1	Bridgehead effects and role of adaptive evolution in invasive populations
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7	Abstract
8	Biological invasions are a major threat to biodiversity, agriculture and human health.
9	Invasive populations can be the source of additional new introductions, leading to a self-
10	accelerating process whereby "invasion begets invasion". This phenomenon, coined
11	"bridgehead effect", has been proposed to stem from the evolution of higher
12	invasiveness in a primary introduced population. There is, however, no conclusive
13	evidence that the success of bridgehead populations stems from the evolution of
14	increased invasiveness. Instead, we argue that a high frequency of secondary
15	introductions can be explained by increased abundance in the bridgehead region or the
16	topology of human transport networks. We outline the type of evidence and
17	experiments that are needed to demonstrate adaptive evolution and higher invasion
18	success of introduced populations.
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20	Key words:
21	Biological invasions, evolution of invasiveness, adaptation, globalization
22	

23 i. Evolution as a potential driver of bridgehead effects

24 All species have limits to their geographical distributions [1]. Historically these limits 25 were determined by a range of environmental factors such as climate or resource 26 availability, biotic interactions and physical barriers to dispersal [1]. However in the 27 current period, particularly since the Industrial Revolution, increased international 28 trade and human movement have resulted in the accidental movement of many species 29 worldwide. This on-going movement of species has brought about the breakdown of 30 biogeographic boundaries that have historically limited the distributions of organisms 31 [2] and some of these species become "invasive" (i.e., an introduced population maintains itself without human assistance, spreads further and has impacts on 32 33 biodiversity, health, agriculture or ecosystem functioning) [3]. As biological invasions 34 are a leading cause of global biodiversity loss and the erosion of ecosystem functions worldwide [4], it is important to develop a better understanding of the invasion process 35 36 [5].

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38 In several species, it has been observed that introduced populations are themselves the source of additional new introductions (e.g. [6–9]), leading to a self-accelerating process 39 40 whereby "invasion begets invasion". This phenomenon has been called the "bridgehead effect", using the analogy of a military unit establishing a foothold at the far side of a 41 42 bridge, prior to further incursions into hostile territories [10]. Recently, the bridgehead 43 effect has been highlighted by a horizon scan identifying the most important issues 44 likely to affect how invasion processes and dynamics are studied in the near future [11], 45 because it could drive steep rises in global invasion rates [12]. A potential explanation 46 for an introduced population being the source of several secondary introductions is that 47 this bridgehead populations evolve "higher invasiveness"; that is, they acquire new

traits increasing the probability of successful establishment and further spread relative 48 49 to native populations [10]. While this idea of adaptive evolution as a driver of 50 bridgehead effects is appealing, we argue here that there is currently no empirical 51 support for this hypothesis. Most studies that suggested a role of adaptive evolution are 52 based on the observation that introduced populations are the source of one or several 53 secondary introductions of invasive species. In the first part of this review we show that 54 none of these studies tested for the appearance of new adaptive traits in the bridgehead 55 population. Moreover, we argue that a high frequency of secondary introductions can be 56 explained by the topology of human transport networks, with species more likely to be 57 introduced to, and spread from, highly connected hubs. Next, we discuss the few studies that actually demonstrated genetic changes in a bridgehead population, emphasising 58 59 that these observed genetic changes have not been shown to increase invasiveness. Finally, we outline the evidence that would be needed to demonstrate adaptive 60 evolution and higher invasiveness of introduced populations. 61

62 ii. Evidence for adaptive evolution in invasive populations

We conducted a literature search on Web of Science in July 2017 using the key words 63 "biological invasions" OR "invasive" Or "introduc*" OR "alien" AND "bridgehead effects" 64 OR "secondary introductions" OR "secondary spread" OR "multiple introductions". For 65 66 our review, we retained all papers that described invasion histories with secondary 67 spread from an initial invasive population and which have hypothesized that this could 68 be explained by evolution of higher invasiveness in the bridgehead population 69 compared to native populations (listed in Table 1). These papers have used either a 70 spatial spread pattern or genetic changes in the invaded range to suggest adaptive 71 evolution as a driver of bridgehead effects.

