273 water or on those two categories plus one additional variable, thus potentially neglecting several
274 other ecophysiological relevant aspects (e.g., substrate, radiation and/or biotic interactions.
275 Although it is important to highlight that not all of these categories might be meaningful for all
276 SDMs; see the next paragraph). While data availability is likely a potential reason for the omission
277 of ecophysiological meaningful predictors, the wide range of variables used in some exemplar
278 studies (see next sections and Appendix S3) indicates that some influential and available predictors

279 may tend to be neglected. Furthermore, there was no difference in the number of predictor classes
280 used in studies from the “data rich” continents (Europe, North America) and the “data poor”
281 continents (Fig. 4), suggesting that data availability may not be a sufficient explanation for the
282 absence of important predictors in the models.

283

284 The intentional use of an ecophysiologicaly incomplete set of predictors in correlative modelling is
285 acceptable, for instance, if the study deliberately focuses on the climatic niche or climatic range
286 only, provided that this is clearly acknowledged. Therefore, it is important to distinguish here
287 between two classes of studies according to their ultimate goal: studies which aim would require
288 including all potentially important variables (e.g. fine-scale predictions for conservation, or
289 addressing aspects of species’ ecology in general), and studies which aim does not necessarily
290 require more than one type of predictors (e.g. climate-change studies only interested in fitting
291 species’ climatic niches and climatic ranges). Also, in some other cases, a comprehensive set of
292 meaningful predictors may not be essential in SDMs (e.g., when illustrating the development of
293 new methodologies, or if models representing a specific aspect of the niche are explicitly desired;
294 Thuiller et al. 2005). Nevertheless, in all type of SDMs, it is important to justify the choice of
295 predictors, and interpret the results in accordance with used predictors. Indeed, only few of the
296 studies reviewed here acknowledged the ecophysiologicaly incomplete set of environmental drivers
297 used as predictors (e.g., Bertrand et al. 2012; Aguirre-Gutiérrez et al. 2013; Ikeda et al. 2014;
298 Riordan & Rundel 2014, Petitpierre et al. in review), and many studies provided no ecological
299 rationale for the choice of predictors. In the next sections, focusing our discussion on SDMs aiming
300 to comprehensively capture species ecological niche, we aim to provide such rationale, discuss
301 ways to account for the needed predictors in SDMs, and identify missing predictors for which
302 development and mapping are still needed at a fine scale. However, we do not provide any
303 estimates of an adequate number of predictors, which depends on the number and distribution of

304 species occurrences and the algorithm or approach used (see e.g., Wisz et al. (2008) and Franklin
305 (2009)).

306

307 **Temperature**

308

309 Temperature and water-related variables were the most commonly used predictors among the
310 reviewed studies (Fig. 3). While temperature is frequently accounted for in the models and plays an
311 indisputable role in regulating plant species growth and thus, distribution (see Appendix S1), two
312 noteworthy issues concerning temperature were identified from our literature analyses. First, there
313 is a large variety of temperature data products available, with the class of temperature variable used
314 having an impact on model performance (Barbet-Massin & Jetz 2014; Slavich et al. 2014). For
315 example, the impact of mean temperature on plants differs from that of extremes or seasonality in
316 both ecological meaning and modelling performance (Zimmermann et al. 2009). In seasonally
317 variable environments especially, annual mean temperature does not represent the growing season
318 or over-wintering conditions, which potentially play a more central role in governing the
319 distribution of plants (Aerts et al. 2006; Paulsen & Körner 2014). One solution to choose between
320 different temperature-related variables might be to include multiple variables in a model, as
321 exemplified by many studies using climatic data provided by WorldClim (Hijmans et al. 2005).
322 However, this raises problems of multicollinearity (Graham 2003; Dormann et al. 2013) and
323 conflicts with the objective of parsimony (Mac Nally 2000). Ultimately, the environmental
324 conditions of the study area and the requirements of the species should determine the most suitable
325 temperature-related variable(s) – a viewpoint only rarely considered or tested in the modelling
326 studies.

327

328 Second, while there is a multitude of temperature data readily available for modelling, their
329 resolution and accuracy can be coarse compared with the species data (Dingman et al. 2013;
330 Franklin et al. 2013; Potter et al. 2013; Pradervand et al. 2014). Temperature measurements are
331 typically obtained by interpolating sparse measurements and neglecting the impact of local
332 topography, land cover or water bodies on local temperatures experienced by plants (Scherrer &
333 Körner 2011; Franklin et al. 2013; Aalto et al. 2014; Slavich et al. 2014). Alternatively, improved
334 temperature maps could be obtained by a combination of increased field measurements (e.g.,
335 thermal loggers), predictive methods, high-resolution digital elevation models (DEMs) and thermal
336 remote sensing rather than spatial interpolations (Scherrer and Körner 2010, Dingman et al. 2013;
337 Pradervand et al. 2014). Thus, while the availability of temperature data is not a primary problem,
338 their usability and ecological significance in SDMs could be improved by increasing their
339 resolution and accuracy.

340

341 **Water**

342

343 Predictors representing water availability for plants are often derived from precipitation, a class of
344 climatic predictors inheriting similar challenges to those discussed for temperature. In addition,
345 precipitation is a poor surrogate for plant available water, especially in high-resolution studies that
346 cover small areas, due to the effects of local topography and soil substrate on the amount and
347 distribution of soil moisture (le Roux et al. 2013c; Piedallu et al. 2013). Therefore, while water as a
348 category of predictors is almost always acknowledged in the models, the ecophysiological
349 significance of the water predictors being used might be poor in many cases. Some studies have
350 used water balance (precipitation minus evapotranspiration), which represents a more accurate
351 measure of plant available water compared with precipitation. Some soil moisture indices derived
352 from climate data and geographic information systems (GIS) modelling are available (e.g.,

353 Trabucco & Zomer 2010), but these proxies also neglect the impact of terrain on plant available
354 moisture. Using high-resolution topographic information in combination with climate and soil
355 measurements could provide a more promising basis for modelling high-resolution soil moisture
356 data (Aalto et al. 2013; Pradervand et al. 2014).

357

358 Ideally, soil moisture measurements taken in the field should most accurately represent the water
359 available to plants. Studies that incorporate field-quantified soil moisture values in their models
360 have improved predictive power, especially at high spatial resolutions (le Roux *et al.*, 2013c).
361 However, collecting these high-resolution and accurate soil moisture data over large areas is rarely
362 feasible. Remote sensing combined with GIS provides ready-to-use (coarse-scale) indices of
363 moisture or wetness (e.g., the surface saturation degree of ASCAT soil wetness indices, see Brocca
364 et al. 2010; Lakshmi 2013; Wagner et al. 2013), and other recent developments such as Synthetic
365 Aperture Radars (Elbially et al. 2014), hyperspectral aerial images (Pottier et al. 2014) and spatial
366 modelling (Aalto et al. 2013) show promise in estimating actual soil moisture at higher resolutions.
367 To conclude, although often accounted for in SDMs with distal predictors, water-related variables
368 could be improved through combined approaches mixing refined field measures, GIS modelling and
369 remote sensing.

370

371 **Nutrients**

372

373 The role of soil and its nutrients on plant performance is acknowledged by most ecologists (Epstein
374 & Bloom 2005; see also Appendix S1) as well as their role on model performance by many
375 modellers (almost half in our study; see also Coudun et al. 2006; Coudun & Gégout 2007; Bertrand
376 et al. 2012; Dubuis et al. 2013). It seems hardly feasible to obtain high-resolution field
377 measurements of nutrient content and geo-chemical properties of soils across a whole study area.

378 Thus, most studies that included substrate variables used either geological or geomorphological
379 surrogates such as bedrock, pH or landforms, or factors related to soil structure, such as texture or
380 soil depth (Bertrand et al. 2012; Dubuis et al. 2013). This highlights the need for more sophisticated
381 indices of soil nutrient content, analogous to those being developed for soil moisture. The use of
382 soil ecological indicator values (e.g., Ellenberg) also highlights such a need (Coudun et al. 2006).
383 Improved spatial predictors of soil characteristics are thus still required, such as those derived from
384 remote sensing (Parviainen et al. 2013) or potentially from statistical modelling (Lagacherie 1992),
385 to further improve plant SDMs (Dubuis et al. 2013).

386

387 **Light**

388

389 The importance of light for plants and its use as a predictor in SDMs were previously discussed by
390 Austin and Van Niel (2011a). Solar radiation can be calculated using DEM and, if available, canopy
391 cover in efficient GIS tools (McCune & Keon 2002). However, light-related variables were only
392 included in less than one third of the studies we reviewed, meaning that more than two thirds of the
393 reviewed studies neglected an important factor controlling plant distributions, especially at local
394 scales. In the studies accounting for light, it was mostly represented by the sum of (potential) solar
395 radiation over various seasons. In these cases, the radiation variable actually expresses heat rather
396 than photosynthetically active radiation (PAR) and therefore acts similarly to temperature. To
397 obtain a real measure of PAR, light must be measured specifically, and the effects of cloud cover
398 and canopy interception must be taken into account (Aguilar et al. 2012; Wang et al. 2014).