72 Spread patterns as evidence of adaptive evolution in bridgehead

73 populations

74 The vast majority of the studies suggesting a role of adaptive evolution used genetic 75 markers to retrace the invasion history of a single invading species. Because a single 76 bridgehead population was found to be the source of several secondary introductions 77 (Table 1), while no (or few) new invasions occurred through direct introductions from 78 the native range, the authors hypothesise that an adaptation for increased invasiveness 79 may have occurred (Table 1). While some of these studies are careful to underline that 80 the appearance of new adaptations in the bridgehead population is only one of the 81 possible explanations, others such as Zepeda et al. [13] are less cautious, making 82 statements such as "Biological invasions are rapid evolutionary events in which 83 populations are usually subject to a founder event during introduction followed by rapid 84 adaptation to the new environment". For example, in the study that originally coined the term "bridgehead effect", the authors concluded that adaptive evolution most likely 85 86 occurred in the introduced population of the Asian harlequin ladybird in the USA because this population was the source of several secondary invasions in Europe, Africa 87 88 and South America while all attempts to establish new populations for biocontrol 89 purposes by releasing individuals from the native population failed [10]. However, there 90 is currently no empirical support for adaptive evolution in the introduced US population. 91 Moreover, and importantly, successful establishment is a rare event with the effect that 92 in most cases a species needs to be introduced many times in sufficient numbers before 93 a self-sustaining population establishes [14]. Therefore, the failure of individuals 94 originating from their native range to establish a new population does not demonstrate 95 an evolutionary shift in the invasive populations.

97 These studies have not demonstrated adaptation. But many suggest that adaptive 98 evolution could drive the observed secondary spread, arguing that this would be 99 "evolutionary parsimonious" because a single evolutionary shift in a single population, 100 the bridgehead population would be required to explain increased invasion success (e.g. 101 [10,15–19]). However, this argument is problematic because adaptation is often not 102 necessary for establishment of a species outside of its native range. Recent studies 103 indicate that in the majority of cases species invade habitats with environmental 104 conditions similar to those in their native range [20–22]. Even in cases where the 105 introduced populations occupy different environmental conditions, this is often within 106 the species fundamental niche, with the effect that establishment requires no adaptation 107 [23]. Moreover, there are two alternative explanations for bridgehead populations 108 being the source of the majority of new introductions, in the absence of the evolution of 109 any new adaptive traits. The first is associated with the fact that introduced populations 110 frequently reach much higher densities than those in native range, because of increased 111 resource availability, filling of an empty niche or release from natural enemies such as 112 herbivores, predators, pathogens and parasites absent from the introduced range 113 (reviewed in [24]). All of these can lead to increased biomass and abundance in the 114 introduced range compared to the native range [25] with subsequent greater likelihood 115 of movement to new non-native regions. The second explanation is associated with the 116 network properties of human commerce. Dispersal of invasive species is tightly linked to 117 trade networks. For example, the invasion history of *Solenopis geminata* seems to closely 118 follow the Spanish trade routes in the 16th century [26]. Current human transport 119 networks are heterogeneous with most nodes (cities, ports, countries) having few 120 connections and a few nodes (e.g., transport hubs) having many connections and these 121 nodes are transport hubs [27]. Moreover, these networks have small-world properties,

meaning that any node in the network can be reached from any other node in a few
steps [27]. Given these network properties, secondary spread can be expected to be the
rule rather than the exception [27,28].

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In summary, the statement that adaptive evolution in bridgehead populations is the
most "evolutionary parsimonious" explanation for the bridgehead phenomenon is
mistaken; the most parsimonious scenario is simply no adaptation at all. Therefore, in
the absence of any convincing evidence for adaptive evolution, increased abundance in a
primary invaded area or the peculiar topology of the transport network should be the
null hypothesis for explaining bridgehead effects.

133 Genetic changes taken as evidence of adaptive evolution in bridgehead

134 populations

Two studies suggested that genetic changes could underlie greater invasiveness of
bridgehead populations [13,29]. However, none of these have demonstrated a causal
link between the trait change and invasion success and cannot exclude that other
processes have played a role in observed trait changes.

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In the first study the authors provide a detailed study of how the tobacco aphid (*Myzus persicae nicotianae*) was first introduced in North America and then to South America
[13]. They show that introductions were associated with bottlenecks. The authors also
suggest that shifts in the mode of reproduction may mitigate the effect of reduced
genetic diversity mentioning the Argentine ant as an example. In this species a reduction
in genetic diversity has been suggested to be at the basis of the success through a shift in

146 social structure of colonies [30]. However, a later study revealed that there is no 147 difference in social structure between native and introduced populations of the 148 Argentine ant, the only difference being the size of the supercolonies which is larger in 149 the introduced range, probably as the result of ecological factors [31]. In the case of their 150 study on aphids, Zepeda-Paulo et al. [13] mention that a shift to asexual reproduction 151 might have been important to circumvent the loss of genetic variation associated with 152 bottlenecks during introductions. However, there is yet not data showing that the 153 establishment probability of asexual aphid populations is greater than sexual 154 populations.