399 Nevertheless, inclusion of a solar radiation variable often improves model prediction by adding
400 information on fine-scale energy input, especially in topographically heterogeneous areas (Austin &
401 Van Niel 2011a). At a given elevation, slopes with different aspects can have very different soil and
402 vegetation temperatures (Scherrer & Körner 2010; Gunton et al. 2015). In contrast to average

403 temperatures based mostly on adiabatic lapse rates, solar radiation can include information
404 regarding aspect, relief shading and daylight period (Kumar et al. 1997; Austin & Van Niel 2011a).
405 However, as mentioned before, the use of solar radiation as a predictor can lead to misleading
406 interpretations, as its impact on plants might strongly depend on season, canopy structure and cloud
407 cover. Thus, the radiation variables should firstly be incorporated into SDMs, seasonal variations
408 should be accounted for, and the effects of canopy and cloud cover should be included when
409 studying understory vegetation (Nieto-Lugilde et al. 2015).

410

411 **Biotic interactions**

412

413 Biotic interactions play a role in altering the potential environmental niche, for example, through
414 competition, facilitation and herbivory (Brooker & Callaghan 1998; Callaway et al. 2002; Araújo &
415 Luoto 2007; Pellissier et al. 2010; Mod et al. 2014). As the importance of biotic interactions and
416 how to measure their importance (Godsoe & Harmon 2012) and account for them in SDMs are still
417 under discussion (Kissling et al. 2012; Wisz et al. 2013), many SDMs do not include biotic factors.
418 Implicitly, these SDMs assume that the important biotic interactions (in a given area or habitat) are
419 already indirectly accounted for at the sampling stage (when gathering observations) because biotic
420 interactions influence the realized distribution of the species (McGill et al. 2006) and are thus
421 captured in the realized environmental niche (Araújo & Guisan 2006). Nonetheless, biotic
422 components were used in approximately one-fifth of the studies, indicating their increasing
423 importance in SDMs. However, explicit information on biological interactions remains difficult to
424 obtain in a spatially explicit form, as the biotic factors governing the assemblage of individual
425 species into communities are still largely unknown (Kissling et al. 2012, Wisz et al. 2013), and
426 associated assembly rules remain to be developed (Guisan & Rahbek 2011). However, surrogates
427 such as dominant species cover have been shown to provide some measure of biotic interactions (le

428 Roux et al. 2014), and incorporating these surrogates has improved both the explanatory and
429 predictive power of SDMs (Meier et al. 2010; Pellissier et al. 2010). Various methods to account for
430 biotic interactions in SDMs are presented in Kissling et al. (2012), Wisz et al. (2013) and Pollock et
431 al. (2014).

432

433 **Disturbance**

434

435 The type and necessity of including disturbance variables in models are highly environment-
436 specific. Frost-related disturbances can strongly impact vegetation in arctic and alpine areas by
437 destroying some species and subsequently, creating space for other species (le Roux et al. 2013a; le
438 Roux & Luoto 2014). In dryer areas, fire may play such a role (Tucker et al. 2012, but see
439 Crimmins et al. 2013). Disturbance has been incorporated in some models, for example, as the
440 proportion of the area that is disturbed (le Roux et al. 2013a), as an index of geomorphic
441 disturbances (Randin et al. 2009a), or as time elapsed since the last fire (Moretti et al. 2006). The
442 use of predictors related to natural disturbances in SDMs may be particularly important when
443 analysing the potential impacts of changing climate because changes in the intensity of these
444 processes associated with climatic shifts may represent key mechanisms by which changes in
445 temperature and rainfall patterns affect vegetation assemblages (le Roux & Luoto 2014, although
446 see Crimmins et al. 2013). Similar to other disturbances, the use of anthropogenic predictors is
447 situational, depending on the study environment, species and study target. For semi-natural or urban
448 landscapes and/or species highly associated with humans, the use of anthropogenic predictors might
449 be crucial to obtain reasonable predictions (Kouba et al. 2011; Senan et al. 2012).

450

451 **Topography and land use**

452

453 Variables representing topography are often included in plant distribution models (see also Franklin
454 1995). Including these variables has been demonstrated to improve plant SDMs (e.g., Sormunen et
455 al. 2011), but interpreting the actual drivers of plant distributions related to these variables can be
456 difficult. Because the effects of topographic variables on plant distributions are distal (i.e., they do
457 not directly impact plants, but they do alter light, moisture, temperature and nutrient conditions;
458 Moeslund et al. 2013), it is not possible to interpret the causal relationships between these variables
459 and the target species (Austin 2007). Correlation between indirect gradients and species distribution
460 results only from location dependence (Austin 2002). Despite the demonstrated ability of
461 topographic variables to improve local models, the use of these indirect variables hampers
462 understanding of proximal species-environment relationships and reduces transferability (Randin et
463 al. 2006). Field quantification of environmental variables or the use of purely proximal variables
464 (sensu Austin 2002) would assist in identifying the actual environmental factors that species
465 respond to and would thus provide more detailed understanding of species distributions and
466 ultimately, yield more realistic SDMs. Therefore, using in-situ measured direct and resource
467 variables instead of indirect gradients (such as elevation, aspect and topographic position) would be
468 advisable (Austin 2002; Pradervand et al. 2014), especially when SDMs are also used to explain
469 species distributions. Land use was occasionally included in the models we reviewed. Its inclusion
470 usually improves the explanatory and predictive power of SDMs (Von Holle & Motzkin 2007) but
471 only for predicting species abundances in some cases (Randin et al. 2009b). However, interpreting
472 the proximal impact of land-use predictors on plant distributions suffers the same problems
473 discussed for topographic variables (i.e., being often not proximal).

474

475 **Implications for future studies**

476

477 As hypothesized, limited data availability could be one justification for omitting potentially
478 influential ecophysiological predictors in SDMs despite their demonstrated advantages for the
479 explanatory and predictive power (e.g., Austin & Van Niel 2011b, Bertrand et al. 2012, le Roux et
480 al. 2014). The other hypothesized explanation was the intended omission, e.g., in studies of climatic
481 niches and ranges (e.g., Thuiller 2005, Petitpierre et al. 2012). However, data unavailability and
482 intended omission can hardly explain all instances (especially in data-rich areas of Europe, North-
483 America and Australia, Fig. 4) where important non-climatic factors were excluded (see similar
484 statement made 20 years previously by Franklin 1995). Indeed, many of the studies provided no
485 justification for the choice of predictors or only provided a reference to another study relying on a
486 similar set of predictors without considering the influence of the study area or the ecophysiological
487 requirements of the studied species to determine a meaningful set of predictors. Furthermore,
488 despite increasing recognition of the importance of a variety of environmental variables for
489 predicting plant distributions (e.g. Austin & Van Niel 2011a, Dubuis et al. 2013) and the increasing
490 availability of numeric data (including from remote sensing), the number of ecophysiological
491 significant variable categories considered in SDMs seems rather to have decreased during the 21st
492 century. Therefore, we argue that in the future, an ecologically sound reasoning for the choice of
493 predictors in the SDMs should become common practice, and the models and predictions should
494 always be interpreted in perspective of the set of predictors used.

495

496 In addition, our literature review highlighted that some variable classes are poorly represented in
497 terms of data quantity (e.g. global coverage) and quality (e.g. resolution). More attention should be
498 paid to ensure that all relevant environmental predictors are made available for modelling at the
499 scale investigated. Although measuring or deriving proximal predictors over large areas can be
500 difficult for single researchers, large international efforts are increasingly developed to use remote
501 sensing products for such purpose (Zimmermann et al. 2007, Estes et al. 2010). More research

502 should also be dedicated to produce finer-scale and more proximal data to improve our
503 understanding of the factors driving species distributions (Gunton et al. 2015) and therefore, the
504 production of more realistic predictions. Here too, remote sensing and GIS can produce promising
505 data products (Bradley et al. 2012, Pottier et al. 2014, He et al. 2015), and ecologists and ecological
506 modellers should give more attention to collaborative research within the geo-environmental
507 sciences.

508

509

510 **CONCLUSIONS**

511

512 Our study reveals that the rationale, selection and use of environmental predictors in many plant
513 species distribution models do not systematically match established ecophysiological theory,
514 perspectives on ecologically meaningful variable selection or demonstrated improvements in
515 SDMs, and therefore calls for the need to add several meaningful variables in SDMs. Except for the
516 pure climatic niche studies and methodological experiments, many plant SDMs so far have omitted
517 important environmental variables, and the number of predictors representing the essential
518 ecophysiological aspects pertaining to plants has not increased during the 21st century, despite
519 increased numerical data availability. In particular, nutrients, actual light, disturbance and biotic
520 interactions should be incorporated more systematically into SDMs, together with the most
521 commonly used temperature and water variables. Furthermore, the type of temperature and water
522 variables to be used should also be given more careful attention. The development of new
523 environmental variables will require improved collaborative research between ecological and geo-
524 environmental sciences as well as access to advanced technology, such as remote sensing and GIS
525 modelling approaches. Developing new sets of ecophysiological more meaningful predictors
526 provides the basis for a paradigm change in SDM research.

527

528

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535

536

537 **AUTHOR CONTRIBUTIONS**

538

539 A.G. and M.L. conceived the idea and outline for this manuscript; H.M. and D.S. performed the
540 literature review; H.M. and D.S. led the writing, with A.G. and M.L.

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546 **REFERENCES**

547

548 Aalto, J., le Roux, P.C. & Luoto, M. 2013. Vegetation mediates soil temperature and moisture in
549 arctic-alpine environments. *Arctic, Antarctic, and Alpine Research* 45: 429-439.

550 Aalto, J., le Roux, P.C. & Luoto, M. 2014. The meso-scale drivers of temperature extremes in high-
551 latitude fennoscandia. *Climate Dynamics* 42: 237-252.

- 552 Aerts, R., Cornelissen, J.H.C. & Dorrepaal, E. 2006. Plant performance in a warmer world: General
553 responses of plants from cold, northern biomes and the importance of winter and spring
554 events. *Plant Ecology* 182: 65-77.
- 555 Aguiar, L.J.G., Fischer, G.R., Ladle, R.J., Malhado, A.C.M., Justino, F.B., Aguiar, R.G. & Costa,
556 J.M.N. 2012. Modeling the photosynthetically active radiation in South West Amazonia
557 under all sky conditions. *Theoretical and Applied Climatology* 108: 631-640.
- 558 Aguirre-Gutiérrez, J., Carvalheiro, L.G., Polce, C., van Loon, E.E., Raes, N., Reemer, M. &
559 Biesmeijer, J.C. 2013. Fit-for-purpose: Species distribution model performance depends on
560 evaluation criteria – Dutch hoverflies as a case study. *PLoS One* 8: e63708.
- 561 Alagador, D., Cerdeira, J.O. & Araújo, M.B. 2014. Shifting protected areas: Scheduling spatial
562 priorities under climate change. *Journal of Applied Ecology* 51: 703-713.
- 563 Araujo, M.B. & Guisan, A. 2006. Five (or so) challenges for species distribution modelling. *Journal*
564 *of Biogeography* 33: 1677-1688.
- 565 Araújo, M.B. & Luoto, M. 2007. The importance of biotic interactions for modelling species
566 distributions under climate change. *Global Ecology and Biogeography* 16: 743-753.
- 567 Araújo, M.B., Alagador, D., Cabeza, M., Nogués-Bravo, D. & Thuiller, W. 2011. Climate change
568 threatens european conservation areas. *Ecology Letters* 14: 484-492.
- 569 Austin, M.P. 1980. Searching for a model for use in vegetation analysis. *Vegetatio* 42: 11-21.
- 570 Austin, M.P. 2002. Spatial prediction of species distribution: An interface between ecological
571 theory and statistical modelling. *Ecological Modelling* 157: 101-118.
- 572 Austin, M.P. 2007. Species distribution models and ecological theory: A critical assessment and
573 some possible new approaches. *Ecological Modelling* 200: 1-19.
- 574 Austin, M.P. & Smith, T.M. 1989. A new model for the continuum concept. *Vegetatio* 83: 35-47.
- 575 Austin, M.P. & Van Niel, K.P. 2011a. Improving species distribution models for climate change
576 studies: Variable selection and scale. *Journal of Biogeography* 38: 1-8.

- 577 Austin, M.P. & Van Niel, K.P. 2011b. Impact of landscape predictors on climate change modelling
578 of species distributions: A case study with *Eucalyptus fastigata* in southern New South
579 Wales, Australia. *Journal of Biogeography* 38: 9-19.
- 580 Bader, M.K.F., Leuzinger, S., Keel, S.G., Siegwolf, R.T.W., Hagedorn, F., Schleppei, P. & Körner,
581 C. 2013. Central European hardwood trees in a high-CO₂ future: Synthesis of an 8-year
582 forest canopy CO₂ enrichment project. *Journal of Ecology* 101: 1509-1519.
- 583 Barbet-Massin, M. & Jetz, W. 2014. A 40-year, continent-wide, multispecies assessment of relevant
584 climate predictors for species distribution modelling. *Diversity and Distributions* 20: 1285-
585 1295.
- 586 Bertrand, R., Perez, V. & Gegout, J.C. 2012. Disregarding the edaphic dimension in species
587 distribution models leads to the omission of crucial spatial information under climate
588 change: The case of *Quercus pubescens* in France. *Global Change Biology* 18: 2648-2660.
- 589 Booth, T. H., Nix, H. A., Busby, J. R. & Hutchinson, M. F. 2014. BIOCLIM: the first species
590 distribution modelling package, its early applications and relevance to most current
591 MAXENT studies. *Diversity and Distributions* 20: 1-9.
- 592 Bradley, B.A., Olsson, A.D., Wang, O., Dickson, B.G., Pelech, L., Sesnie, S.E. & Zachmann, L.J.
593 2012. Species detection vs. Habitat suitability: Are we biasing habitat suitability models
594 with remotely sensed data? *Ecological Modelling* 244: 57-64.
- 595 Brocca, L., Melone, F., Moramarco, T., Wagner, W. & Hasenauer, S. 2010. Ascat soil wetness
596 index validation through in situ and modeled soil moisture data in central Italy. *Remote
597 Sensing of Environment* 114: 2745-2755.
- 598 Broennimann, O. & Guisan, A. 2008. Predicting current and future biological invasions: Both
599 native and invaded ranges matter. *Biology Letters* 4: 585-589.
- 600 Brooker, R.W. & Callaghan, T.V. 1998. The balance between positive and negative plant
601 interactions and its relationship to environmental gradients: A model. *Oikos* 81: 196-207.

- 602 Callaway, R.M., Brooker, R.W., Choler, P., Kikvidze, Z., Lortie, C.J., Michalet, R., Paolini, L.,
603 Pugnaire, F.I., Newingham, B., Aschehoug, E.T., Armas, C., Kikodze, D. & Cook, B.J.
604 2002. Positive interactions among alpine plants increase with stress. *Nature* 417: 844-848.
- 605 Coudun, C. & Gégout, J.-C. 2007. Quantitative prediction of the distribution and abundance of
606 *Vaccinium myrtillus* with climatic and edaphic factors. *Journal of Vegetation Science*, 18:
607 517-524.
- 608 Coudun, C., Gégout, J.-C., Piedallu, C. & Rameau, J.-C. 2006. Soil nutritional factors improve
609 models of plant species distribution: An illustration with *Acer campestre* (L.) in France.
610 *Journal of Biogeography* 33: 1750-1763.
- 611 Crimmins, S.M., Dobrowski, S.Z., Mynsberge, A.R. & Safford, H.D. 2013. Can fire atlas data
612 improve species distribution model projections? *Ecological applications* 24: 1057-1069.
- 613 D'Amen, M., Rahbek, C., Zimmermann, N.E. & Guisan A. in press. Spatial predictions at the
614 community level: from current approaches to future frameworks. *Biological Reviews*
615 10.1111/brv.12222.
- 616 Dormann, C. 2007. Promising the future? Global change projections of species distributions. *Basic*
617 *and Applied Ecology* 8: 387-397.
- 618 Dormann, C.F., Elith, J., Bacher, S., Buchmann, C., Carl, G., Carré, G., Marquéz, J.R.G., Gruber,
619 B., Lafourcade, B., Leitão, P.J., Münkemüller, T., McClean, C., Osborne, P.E., Reineking,
620 B., Schröder, B., Skidmore, A.K., Zurell, D. & Lautenbach, S. 2013. Collinearity: A review
621 of methods to deal with it and a simulation study evaluating their performance. *Ecography*
622 36: 27-46.
- 623 Dubuis, A., Giovanettina, S., Pellissier, L., Pottier, J., Vittoz, P. & Guisan, A. 2013. Improving the
624 prediction of plant species distribution and community composition by adding edaphic to
625 topo-climatic variables. *Journal of Vegetation Science* 24: 593-606.

- 626 Elbially, S., Mahmoud, A., Pradhan, B. & Buchroithner, M. 2014. Application of spaceborne
627 synthetic aperture radar data for extraction of soil moisture and its use in hydrological
628 modelling at gottleuba catchment, Saxony, Germany. *Journal of Flood Risk Management* 7:
629 159-175.
- 630 Elith, J. & Leathwick, J.R. 2009. Species distribution models: Ecological explanation and prediction
631 across space and time. *Annual Review of Ecology, Evolution, and Systematics* 40: 677-697.
- 632 Epstein, E. & J., B.A. (2005) *Mineral nutrition of plants: principles and perspectives*, 2 edn.
633 Sinauer Associates, Sunderland.
- 634 Estes, L. D., P. R. Reillo, Mwangi, A. G., Okin, G. S. & Shugart, H. H. 2010. Remote sensing of
635 structural complexity indices for habitat and species distribution modeling. *Remote Sensing*
636 *of Environment* 114: 792-804.
- 637 Fitter, A.H. & Hay, R.K.M. (2002) *Environmental physiology of plants*, 3 ed. Academic Press,
638 London & San Diego. Franklin, J. 1995. Predictive vegetation mapping: Geographic
639 modelling of biospatial patterns in relation to environmental gradients. *Progress in Physical*
640 *Geography* 19: 474-499.
- 641 Franklin, J. 2009. *Mapping species distributions: Spatial inference and prediction*. Cambridge
642 University Press Cambridge.
- 643 Fukami, T., Martijn Bezemer, T., Mortimer, S.R. & van der Putten, W.H. 2005. Species divergence
644 and trait convergence in experimental plant community assembly. *Ecology Letters* 8: 1283-
645 1290.
- 646 Godsoe, W. & Harmon, L.J. 2012. How do species interactions affect species distribution models?
647 *Ecography* 35: 811–820.
- 648 Graham, M.H. 2003. Confronting multicollinearity in ecological multiple regression. *Ecology* 84:
649 2809-2815.