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156 The second study suggested that increased genetic diversity, instead of reduced genetic 157 diversity, may have conferred greater adaptive potential in a population of glossy 158 buckthorn introduced from a bridgehead region [29]. Due to admixture resulting from 159 several primary introductions, the bridgehead population had high allelic richness at 160 loci with putative ecological functions. Moreover, there was evidence of selection on loci 161 putatively involved in reproductive functions, which was interpreted by the authors as 162 an adaptive shift towards more "reproductive" phenotypes. However, this study did not 163 identify any phenotypic differences between individuals from the native and introduced 164 range nor did it demonstrate that the introduced population had a greater colonization 165 success compared to native populations. Although previous work has shown that 166 admixture in invasive populations may increase fitness relative to native populations 167 [32], it is currently not possible to conclude from the evidence in this study if adaptation 168 played any role in the secondary spread of glossy buckthorn.

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170 iii. Evidence needed

171 In order to demonstrate that adaptive evolution is an important driver of secondary 172 introductions, one would need to first establish that a trait conferring greater invasion 173 success evolved in the bridgehead population. Thus far, there is only limited evidence 174 for adaptive evolution in introduced populations (reviewed in [33]), with the notable 175 exception of a well-designed common garden experiment comparing the fitness of 176 invasive plant populations along a climatic gradient over several growing seasons [34]. 177 But there is currently no specific evidence of adaptive evolution in bridgehead 178 populations, favouring secondary spread. Here, we outline what type of evidence would

179 be required to test this hypothesis.

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Demonstrating that a trait conferring greater invasion success evolved in a bridgehead 181 182 population requires several steps (Fig. 1). The first is to determine the exact origin of the 183 introduced population. Because organisms display geographical variation in many 184 phenotypic traits, demonstrating a difference between a native and an introduced 185 population does not allow one to make any conclusion about evolution in the introduced 186 range as it may simply reflect variation in the native range. For example, most native 187 populations of little fire ants thrive under tropical climatic conditions, while some 188 introduced populations have established in colder locations [35]. This could be taken as 189 evidence for a potential shift in climatic tolerances in the introduced population. 190 However, a genetic study revealed that the source population of the introduction is 191 located at the edge of the native distribution and was therefore already pre-adapted to 192 establishing under a Mediterranean climate [35].

194 The second step is to demonstrate that phenotypic differences between a source and 195 introduced population have a genetic basis. Many traits are plastic and differences 196 between any two populations may simply reflect the effect of ecological and biotic 197 factors. The third step consists in demonstrating that a genetic change that occurred in 198 the primarily introduced population provides a selective advantage. This requires 199 carrying-out well designed reciprocal transplant experiments to compare the native 200 source and the bridgehead populations [33]. Ideally, such studies should measure fitness 201 over multiple growing seasons or generations [33]. As this is not always feasible, an 202 increasing number of studies use genomic screens to search for signatures of natural 203 selection at loci possibly involved in adaptation [36,37]. For example, a recent study 204 revealed that selection on loci associated with flowering time could play a role in 205 adaptive evolution in the invasion of Pyrenean rocket [38]. It is however important to 206 note that genomic screens are only correlative and that experiments comparing the 207 fitness of organisms with identified genetic differences remain necessary to 208 demonstrate adaptation.

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210 The final step is to show that a genetic difference between a source and an introduced 211 population is responsible for the success of the bridgehead population as a source of 212 secondary invasions (i.e. that it is a pre-adaptation for further invasions). For example, 213 an adaptation to the climatic conditions of a bridgehead population will act as a pre-214 adaptation only if the habitat where secondary introductions occur has climatic 215 conditions more similar to those of bridgehead population than to those of the native 216 population. In general, demonstrating that the bridgehead population is pre-adapted for 217 secondary spread is difficult since it would require conducting controlled introductions, 218 which raise ethical issues (fears that the invasive species may escape into the wild) and

because, it is logistically challenging, as the experiment would have to be repeated atdifferent locations to preclude the role of idiosyncratic ecological factors.