- 650 Grime, J.P. 1977. Evidence for the existence of three primary strategies in plants and its relevance
651 to ecological and evolutionary theory. *The American naturalist* 111: 1169-1194.
- 652 Grinnell, J. 1917. The niche-relationships of the california thrasher. *The Auk*: 427-433.
- 653 Guisan, A. & Zimmermann, N.E. 2000. Predictive habitat distribution models in ecology.
654 *Ecological Modelling* 135: 147-186.
- 655 Guisan, A. & Thuiller, W. 2005. Predicting species distribution: Offering more than simple habitat
656 models. *Ecology Letters* 8: 993-1009.
- 657 Guisan, A. & Rahbek, C. 2011. SESAM - a new framework integrating macroecological and
658 species distribution models for predicting spatio-temporal patterns of species assemblages.
659 *Journal of Biogeography* 38: 1433-1444.
- 660 Guisan, A., Lehmann, A., Ferrier, S., Austin, M., Overton, J.M.C., Aspinall, R. & Hastie, T. 2006.
661 Making better biogeographical predictions of species' distributions. *Journal of Applied*
662 *Ecology* 43: 386-392.
- 663 Guisan, A., Tingley, R., Baumgartner, J.B., Naujokaitis-Lewis, I., Sutcliffe, P.R., Tulloch, A.I.T.,
664 Regan, T.J., Brotons, L., McDonald-Madden, E., Mantyka-Pringle, C., Martin, T.G.,
665 Rhodes, J.R., Maggini, R., Setterfield, S.A., Elith, J., Schwartz, M.W., Wintle, B.A.,
666 Broennimann, O., Austin, M., Ferrier, S., Kearney, M.R., Possingham, H.P. & Buckley,
667 Y.M. 2013. Predicting species distributions for conservation decisions. *Ecology Letters* 16:
668 1424-1435.
- 669 Gunton, R.M., Polce, C. & Kunin, W.E. 2015. Predicting ground temperatures across european
670 landscapes. *Methods in Ecology and Evolution* 6: 532-542.
- 671 Harwood, T.D., Mokany, K. & Paini, D.R. 2014. Microclimate is integral to the modeling of plant
672 responses to macroclimate. *Proceedings of the National Academy of Sciences of the United*
673 *States of America*, 111: E1164-E1165.

- 674 He, K.S., Bradley, B.A., Cord, A.F., Rocchini, D., Tuanmu, M-N., Schmidlein, S., Turner, W.,
675 Wegmann, M. & Pettorelli, N. 2015. Will remote sensing shape the next generation of
676 species distribution models? *Remote Sensing in Ecology and Conservation* 1: 4-18.
- 677 Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G. & Jarvis, A. 2005. Very high resolution
678 interpolated climate surfaces for global land areas. *International journal of climatology* 25:
679 1965-1978.
- 680 Huston, M.A. (2002) Introductory essay: Critical issues for improving predictions. *Predicting*
681 *species occurrences: Issues of accuracy and scale* (eds. by J.M. Scott, P.J. Heglund and
682 M.L. Morrison), pp. 7-21.
- 683 Hutchinson, G.E. 1957. Concluding remarks. *Cold Spring Harbor Symposia on Quantitative*
684 *Biology* 22: 415-427.
- 685 Hutchinson, G.E. 1978. *An introduction to population ecology*. Yale University Press, New Haven
686 and London.
- 687 Ikeda, D.H., Grady, K.C., Shuster, S.M. & Whitham, T.G. 2014. Incorporating climate change and
688 exotic species into forecasts of riparian forest distribution. *Plos One* 9: e107037.
- 689 Inauen, N., Körner, C. & Hiltbrunner, E. 2012. No growth stimulation by CO₂ enrichment in alpine
690 glacier forefield plants. *Global Change Biology* 18: 985-999.
- 691 Jiménez-Valverde, A., Peterson, A.T., Soberón, J., Overton, J.M., Aragón, P. & Lobo, J.M. 2011.
692 Use of niche models in invasive species risk assessments. *Biological Invasions* 13: 2785-
693 2797.
- 694 Kissling, W.D., Dormann, C.F., Groeneveld, J., Hickler, T., Kühn, I., McNerny, G.J., Montoya,
695 J.M., Römermann, C., Schiffers, K., Schurr, F.M., Singer, A., Svenning, J.-C.,
696 Zimmermann, N.E. & O'Hara, R.B. 2012. Towards novel approaches to modelling biotic
697 interactions in multispecies assemblages at large spatial extents. *Journal of Biogeography*
698 39: 2163-2178.

- 699 Kouba, Y., Alados, C.L. & Bueno, C.G. 2011. Effects of abiotic and anthropogenic factors on the
700 spatial distribution of *Quercus faginea* in the spanish central pyrenees. *Plant Ecology* 212:
701 999-1007.
- 702 Kumar, L., Skidmore, A.K. & Knowles, E. 1997. Modelling topographic variation in solar radiation
703 in a gis environment. *International Journal of Geographical Information Science* 11: 475-
704 497.
- 705 Körner, C. 2006. Plant CO₂ responses: An issue of definition, time and resource supply. *New*
706 *Phytologist* 172: 393-411.
- 707 Körner, C. (2014) Pflanzen im lebensraum. *Strasburger - lehrbuch der pflanzenwissenschaften* (eds.
708 by J.W. Kadereit, C. Körner, B. Kost and U. Sonnewald), pp. 759-810. Springer Spektrum,
709 Berlin.
- 710 Lagacherie, P. 1992. *Formalisation des lois de distribution des sols pour automatiser la*
711 *cartographie pédologique à partir d'un secteur pris comme référence*. PhD Thesis,
712 Université de Montpellier, Montpellier.
- 713 Lakshmi, V. 2013. Remote sensing of soil moisture. *ISRN Soil Science* 33, 1-33.
- 714 Lambers, H., Chapin III, F.S. & Pons, T.L. 2008. *Plant water relations*. Springer, New York, US.
- 715 Larcher, W. 1975. *Physiological plant ecology*. 2nd ed. Springer, London, UK.
- 716 le Roux, P.C. & Luoto, M. 2014. Earth surface processes drive the richness, composition and
717 occurrence of plant species in an arctic–alpine environment. *Journal of Vegetation Science*
718 25: 45-54.
- 719 le Roux, P.C., Virtanen, R. & Luoto, M. 2013a. Geomorphological disturbance is necessary for
720 predicting fine-scale species distributions. *Ecography* 36: 800-808.
- 721 le Roux, P.C., Aalto, J. & Luoto, M. 2013c. Soil moisture's underestimated role in climate change
722 impact modelling in low-energy systems. *Global Change Biology* 19: 2965-2975.

- 723 le Roux, P.C., Pellissier, L., Wisz, M.S. & Luoto, M. 2014. Incorporating dominant species as
724 proxies for biotic interactions strengthens plant community models. *Journal of Ecology* 102:
725 767-775.
- 726 le Roux, P.C., Lenoir, J., Pellissier, L., Wisz, M.S. & Luoto, M. 2013b. Horizontal, but not vertical,
727 biotic interactions affect fine-scale plant distribution patterns in a low energy system.
728 *Ecology* 94: 671-682.
- 729 Mac Nally, R. 2000. Regression and model-building in conservation biology, biogeography and
730 ecology: The distinction between – and reconciliation of – ‘predictive’ and ‘explanatory’
731 models. *Biodiversity & Conservation* 9: 655-671.
- 732 McCune, B. & Keon, D. 2002. Equations for potential annual direct incident radiation and heat
733 load. *Journal of Vegetation Science* 13: 603-606.
- 734 McGill, B.J. 2010. Matters of scale. *Science* 328: 575-576.
- 735 McGill, B.J., Enquist, B.J., Weiher, E. & Westoby, M. 2006. Rebuilding community ecology from
736 functional traits. *Trends in Ecology & Evolution* 21: 178-185.
- 737 Meier, E.S., Edwards Jr, T.C., Kienast, F., Dobbertin, M. & Zimmermann, N.E. 2011. Co-
738 occurrence patterns of trees along macro-climatic gradients and their potential influence on
739 the present and future distribution of *fagus sylvatica* l. *Journal of Biogeography* 38: 371-
740 382.
- 741 Meier, E.S., Kienast, F., Pearman, P.B., Svenning, J.-C., Thuiller, W., Araújo, M.B., Guisan, A. &
742 Zimmermann, N.E. 2010. Biotic and abiotic variables show little redundancy in explaining
743 tree species distributions. *Ecography* 33: 1038-1048.
- 744 Mellert, K.H., Fensterer, V., Kuechenhoff, H., Reger, B., Koelling, C., Klemmt, H.J. & Ewald, J.
745 2011. Hypothesis-driven species distribution models for tree species in the Bavarian Alps.
746 *Journal of Vegetation Science* 22: 635-646.

- 747 Mod, H.K., le Roux, P.C. & Luoto, M. 2014. Outcomes of biotic interactions are dependent on
748 multiple environmental variables. *Journal of Vegetation Science* 25: 1024-1032.
- 749 Moeslund, J.E., Arge, L., Bøcher, P.K., Dalgaard, T. & Svenning, J.-C. 2013. Topography as a
750 driver of local terrestrial vascular plant diversity patterns. *Nordic Journal of Botany* 31: 129-
751 144.
- 752 Moretti, M., Conedera, M., Moresi, R. & Guisan, A. 2006. Modelling the influence of change in fire
753 regime on the local distribution of a mediterranean pyrophytic plant species (*Cistus*
754 *salviifolius*) at its northern range limit. *Journal of Biogeography* 33: 1492-1502.
- 755 Mouquet, N., Lagadeuc, Y., Devictor, V., Doyen, L., Duputié, A., Eveillard, D., Faure, D., Garnier,
756 E., Gimenez, O., Huneman, P., Jabot, F., Jarne, P., Joly, D., Julliard, R., Kéfi, S., Kergoat,
757 G.J., Lavorel, S., Le Gall, L., Meslin, L., Morand, S., Morin, X., Morlon, H., Pinay, G.,
758 Pradel, R., Schurr, F.M., Thuiller, W. & Loreau, M. 2015. Predictive ecology in a changing
759 world. *Journal of Applied Ecology* 52: 1293-1310.
- 760 Nieto-Lugilde D., Lenoir J., Abdulkhak S., Aeschimann D., Dullinger S., Gégout J.-C., Guisan A.,
761 Pauli H., Renaud J., Theurillat J.-P. Thuiller, W., Van Es, J., Vittoz, P., Willner, W.,
762 Wohlgemuth, T., Zimmermann, N.E. & Svenning, J.-C. 2015. Tree cover at fine and coarse
763 spatial grains interacts with shade tolerance to shape plant species distributions across the
764 Alps. *Ecography* 37: 1-12.
- 765 Norby, R.J. & Zak, D.R. 2011. Ecological lessons from free-air CO₂ enrichment (FACE) experiments.
766 *Annual Review of Ecology, Evolution, and Systematics* 42: 181.
- 767 Ohmann, J.L., Gregory, M.J., Henderson, E.B. & Roberts, H.M. 2011. Mapping gradients of
768 community composition with nearest-neighbour imputation: extending plot data for
769 landscape analysis. *Journal of Vegetation Science* 22: 660-676.

- 770 Parviainen, M., Zimmermann, N., Heikkinen, R. & Luoto, M. 2013. Using unclassified continuous
771 remote sensing data to improve distribution models of red-listed plant species. *Biodiversity
772 and Conservation* 22: 1731-1754.
- 773 Paulsen, J. & Körner, C. 2014. A climate-based model to predict potential treeline position around
774 the globe. *Alpine botany* 124: 1-12.
- 775 Pearson, R.G. & Dawson, T.P. 2003. Predicting the impacts of climate change on the distribution of
776 species: Are bioclimate envelope models useful? *Global Ecology and Biogeography* 12:
777 361-371.
- 778 Pellissier, L., Bråthen, K.A., Pottier, J., Randin, C.F., Vittoz, P., Dubuis, A., Yoccoz, N.G., Alm, T.,
779 Zimmermann, N.E. & Guisan, A. 2010. Species distribution models reveal apparent
780 competitive and facilitative effects of a dominant species on the distribution of tundra plants.
781 *Ecography* 33: 1004-1014.
- 782 Peterson, A.T., Soberón, J., Pearson, R.G., Anderson, R.P., Martínez-Meyer, E., Nakamura, M. &
783 Araújo, M.B. 2011. *Ecological niches and geographic distributions*. Princeton University
784 Press, Princeton, USA.
- 785 Petitpierre, B., C. Kueffer, Broennimann, O., Randin, C., Daehler, C. & Guisan, A. 2012. Climatic
786 niche shifts are rare among terrestrial plant invaders. *Science* 335: 1344-1348.
- 787 Petitpierre, B., Broenniman, O., Kueffer, C., Daehler, C. and Guisan A. in review. Selecting the best
788 set of predictors to build transferable species distribution models. *Global Ecology and
789 Biogeography*.
- 790 Piedallu, C., Gegout, J.-C., Perez, V. & Lebourgeois, F. 2013. Soil water balance performs better
791 than climatic water variables in tree species distribution modelling. *Global Ecology and
792 Biogeography* 22: 470-482.
- 793 Pollock, L.J., Tingley, R., Morris, W.K., Golding, N., O'Hara, R.B., Parris, K.M., Veski, P.A. &
794 McCarthy, M.A. 2014. Understanding co-occurrence by modelling species simultaneously

- 795 with a Joint Species Distribution Model (JSDM). *Methods in Ecology and Evolution* 5: 397-
796 406.
- 797 Potter, K.A., Woods, H.A. & Pincebourde, S. 2013. Microclimatic challenges in global change
798 biology. *Global Change Biology* 19: 2932-2939.
- 799 Pottier, J., Malenovský, Z., Psomas, A., Homolová, L., Schaepman, M.E., Choler, P., Thuiller, W.,
800 Guisan, A. & Zimmermann, N.E. 2014. Modelling plant species distribution in alpine
801 grasslands using airborne imaging spectroscopy. *Biology Letters* 10
- 802 Pradervand, J.-N., Dubuis, A., Pellissier, L., Guisan, A. & Randin, C. 2014. Very high resolution
803 environmental predictors in species distribution models: Moving beyond topography?
804 *Progress in Physical Geography* 38: 79-96.
- 805 Randin, C.F., Vuissoz, G., Liston, G.E., Vittoz, P. & Guisan, A. 2009a. Introduction of snow and
806 geomorphic disturbance variables into predictive models of alpine plant distribution in the
807 western swiss alps. *Arctic Antarctic and Alpine Research* 41: 347-361.
- 808 Randin, C.F., Jaccard, H., Vittoz, P., Yoccoz, N.G. & Guisan, A. 2009b. Land use improves spatial
809 predictions of mountain plant abundance but not presence-absence. *Journal of Vegetation*
810 *Science* 20: 996-1008.
- 811 Randin, C.F., Dirnböck, T., Dullinger, S., Zimmermann, N.E., Zappa, M. & Guisan, A. 2006. Are
812 niche-based species distribution models transferable in space? *Journal of Biogeography* 33:
813 1689-1703.
- 814 Riordan, E.C. & Rundel, P.W. 2014. Land use Compounds habitat losses under projected climate
815 change in a threatened california ecosystem. *Plos One* 9: e86487.
- 816 Scherber, C., Eisenhauer, N., Weisser, W.W., Schmid, B., Voigt, W., Fischer, M., Schulze, E.-D.,
817 Roscher, C., Weigelt, A., Allan, E., Beszler, H., Bonkowski, M., Buchmann, N., Buscot, F.,
818 Clement, L.W., Ebeling, A., Engels, C., Halle, S., Kertscher, I., Klein, A.-M., Koller, R.,
819 König, S., Kowalski, E., Kummer, V., Kuu, A., Lange, M., Lauterbach, D., Middelhoff, C.,

- 820 Migunova, V.D., Milcu, A., Muller, R., Partsch, S., Petermann, J.S., Renker, C., Rottstock,
821 T., Sabais, A., Scheu, S., Schumacher, J., Temperton, V.M. & Tschardtke, T. 2010. Bottom-
822 up effects of plant diversity on multitrophic interactions in a biodiversity experiment. *Nature*
823 468: 553-556.
- 824 Scherrer, D. & Körner, C. 2010. Infra-red thermometry of alpine landscapes challenges climatic
825 warming projections. *Global Change Biology* 16: 2602-2613.
- 826 Scherrer, D. & Körner, C. (2011). Topographically controlled thermal-habitat differentiation buffers
827 alpine plant diversity against climate warming. *Journal of Biogeography* 38: 406-416.
- 828 Schulze, E., Beck, E. & Muller-Hohenstein, K. 2005. *Plant Ecology*, 1st ed. Springer-Verlag,
829 Berlin. Senan, A.S., Tomasetto, F., Farcomeni, A., Somashekar, R.K. & Attorre, F. 2012.
830 Determinants of plant species invasions in an arid island: Evidence from Socotra island
831 (Yemen). *Plant Ecology* 213: 1381-1392.
- 832 Slavich, E., Warton, D.I., Ashcroft, M.B., Gollan, J.R. & Ramp, D. 2014. Topoclimate versus
833 macroclimate: how does climate mapping methodology affect species distribution models
834 and climate change projections? *Diversity and Distributions*, 20: 952-963.
- 835 Soberón, J. & Peterson, A.T. 2005. Interpretation of models of fundamental ecological niches and
836 species' distributional areas. *Biodiversity Informatics* 2: 1-10.
- 837 Sormunen, H., Virtanen, R. & Luoto, M. 2011. Inclusion of local environmental conditions alters
838 high-latitude vegetation change predictions based on bioclimatic models. *Polar Biology* 34:
839 883-897.
- 840 Thibaud, E., Petitpierre, B., Broennimann, O., Davison, A.C. & Guisan, A. 2014. Measuring the
841 relative effect of factors affecting species distribution model predictions. *Methods in*
842 *Ecology and Evolution* 5: 947-955.
- 843 Thuiller, W., Araújo, M.B. & Lavorel, S. 2004. Do we need land-cover data to model species
844 distributions in Europe? *Journal of Biogeography* 31: 353-361.