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222 An alternative approach to testing experimentally if adaptive evolution is a driver of 223 secondary spread is to investigate whether historical introductions from the introduced 224 range more frequently lead to successful establishment than introductions from the 225 native range. Given that this would require accounting for variation in propagule 226 pressure from the native and invaded range [39], it may be difficult to obtain sufficient 227 data for a single species because numbers of introductions are generally limited. Larger 228 sample size could possibly be obtained using data from multiple species. Alternatively, it 229 might also be possible to use data from biocontrol attempts, where species have been 230 introduced by humans on purpose. The idea would be to test if introductions from the 231 introduced range of a species are more successful than introductions directly from the 232 native range, while accounting for propagule pressure. The advantage of this type of 233 data is that there are observations of both successes and failures and perhaps even 234 estimates of the actual propagule pressure, which would be important for this analysis. 235

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Perspectives and Conclusion

The bridgehead effect has recently become a focus of invasion science because it has the potential to drive global increases in future invasion rates. A popular explanation for the bridgehead effect is that it stems from the evolution of higher invasion success of the bridgehead population. In many studies, authors are careful to suggest that alternative hypotheses are possible. However, collectively these studies placing the bridgehead effect in the context of an "evolutionary parsimonious scenario" give the impression that

244 the evolution of traits conferring greater invasiveness in the bridgehead population is an 245 important factor of the invasion process, in particular by favouring secondary spread. 246 However, in the absence of empirical evidence for adaptation, the null hypothesis should 247 be that secondary spread from bridgehead populations simply stems from increased abundance in bridgehead populations or general properties of transport networks. Yet, 248 249 for the purpose of the management of biological invasions the point made previous 250 authors [10] that there should be heightened vigilance against invasive bridgehead 251 populations is still valid. This is because bridgehead population are generally more 252 likely to generate new introductions than native populations, independently of the 253 underlying process generating this secondary spread. 254 255 Acknowledgements 256 We thank A. Liebhold, S. Ollier, D. Simberloff and three anonymous referees for their 257 comments and the Swiss NSF and two ERC advanced grants for financial support. 258 References 259 260 1 Gaston, K. (2009) Geographic range limits of species. Proc. R. Soc. B Biol. Sci. 276, 261 1391-1393 262 2 Capinha, C. et al. (2015) The dispersal of alien species redefines biogeography in 263 the Anthropocene. *Science.* 348, 1248–1251 264 3 Simberloff, D. et al. (2013) Impacts of biological invasions: what's what and the 265 way forward. Trends Ecol. Evol. 28, 58-66 266 Mack, R. et al. (2000) Biotic invasions: causes, epidemiology, global consequences, 4 267 and control. *Ecol. Appl.* 10, 689–710 268 5 Wilson, J.R.U. et al. (2009) Something in the way you move: dispersal pathways

- affect invasion success. *Trends Ecol. Evol.* 24, 136–144
- 270 6 Darling, J.A. *et al.* (2008) Genetic patterns across multiple introductions of the
- 271 globally invasive crab genus *Carcinus*. *Mol. Ecol.* 17, 4992–5007
- 272 7 Ficetola, G.F. *et al.* (2008) Population genetics reveals origin and number of
- founders in a biological invasion. *Mol. Ecol.* 17, 773–782
- 274 8 Grapputo, A. *et al.* (2005) The voyage of an invasive species across continents:
- 275 genetic diversity of North American and European Colorado potato beetle
- 276 populations. *Mol. Ecol.* 14, 4207–4219
- Miller, N. *et al.* (2005) Multiple transatlantic introductions of the western corn
 rootworm. *Science.* 310, 992–992
- 279 10 Lombaert, E. *et al.* (2010) Bridgehead effect in the worldwide invasion of the
- biocontrol harlequin ladybird. *PLoS One* 5, e9743
- 281 11 Ricciardi, A. *et al.* (2017) Invasion science: a horizon scan of emerging challenges
 282 and opportunities. *Trends Ecol. Evol.* 32, 464–474
- 283 12 Garnas, J.R. *et al.* (2016) Complex patterns of global spread in invasive insects:
- eco-evolutionary and management consequences. *Biol. Invasions* 18, 935–952
- 285 13 Zepeda-Paulo, F.A. *et al.* (2010) The invasion route for an insect pest species: The
 286 tobacco aphid in the New World. *Mol. Ecol.* 19, 4738–4752
- 287 14 Lockwood, J.L. *et al.* (2005) The role of propagule pressure in explaining species
 288 invasions. *Trends Ecol. Evol.* 20, 223–228
- 289 15 Guillemaud, T. et al. (2011) Biological invasions in agricultural settings: insights
- 290 from evolutionary biology and population genetics. *Comptes Rendus Biol.* 334,
 291 237–246
- 292 16 Lawson Handley, L.J. *et al.* (2011) Ecological genetics of invasive alien species.
- *BioControl* 56, 409–428

- Lombaert, E. *et al.* (2014) Complementarity of statistical treatments to reconstruct
 worldwide routes of invasion: the case of the Asian ladybird *Harmonia axyridis*. *Mol. Ecol.* 23, 5979–5997
- 297 18 Estoup, A. and Guillemaud, T. (2010) Reconstructing routes of invasion using
 298 genetic data: Why, how and so what? *Mol. Ecol.* 19, 4113–4130
- 299 19 Leduc, A. *et al.* (2015) Bridgehead invasion of a monomorphic plant pathogenic
- 300 bacterium : *Xanthomonas citri* pv . *citri*, an emerging citrus pathogen in Mali and
 301 Burkina Faso. *Environ. Microbiol.* 17, 4429–4442
- 302 20 Petitpierre, B. *et al.* (2012) Climatic Niche Shifts Are Rare Among Terrestrial Plant
 303 Invaders. *Science.* 335, 1344–1348
- 304 21 Strubbe, D. *et al.* (2013) Niche conservatism in non-native birds in Europe: niche
 305 unfilling rather than niche expansion. *Glob. Ecol. Biogeogr.* 22, 962–970
- **306** 22 Strubbe, D. *et al.* (2015) Niche conservatism among non-native vertebrates in
- **307** Europe and North America. *Ecography.* 38, 321–329
- **308** 23 Tingley, R. *et al.* (2014) Realized niche shift during a global biological invasion.
- 309 Proc. Natl. Acad. Sci. U. S. A. 111, 10233–10238
- 310 24 Catford, J. a. *et al.* (2009) Reducing redundancy in invasion ecology by integrating
- 311 hypotheses into a single theoretical framework. *Divers. Distrib.* 15, 22–40
- 312 25 Prior, K. *et al.* (2015) Insights from community ecology into the role of enemy
- 313 release in causing invasion success: the importance of native enemy effects. *Biol.*
- 314 *Invasions* 17, 1283–1297
- 315 26 Gotzek, D. *et al.* (2015) Global invasion history of the tropical fire ant: a stowaway
 316 on the first global trade routes. *Mol. Ecol.* 24, 374–388
- 317 27 Banks, N.C. *et al.* (2015) The role of global trade and transport network topology
- in the human-mediated dispersal of alien species. *Ecol. Lett.* 18, 188–199

- 319 28 Floerl, O. *et al.* (2009) The importance of transport hubs in stepping-stone
 320 invasions. *J. Appl. Ecol.* 46, 37–45
- 321 29 De Kort, H. *et al.* (2016) Transatlantic invasion routes and adaptive potential in
- 322 North American populations of the invasive glossy buckthorn, *Frangula alnus*.
- *Ann. Bot.* 118, 1089–1099
- 324 30 Tsutsui, N.D. *et al.* (2000) Reduced genetic variation and the success of an invasive
 325 species. *Proc. Natl. Acad. Sci. U. S. A.* 97, 5948–5953
- 326 31 Vogel, V. et al. (2009) Dynamics and genetic structure of Argentine ant
- 327 supercolonies in their native range. *Evolution.* 63, 1627–1639
- 328 32 Lavergne, S. and Molofsky, J. (2007) Increased genetic variation and evolutionary
- 329 potential drive the success of an invasive grass. *Proc. Natl. Acad. Sci. U. S. A.* 104,
- 330 3883-3888
- 331 33 Colautti, R.I. and Lau, J.A. (2015) Contemporary evolution during invasion:
- evidence for differentiation, natural selection, and local adaptation. *Mol. Ecol.* 24,
- 333 1999–2017
- 334 34 Colautti, R.I. and Barrett, S.C.H. (2013) Rapid adaptation to climate facilitates
- range expansion of an invasive plant. *Science.* 342, 364–366
- 336 35 Rey, O. *et al.* (2012) Where do adaptive shifts occur during invasion? A
- 337 multidisciplinary approach to unravelling cold adaptation in a tropical ant species
- invading the Mediterranean area. *Ecol. Lett.* 15, 1266–1275
- 339 36 Savolainen, O. *et al.* (2013) Ecological genomics of local adaptation. *Nat. Rev.*
- 340 *Genet.* 14, 807–820
- 341 37 Dlugosch, K.M. *et al.* (2015) The devil is in the details: genetic variation in
- introduced populations and its contributions to invasion. *Mol. Ecol.* 24, 2095–
- 343 2111