- 845 Thuiller, W., Richardson, D.M., Pysek, P., Midgley, G.F., Hughes, G.O. & Rouget, M. 2005.
846 Niche-based modelling as a tool for predicting the risk of alien plant invasions at a global
847 scale. *Global Change Biology* 11: 2234-2250.
- 848 Trabucco, A. & Zomer, R.J. 2010. *Global Soil Water Balance Geospatial Database*. CGIAR
849 Consortium for Spatial Information. Published online, available from the CGIAR-CSI
850 GeoPortal at: <http://www.cgiar-csi.org>
- 851 Tucker, C.M., Rebelo, A.G. & Manne, L.L. 2012. Contribution of disturbance to distribution and
852 abundance in a fire-adapted system. *Ecography* 35: 348-355.
- 853 Wagner, W., Hahn, S., Kidd, R., Melzer, T., Bartalis, Z., Hasenauer, S., Figa-Saldaña, J., de
854 Rosnay, P., Jann, A., Schneider, S., Komma, J., Kubu, G., Brugger, K., Aubrecht, C., Züger,
855 J., Gangkofner, U., Kienberger, S., Brocca, L., Wang, Y., Blöschl, G., Eitzinger, J.,
856 Steinnocher, K., Zeil, P. & Rubel, F. 2013. The ascat soil moisture product: A review of its
857 specifications, validation results, and emerging applications. *Meteorologische Zeitschrift* 22:
858 5-33.
- 859 Wang, L., Gong, W., Ma, Y., Hu, B. & Zhang, M. 2014. Photosynthetically active radiation and its
860 relationship with global solar radiation in Central China. *International Journal of*
861 *Biometeorology* 58: 1265-77.
- 862 Wisz, M.S., Hijmans, R. Li, J., Peterson, A.T., Graham, C.H. & Guisan, A. 2008. Effects of sample
863 size on the performance of species distribution models. *Diversity and Distributions* 14: 763-
864 773.
- 865 Wisz, M.S., Pottier, J., Kissling, W.D., Pellissier, L., Lenoir, J., Damgaard, C.F., Dormann, C.F.,
866 Forchhammer, M.C., Grytnes, J.A., Guisan, A., Heikkinen, R.K., Høye, T.T., Kuhn, I.,
867 Luoto, M., Maiorano, L., Nilsson, M.C., Normand, S., Ockinger, E., Schmidt, N.M.,
868 Termansen, M., Timmermann, A., Wardle, D.A., Aastrup, P. & Svenning, J.C. 2013. The

869 role of biotic interactions in shaping distributions and realised assemblages of species:
870 Implications for species distribution modelling. *Biological Reviews* 88: 15-30.

871 Von Holle, B. & Motzkin, G. 2007. Historical land use and environmental determinants of
872 nonnative plant distribution in coastal southern new england. *Biological Conservation* 136:
873 33-43.

874 Zimmermann, N. E., Edwards, T. C. Moisen, G. G., Frescino, T. S. & Blackard, J. A. 2007. Remote
875 sensing-based predictors improve distribution models of rare, early successional and
876 broadleaf tree species in Utah. *The Journal of Applied Ecology* 44: 1057-
877 1067. Zimmermann, N.E., Edwards, T.C., Graham, C.H., Pearman, P.B. & Svenning, J.C.
878 2010. New trends in species distribution modelling. *Ecography* 33: 985-989.

879 Zimmermann, N.E., Yoccoz, N.G., Edwards, T.C., Meier, E.S., Thuiller, W., Guisan, A., Schmatz,
880 D.R. & Pearman, P.B. 2009. Climatic extremes improve predictions of spatial patterns of
881 tree species. *Proceedings of the National Academy of Sciences of the United States of*
882 *America* 106 Suppl 2: 19723-19728.

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886 **Supporting Information**

887 Online Supporting Information may be found in the online version of this article:

888 **Appendix S1** Ecophysiological meaning of different categories of variables for plant species

889 **Appendix S2** Journals and numbers of studies included in the paper.

890 **Appendix S3** Variables included in the different classes and categories

891 **TABLES**

892

893 Table 1. Classification of predictors into eight categories and 16 classes (see Appendix 3 for details
 894 of the variables). The five first columns represent the most important categories, which we refer to
 895 as ‘the five most essential categories’ in the text.

| Cate- gories | Temperature | Water | Substrate | Radiation | Biotic inter- actions | Disturbance | Topo- graphy | Land use |
|-------------------------|--|---|------------------|-------------------|--------------------------------------|-----------------------------------|---------------------------|---------------------|
| | mean (annual, seasonal, monthly) temperature | mean / summed (annual, seasonal, monthly) precipitation | pH, bedrock | radiation, clouds | variables related to other organisms | geomorpho-logical processes, fire | slope, aspect, elevation, | land-use classes |
| Classes | extreme temperatures | extreme precipitation | nutrients | | | anthropo- genic variables | | |
| | seasonality | seasonality | | | | | | |
| | | water balance | | | | | | |
| | | soil moisture | | | | | | |

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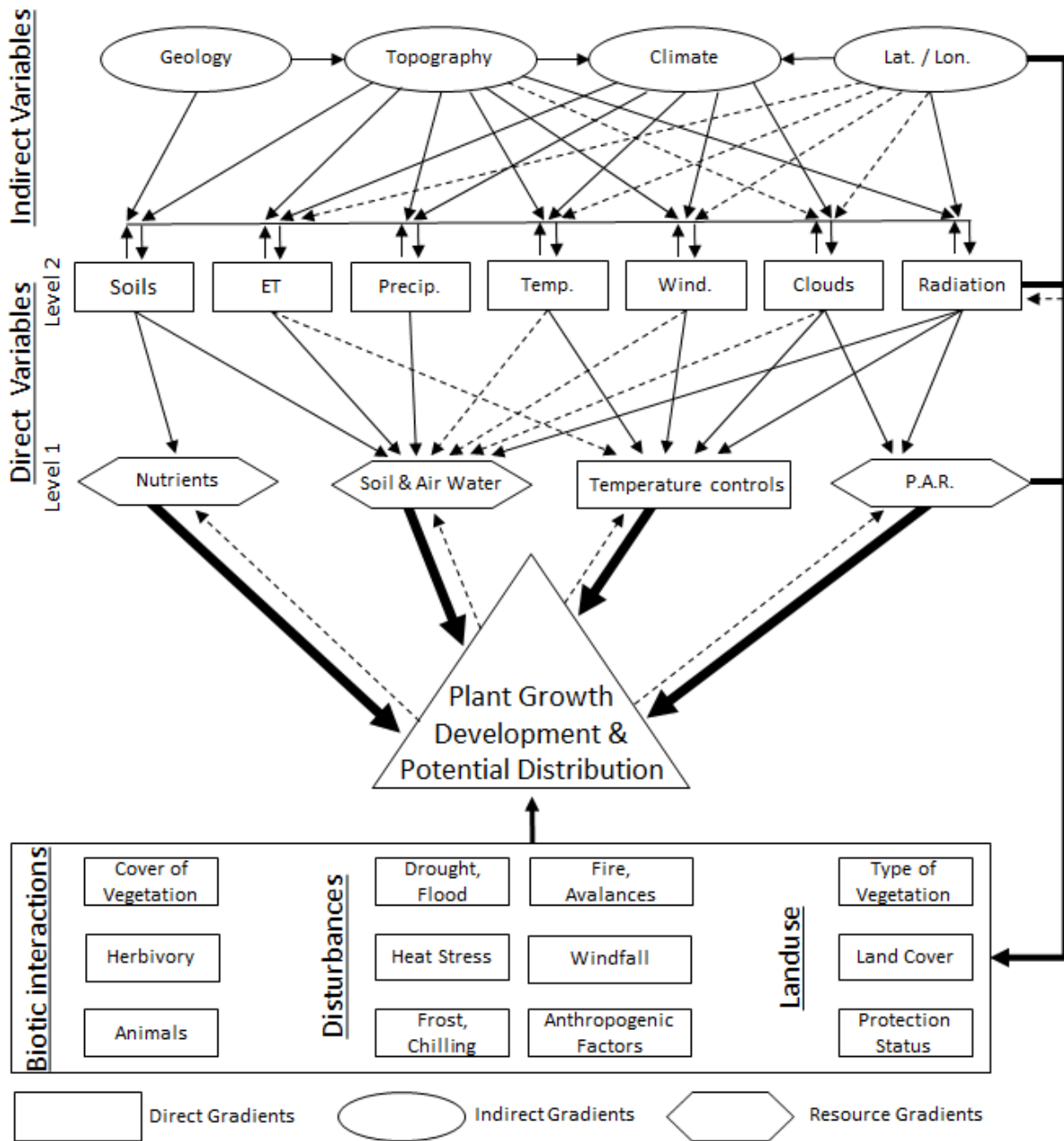
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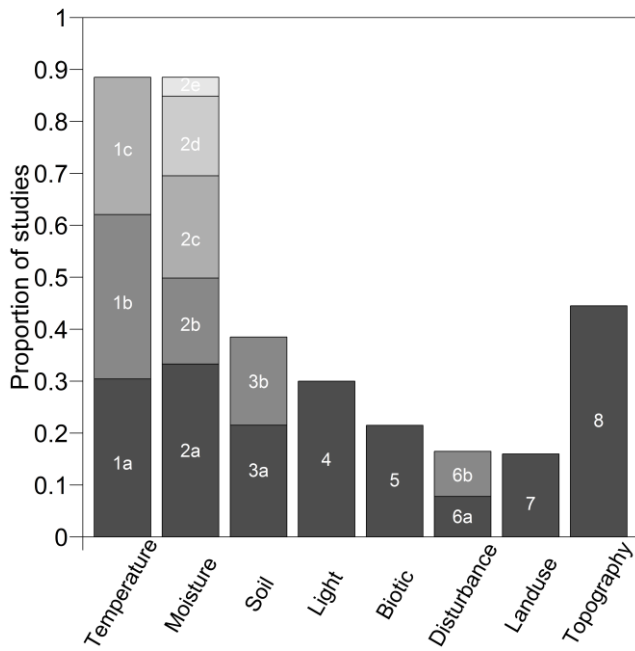
903 FIGURES



904

905 Fig. 1. Example of a conceptual framework of relationships between resources, direct and indirect
 906 environmental gradients and their influence on the growth, performance, and geographical
 907 distribution of vascular plants and vegetation. ET = Evapotranspiration, P.A.R = Photosynthetically
 908 active radiation. Adapted from Guisan & Zimmermann 2000.

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910

911 Fig. 3. Proportion of studies in which each predictor class was used: 1a mean temperature; 1b

912 extreme temperature; 1c seasonality of temperature; 2a mean precipitation; 2b extreme

913 precipitation; 2c seasonality of precipitation; 2d water balance; 2e soil moisture; 3a pH/bedrock; 3b

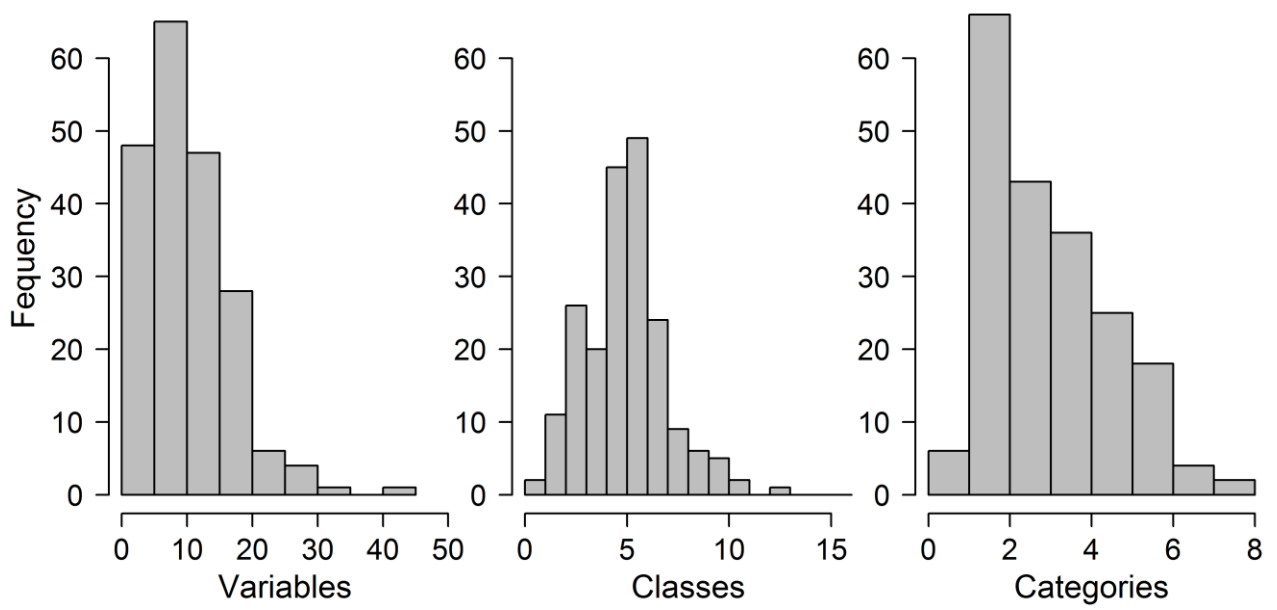
914 nutrients; 4 radiation; 5 biotic interactions; 6a natural disturbances; 6b human disturbances; 7 land

915 use; 8 topography.

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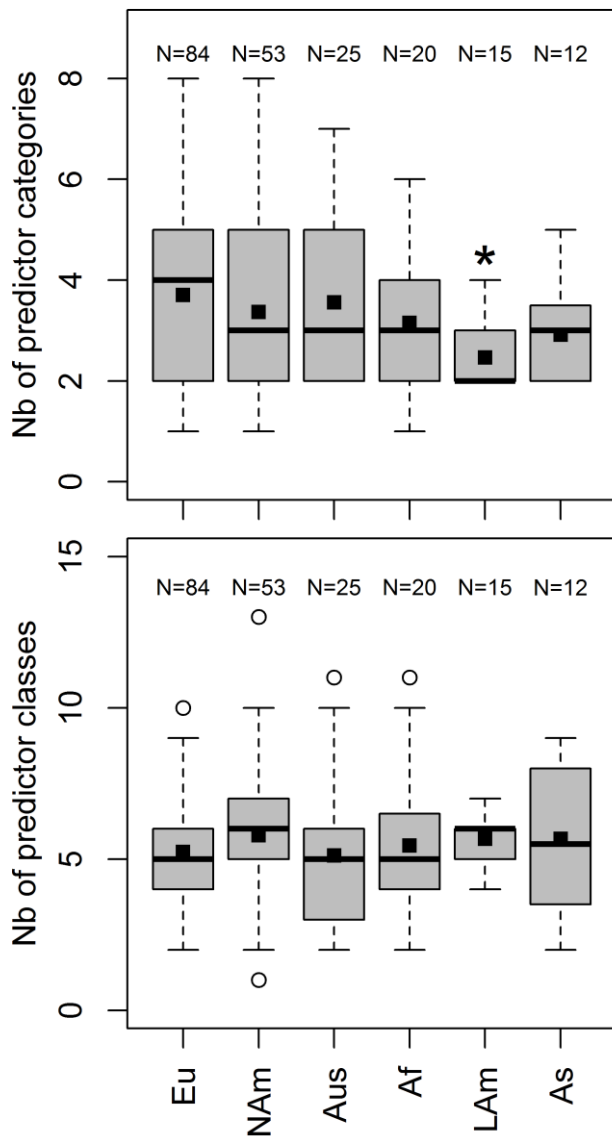
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920 Fig. 2. Frequency of the number of variables, classes (16) and categories (see Table 1) accounted
 921 for in the plant species distribution modelling studies. One outlier value (75) was removed from the
 922 histogram representing the number of variables in the SDMs.

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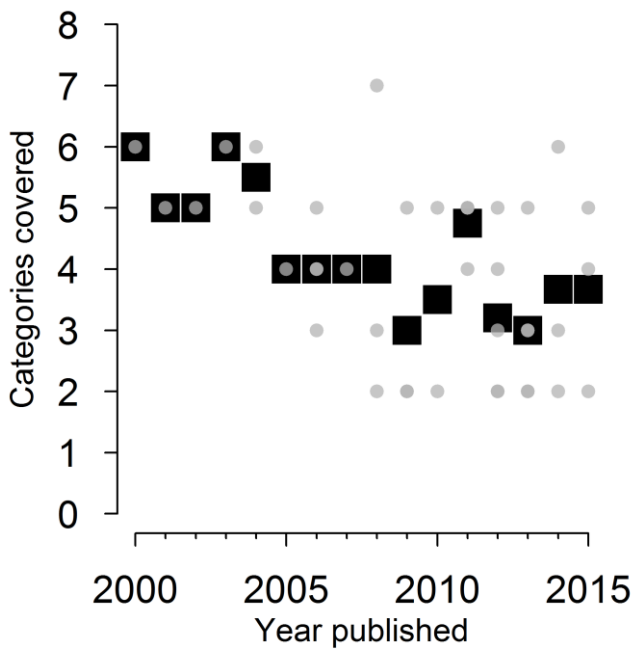
925 Fig. 4. The number of categories and classes accounted for in the plant species distribution models

926 (SDMs) using data from different continents. The boxes represent the median and the 25/75

927 percentile, and the whiskers are 2 SD. The mean is indicated by a black square, and significant

928 differences are marked with an asterisk.

929



930

931 Fig. 5. Number of variable categories (as presented in Table 1) used in the SDM studies published
 932 in two journals from 2000-2015. Spearman's rank correlation between the years and categories
 933 included is -0.40^* . Black squares indicate the mean values of all studies published within a year,
 934 and the grey dots indicate individual studies.

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Supporting information to the paper

Heidi K. Mod, Daniel Scherrer, Miska Luoto & Antoine Guisan. What we use is not what we know: environmental predictors in plant distribution models. *Journal of Vegetation Science*.