- 344 38 Vandepitte, K. *et al.* (2014) Rapid genetic adaptation precedes the spread of an
 345 exotic plant species. *Mol. Ecol.* 23, 2157–2164
- 346 39 Colautti, R.I. *et al.* (2006) Propagule pressure: a null model for biological
 347 invasions. *Biol. Invasions* 8, 1023–1037
- 348 40 Ascunce, M.S. *et al.* (2011) Global invasion history of the fire ant *Solenopsis invicta*.
 349 *Science.* 331, 1066–8
- 350 41 Gau, R.D. *et al.* (2013) Global genetics and invasion history of the potato powdery

351 scab pathogen, *Spongospora subterranea* f.sp. *subterranea*. *PLoS One* 8, e67944

- 352 42 Keller, S.R. *et al.* (2012) Bayesian inference of a complex invasion history revealed
- by nuclear and chloroplast genetic diversity in the colonizing plant, *Silene latifolia*.
- 354 *Mol. Ecol.* 21, 4721–4734
- 355 43 Yang, X.M. *et al.* (2012) Invasion genetics of the western flower thrips in China:
- evidence for genetic bottleneck, hybridization and bridgehead effect. *PLoS One* 7,
- 357 e34567
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				monstrate tha bridgehead ef			
Paper group	Taxon	Identification of the origin of the introduced population(s)	Genetic basis for phenotypic differences	Genetic phenotypic difference demonstrated to confer greater invasiveness	Pre-adaptation for secondary spread	Alternative explanation(s) proposed for the observed pattern	Ref.
	Red imported fire ant (Solenopsis invicta)	Yes	No	No	No	Yes. High levels of secondary transport could reflect higher propagule pressure from the USA relative to native areas because the USA are highly connected to the transport network.	[40]
	Harlequin ladybird (Harmonia axyiris)	No	No	No	No	No	[10]
	(Intrinoma dayiris) Potato powdery scab pathogen (Spongospora subterranea f.sp. subterranea)	No	No	No	No	Yes. High levels of secondary transport from Europe rather than from the native range could be explained by historical colonization and exploration of Europeans, transporting contaminated potato specimens with them.	[41]
Pattern of spread	Colorado potato beetle (<i>Leptinotarsa</i> <i>decemlineata</i>)	No	No	No	No	Yes. High levels of secondary transport from Europe could be due to higher levels of commercial trade than in the native range, resulting in more common long-distance transport.	[8]
Patte	White campion (Silene latifolia)	Yes	No	No	No	No. But stated that a common garden would be necessary to test the hypothesis.	[42]
	Pathogenic citrus bacterium (Xanthomonas citri pv. citri)	Yes	No	No	No	Yes. Secondary transport from Mali could be explained by higher levels of exportation of citrus material from nurseries to neighbouring regions.	[19]
	Western corn rootworm (Diabrotica virgifera virgifera)	No	No	No	No	No. But stated that the hypothesis of adaptation needs further investigation.	[9]
	Western flower thrips (Frankliniella occidentalis)	No	No	No	No	Yes. Secondary spread from Kunming could be explained by the fact that Kunming is a center for floriculture transportation and a major center for international imports of plants in China.	[43]
Genetic changes	Glossy buckthorn (Frangula alnus)	Yes	Indirect evidence	Indirect evidence	No	Yes. Secondary spread could be attributed to role in the early trade of ornamental plants of New York. The invasion of more inland areas may have occurred from nurseries.	[29]
Genetic	Tobacco aphid (Myzus persicae nicotianae)	No	Yes	Indirect evidence	No	No	[13]

Table 1. Previous studies on bridgehead effects arguing that adaptive evolution	on may
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363 drive secondary spread, the type of evidence used to support the evolution of greater

invasion success and alternative explanations proposed for the observed pattern.

365	
366	Figure legend
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368	Figure 1. Evidence needed to demonstrate that adaptive evolution is an important
369	driver of secondary introductions.
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372	