Appendix S1. Ecophysiologicaly relevant variables for plant distribution

Seven environmental factors are generally considered as essential for plant growth and survival: light, water, temperature, nutrients, biotic interactions, disturbance and CO₂ (Guisan & Zimmermann, 2000, Kadereit *et al.*, 2014). All these factors can have direct and indirect effects on plants and in combination with dispersal and historical factors, they define the abundance and distribution of plant species (Soberon & Peterson, 2005).

Temperature is the most common regulatory factor considered in SDM's. Temperature directly effects the speed of growth and in case of strong seasonality defines the growing season length. Additionally, minimum and maximum temperatures can reflect physiological thresholds for plants by frost or heat resistance.

Water has several essential functions in plants including photosynthesis, cooling by transpiration and maintaining turgor. In SDMs "water" is usually reflected by either precipitation alone or in combination with evapotranspiration (e.g. water balance). These environmental variables are considered a proxy for plant available water. However, this might not be the case if soils and topography are heterogeneous, as plant available water is strongly influence by both soil type and topographic position. The seasonality of available water/precipitation might lead to temporal flooding, drought or snow cover and thus requires special adaptations by the present plant species.

Nutrients are taken up with water by roots (often with the help of mycorrhiza). Many micronutrients are essential for plant survival including potassium, calcium, magnesium, sulphur, boron, chlorine, manganese, molybdenum and zinc but most significant for productivity are usually the contents of nitrogen and phosphorus. Nutrients in a wider sense can also influence the pH of the soils, whereas bedrock together with living organism are the primary regulators of available nutrients in soils. Therefore, while deriving nutrient content of the soils might not be effective, bedrock, soil pH and soil texture are often used as surrogates in the SDMs.

Light is often expressed as global radiation and therefore energy (W/m²) driving temperature (air, leaf, and soil) and evapotranspiration. However, for plants light reflects also photo active radiation (PAR) and is thereby directly related to photosynthesis. While radiation can be easily modelled and is relatively independent of the vegetation, PAR is strongly affected by the canopy structure of the vegetation. Therefore, the available light for photosynthesis might be very different in a forest compared to open grassland at otherwise similar global radiation (energy). Additionally, light might contain important signals for plant development (e.g. germination and photoperiodism).

Biotic interactions act among and between species, and have both positive and negative impact by prohibiting or ameliorating growth. Impact of other species can be direct (e.g. competition, herbivory) or indirect (e.g. ameliorating harsh microclimatic conditions, shading, nutrient addition by manure). Biotic interactions have been included to the SDMs as e.g. presence or cover of dominant species, remote sensed vegetation index or interaction matrices for multispecies co-occurrence datasets.

Disturbance's impact is mainly negative for species as soil, water, air or snow movement, fire or anthropogenic activities destroy vegetation. However, some ruderal species benefit from disturbances indirectly as they decrease competition and create space by destroying dominant species, and some specialist species require disturbances, as fire and water-logging for germination. Disturbances have also secondary impact on vegetation, by indirectly impacting soil properties: e.g. cryoturbation bring nutrients closer to soil surface.

CO₂ the carbon source for plants and therefore essential for their survival and productivity. However, the levels of CO₂ among sites don't vary enough to be limiting or having a significant influence on species composition and therefore are ignored in correlative models such as SDM's.

Topography and land use do not have a direct impact on plants, but they affect the distribution of ecophysiosologically meaningful factors (e.g. temperature, light). Topography and land use related variables are easily available and incorporating them often improve SDMs.

References:

Guisan, A. & Zimmermann, N.E. 2000. Predictive habitat distribution models in ecology. *Ecological Modelling* 135: 147-186.

Kadereit, J.W., Körner C., Kost B., Sonnewald U. 2014. *Strasburger - Lehrbuch der Pflanzenwissenschaften*, Berlin, Springer Spektrum.

Soberón, J. & Peterson, A.T. 2005. Interpretation of models of fundamental ecological niches and species' distributional areas.

Biodiversity Informatics 2: 1-10.

Appendix S2. Journals and number of studies included by the query (and subsequent analyses).

Recent search:

Ambio 3 (1)

American Naturalist 4 (1)

Annals of Botany 9 (2)

Applied Vegetation Science 9 (5)

Biodiversity and Conservation 22 (5)

Biological Conservation 49 (10)

Biology Letters 2 (2)

Climatic Change 7 (3)

Conservation Biology 9 (1)

Conservation Letters 1 (0)

Diversity and Distribution 62 (19)

Ecography 52 (21)

Ecological Applications 20 (5)

Ecological Modelling 50 (20)

Ecological Monographs 3 (3)

Ecology 9 (2)

Ecology Letters 11 (1)

Ecosystems 1 (0)

Functional Ecology 3 (0)

Global Change Biology 58 (16)

Global Ecology and Biogeography 45 (11)

Journal of Applied Ecology 13 (1)

Journal of Biogeography 59 (17)

Journal of Ecology 14 (2)

Journal of Vegetation Science 24 (8)

Landscape Ecology 12 (1)

Methods in Ecology and Evolution 15 (1)

Nature Communications 1 (0)

New Phytologist 5 (2)

Oecologia 1 (0)

Oikos 5 (1)

Perspectives in Plant Ecology 7 (0)

Plant Ecology 8 (7)

Plos One 113 (29)

Proceedings of National Academy of Sciences 10 (1)

Proceedings of Royal Society B 14 (2)

Science 2 (0)

Trends in Ecology and Evolution 1 (0)

Temporal search

Journal of Vegetation Science 39 (12)

Journal of Biogeography 122 (28)

Appendix S3. Variables included in different classes and categories.

TEMPERATURE

mean temperature

- (annual / monthly) mean temperature (of coldest / warmest / driest / wettest quarter / summer / winter)
- soil temperature
- warmth index (the annual sum of positive differences between monthly mean temperatures and e.g. 5 degrees, i.e. a measure of the effective warmth for plants)

extreme temperature

- (annual) min / max temperature (of coldest / warmest driest / wettest quarter / month / season)
- mean temperature of coldest / warmest / driest / wettest month
- mean daily max / min temperature (for DJF / MAM / JJA / SON)

temperature seasonality

- seasonality, annual / diurnal range
- growing degree days (all thresholds) / freezing degree days (FDD) (soil / air) / non-FDD / chilling degree days
- isothermality
- heat units (annual sum of daily temperatures exceeding X degrees)
- frost duration
- winter / summer cold / heat wave duration

WATER

mean precipitation

- (annual / monthly) mean / summed precipitation (of coldest / warmest / driest / wettest quarter / season)
- days with rain > 1 mm
- rainfall intensity

extreme precipitation

- mean / summed / min / max precipitation of coldest / warmest / driest / wettest month
- highest 5-day precipitation

precipitation seasonality

- seasonality, annual range
- snow (cover duration, annual snowfall)

- dry / wet season / day length / intensity / frequency
- % of annual precipitation falling during the growing season
- average flood duration
- the standard deviation of hydrographs

water-balance

- (annual / seasonal / monthly) water balance
- (annual / seasonal) evapo-transpiration, vapour pressure
- (mean / annual / seasonal / soil) water / moisture deficit / surplus / availability / stress
- (annual / seasonal / plant available) water / wetness / moisture / aridity index
- water content
- flow accumulation
- average water level
- soil moisture (days; days when soil moisture - air temperature ratio is favourable for plant growth)
- waterlogging index

soil water capacity

- soil water capacity, measured soil moisture
- soil drainage class
- hydraulic soil presence class

SUBSTRATE

bedrock / ph

- bedrock, lithology, rock type
- pH
- surface geology, geological substrate

nutrients

- nutrients, fertility, Cation-exchange capacity, calcareous
- soil material / depth / order / quality / texture / type
- organic matter, loaminess, alluvial, clay / silt / sand content, salt, gypsum
- soil grain size, bulk density
- FAO soil group
- remote sensed Normalized difference soil index, soil production index

- water regime (ordered classes from dry to waterlogged)

LIGHT

- solar radiation (daily, annual, seasonal)
- most / least radiated quarter
- mean hours of sunshine
- clouds

BIOTIC

- NDVI, Landsat bands, Enhanced Vegetation Indices, remote sensed vegetation (indices / classes)
- vegetation height / density / volume/ cover
- canopy / forest / tree cover
- productivity, Net Primary Production
- ecological classification, succession time
- pollinators
- litter
- distance to moorland, moorland presence / absence
- stand basal area
- % of sparsely / dense vegetated brownfield
- % of brownfield with low / high vegetation

DISTURBANCE

natural

- fire, volcanic ash
- geomorphological disturbance
- trampling, grazing
- % area of disturbed terrain

anthropogenic

- population / settlement / building density
- distance to urban areas / roads / harbour / roads
- agriculture, afforestation, soil drainage, roads, human perturbation, forest / etc. management

- human footprint, anthropization degree
- brick rubble
- ownership status (measure of land management)
- predominance of exotic species

TOPOGRAPHY

- altitude (range), terrain curvature, topographic position, slope, flatness, meso-topography, % of steep topography, slope type
- aspect, eastness, northness
- rockiness, ruggedness, topographic wetness index,
- topographic diversity

LAND-USE

- Corine, land-use classes (if only "biotic" land-use -> 'biotic' class)
- distance to potential forest, age of forest